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Air Pollutant Emissions 1990-2015



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Summary - sammendrag

This report documents the methodologies used in the Norwegian emission inventory of acidifying pollutants, particulate matters, heavy metals and persistent organic pollutants submitted under the UNECE Convention on Long-range Transboundary Air Pollution.

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Content

Executive Summary	6
1 Introduction	7
1.1 National Inventory Background	7
1.2 Institutional arrangements	7
1.3 Inventory preparation process	8
1.3.1 Pollutants included, data collection, processing and archiving	8
1.3.2 Definitions and structure	10
1.3.3 Archiving.....	10
1.4 Methods and data sources	10
1.4.1 Structure of the general emission model	11
1.4.2 The four axes: Pollutants, industries, fuels, and sources.....	12
1.5 Key Categories	15
1.6 QA/QC and Verification methods.....	15
1.6.1 QA Procedures	15
1.6.2 General QC procedures	16
1.6.3 Source category-specific QC procedures	19
1.6.4 Verification studies	20
1.7 General uncertainty evaluation	21
1.7.1 Acidifying substances and NMVOC	21
1.7.2 Heavy metals and POPs	21
1.8 General Assessment of Completeness	21
2 Explanation of key trends.....	24
2.1 Acidifying substances and NMVOC	24
2.1.1 Total acidifying emission	24
2.1.2 NO _x	24
2.1.3 SO ₂	26
2.1.4 NH ₃	28
2.1.5 NMVOC	30
2.2 CO	32
2.3 PM, POPs and heavy metals.....	33
2.3.1 PM ₁₀	34
2.3.2 PM _{2.5}	36
2.3.3 Black carbon.....	36
2.3.4 Dioxins.....	38
2.3.5 PAH-4.....	39
2.3.6 HCB.....	41
2.3.7 PCB.....	43
2.3.8 Lead	45
2.3.9 Cadmium	46
2.3.10 Mercury	48
2.3.11 Chromium, arsenic and copper	49
3 ENERGY (NFR sector 1)	51
3.1 Overview	51
3.2 Energy combustion.....	51
3.2.1 Overview	51
3.2.2 Energy industries.....	61
3.2.3 Manufacturing industries and construction	68
3.2.4 Transport.....	70
3.2.5 Other sectors.....	103
3.2.6 International bunkers	106
3.3 Energy production (fugitive emissions from fuels)	107
3.3.1 Overview	107

3.3.2	Fugitive emissions from coal mining and handling	107
3.3.3	Fugitive emissions from uncontrolled combustion and burning coal dumps	108
3.3.4	Oil and natural gas	109
4	INDUSTRIAL PROCESSES AND PRODUCT USE (NFR sector 2)	117
4.1	Overview	117
4.2	Mineral products	117
4.2.1	Cement production	118
4.2.2	Lime production	120
4.2.3	Glass and glassfibre production	121
4.2.4	Mining and extraction of stones and minerals	122
4.2.5	Construction and demolition	123
4.2.6	Ceramics	124
4.2.7	Non-metallurgical Magnesia Production	125
4.2.8	Sandpit and rock-crushing plant	126
4.2.9	Concrete pumice stone	127
4.2.10	Rock wool production	128
4.2.11	Production of mineral white (plaster)	130
4.2.12	Construction and repairing of vessels - Sandblasting	131
4.2.13	Leather preparing	132
4.2.14	Production of asphalt	133
4.3	Chemical Industry	134
4.3.1	Ammonia Production	134
4.3.2	Production of nitric acid	135
4.3.3	Silicon carbide	136
4.3.4	Production of calcium carbide	139
4.3.5	Production of titanium dioxide	141
4.3.6	Production of methanol	142
4.3.7	Production of sulphuric acid	143
4.3.8	Production of plastic	144
4.3.9	Production of explosives	145
4.3.10	Chloralkali production	146
4.3.11	Production of pigments	146
4.3.12	Production of soap	147
4.3.13	Paint and varnish production	148
4.4	Metal production	148
4.4.1	Production of iron and steel	149
4.4.2	Production of ferroalloys	150
4.4.3	Production of primary aluminium	157
4.4.4	Production of secondary aluminium	160
4.4.5	Production of magnesium	162
4.4.6	Production of zinc	163
4.4.7	Production of nickel	164
4.4.8	Manufacture of anodes	165
4.5	Solvents and product use	167
4.5.1	Solvent losses (NMVOC)	167
4.5.2	Creosote-treated materials	171
4.5.3	Road paving with asphalt	172
4.5.4	Other product use	173
4.5.5	Mercury-containing products	174
4.5.6	Tobacco	174
4.6	Other production	176
4.6.1	Pulp and paper	176
4.6.2	Food and beverages industry	177
4.6.3	Ore mines	178
4.7	Wood processing	180
4.7.1	Wood processing	180
5	AGRICULTURE (NFR sector 3)	181

5.1	Overview	181
5.2	Livestock population characterisation	183
5.2.1	Data sources	183
5.2.2	Method for estimating number of cattle	184
5.2.3	Method for estimating number of sheep	185
5.2.4	Deviations from FAO statistics.....	185
5.2.5	Uncertainties	186
5.2.6	Source specific QA/QC.....	186
5.3	Nitrogen in animal manure as basis for emission estimates.....	186
5.4	Emissions from manure management	189
5.4.1	Description	189
5.4.2	NH ₃ emissions from manure management.....	189
5.4.3	NO _x emissions from manure management	195
5.4.4	NMVOC emissions from manure management	196
5.4.5	PM emissions from manure management	198
5.4.6	Completeness.....	199
5.4.7	Source specific QA/QC.....	199
5.5	Crop production and agricultural soils	199
5.5.1	Description	199
5.5.2	NH ₃ emissions from agricultural soils	200
5.5.3	NO _x emissions from agricultural soil.....	205
5.5.4	NMVOC emissions from agricultural soils	206
5.5.5	Particle emissions from farm-level agricultural operations.....	206
5.5.6	Uncertainties	207
5.5.7	Completeness.....	208
5.5.8	Source specific QA/QC.....	208
5.5.9	Use of pesticides	208
5.6	Field burning of agricultural wastes	209
5.6.1	Description	209
5.6.2	Method	209
5.6.3	Activity data	209
5.6.4	Emission factors	210
5.6.5	Uncertainties	211
5.6.6	Completeness.....	211
5.6.7	Source specific QA/QC.....	211
5.7	Other agricultural emission sources	211
5.7.1	Description	211
5.7.2	NH ₃ emissions from treatment of straw	211
6	WASTE (NFR sector 5)	213
6.1	Overview	213
6.2	Solid waste disposal on land	214
6.2.1	Description	214
6.2.2	Method	214
6.2.3	Activity data	214
6.3	Compost production	215
6.3.1	Description	215
6.4	Waste incineration	216
6.4.1	Description	216
6.4.2	Method	216
6.4.3	Activity data	217
6.4.4	Emission factors	218
6.4.5	Uncertainties	218
6.4.6	Source specific QA/QC.....	219
6.5	Waste water handling	219
6.5.1	Description	219
6.5.2	Method	219
6.5.3	Activity data	219

6.6	Other emission sources from the waste sector	220
6.6.1	Description	220
6.6.2	Method	220
6.6.3	Activity data	221
6.6.4	Emission factors	221
6.6.5	Uncertainties	222
6.6.6	Source specific QA/QC.....	222
7	Other and Natural emissions	223
8	Recalculations and Improvements.....	224
8.1	Recalculations	224
8.1.1	Overall description of the recalculations for the long-range transboundary air pollutants	224
8.1.2	Specific description of the recalculations.....	224
8.1.3	Implications of the recalculations for long-range transboundary air pollutants	234
8.2	Planned improvements	236
8.2.1	Overview	236
8.2.2	General.....	236
8.2.3	Energy	236
8.2.4	Industrial processes and product use	237
8.2.5	Agriculture	237
8.2.6	Waste	237
9	Projections.....	238
9.1	Introduction.....	238
9.2	The baseline scenario	238
9.3	Methodology and key assumptions	239
9.3.1	Macroeconomic assumptions and CO ₂ emissions from the mainland economy....	239
10	Reporting on gridded emissions and LPS	241
10.1	Gridded emissions.....	241
10.1.1	EMEP grid squares.....	241
10.1.2	Scope	241
10.1.3	Recent improvements.....	241
10.1.4	Planned improvements.....	242
10.2	LPS.....	242
	References.....	243
	Appendix A	251
	Tier 1 Key Category Analysis- Norway – 2017 submission	251
	Appendix B	267
	Emission factors used in the estimations of emissions from combustion	267
	SO ₂ and heavy metals - Stationary and mobile combustion	268
	Aviation - NO _x , NMVOC, CO, particles and PAH	270
	Road traffic - NO _x , NMVOC, CO, NH ₃ , particles and PAH	272
	Navigation - NO _x , NMVOC, CO, particles and POPs	273
	Other mobile sources including railways - NO _x , NMVOC, CO, NH ₃ , particles and POPs....	275
	Other mobile sources including railways - NO _x , NMVOC, CO, NH ₃ , particles and POPs....	276
	NO _x - Stationary combustion	278
	NMVOC - Stationary combustion.....	280
	CO - Stationary combustion.....	281
	NH ₃ - Stationary combustion	283
	Particulate matter - Stationary combustion	283
	POPs (Persistent Organic Pollutants) - Stationary combustion	286
	Source: Emission factor: Finstad et al. (2001). PAH-profile: EEA (2013)POPs (Persistent Organic Pollutants) - Stationary combustion	291
	Appendix C.....	294

Uncertainty analysis	294
Appendix D	299
Economic sectors in the Norwegian emission model.....	299
Appendix E.....	304
Source classifications used in the Norwegian emission inventory	304

Executive Summary

This report documents the methodologies used in the Norwegian emission inventory of acidifying pollutants, particulate matters, heavy metals and persistent organic pollutants.

The Norwegian emission inventory is a joint undertaking between the Norwegian Environment Agency¹ and Statistics Norway. This report has been prepared by the Norwegian Environment Agency in close collaboration with Statistics Norway's Division for Energy and Environmental Statistics.

The most important changes since the 2016 submission are:

- Emissions from road traffic have been updated due to revised emission factors. NMVOC and CO has decreased throughout the period. NO_x has decreased for most years, but increased for 2011-2014.
- NMVOC and NO_x from agriculture (3B manure management systems and 3D agricultural soils) have been included in the emission inventory.
- Emissions of PM_{2.5}, PM₁₀ and TSP from agriculture have been revised according to EMEP guidelines 2016. Emissions from agricultural operations previously defined under source 3I are now defined under source 3D Agricultural soils. Emissions of PM from 3B (manure management systems) are included for the first time.
- Emissions of CO from production of aluminium have been included for all years 1990-2014. This has led to an increase in emissions of CO by 13.9 to 57.8 per cent.
- PAH-4 have been split into benzo(a)pyren, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene. Emission factors used for estimation of emissions of benzo(a)pyren, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are taken from EEA Guidebook 2013 and 2016, Finstad et al (2001) and Danish IIR (Aarhus University 2016). The PAH profile has been measured at some aluminum producing plants. This has also led to updated distribution of PAH-4 emissions from aluminum production.
- Emissions of SO₂, NO_x, NH₃, NMVOC, CO and heavy metals from car fires have been included. This has led to increased emissions. The increase for 2014 in emission of lead, chromium and copper were 153, 0.7 and 5 kg respectively.

Chapter 8.1 Recalculations gives a more thorough description of changes in the most recent emission calculations.

¹ Former names are "Climate and Pollution Agency" and "Norwegian Pollution Control Authority".

1 Introduction

1.1 National Inventory Background

The Norwegian national inventory for long-range transboundary air pollutants includes emission data for the years 1990-2015. The emissions covered in this report are those embraced by the convention on long-range transboundary air pollution, i.e. they are defined with a territorial delimitation. The calculation methods used and the documentation of these, are, as far as possible, in accordance with the strict demands formulated in the emission convention.

1.2 Institutional arrangements

The Norwegian emissions inventories have been produced for about three decades as collaboration between Statistics Norway (SSB) and the Norwegian Environment Agency.

Statistics Norway is responsible for the official statistics on emissions to air, and contributes to the reporting to the UNECE. Their tasks include:

- collection of activity data
- operation and further development of models for emission estimation
- emission calculations
- filling in most of the tables for international reporting to UNECE
- publishing national official statistics on emissions to air.

The Norwegian Environment Agency is responsible for:

- international reporting to the UNECE
- emission factors for all emission sources
- measured emission data from large industrial plants based on individual reports submitted to the Norwegian Environment Agency on a regular basis
- considering the quality and assuring necessary updating, of emission models like, e.g. the road traffic model.

Activity data² are collected either internally at Statistics Norway (e.g. data on energy use, industrial production, number of animals, etc.) or reported to Statistics Norway, and in some cases to the Norwegian Environment Agency, from external sources such as the Norwegian Petroleum Directorate (OD) and the Norwegian Public Roads Administration (VD). Emission figures are derived from models operated by Statistics Norway. In the modelling activities Statistics Norway makes use of the data collected by the Norwegian Environment Agency on emission factors and emissions from industrial plants.

² Data on the magnitude of human activity resulting in emissions or removals taking place during a given period of time.

The Norwegian Environment Agency is responsible for quality control of the data they deliver to the emission model operated by Statistics Norway, but Statistics Norway makes an additional consistency check (see chapter 1.6). Statistics Norway is responsible for quality control of the activity data and the emission figures from the model, and the Norwegian Environment Agency also participates in this quality control before reporting to the UNECE.

1.3 Inventory preparation process

The Norwegian emission inventory is based on a general emission model and a series of more detailed supplementary models, which cover specific emission sources and pollutants (e.g. road traffic, air traffic, solvents). These smaller models feed results into the general model. All models are operated by Statistics Norway.

Data and information on point sources are recorded at the Norwegian Environment Agency in the database *Forurensning* and published in *Norske utslipp* (<http://www.norskeutslipp.no>). This is the Norwegian Pollutant Release and Transfer Register (PRTR). *Forurensning* is a further development of the old register Inkosys, which was introduced in 1978 as an internal tool for the authorities. The database was upgraded in 1992, and has later been under continuous development in order to harmonise with the PRTR adopted by the OECD in 1996. Each polluting industrial installation or plant is subjected to licensing and is obliged to produce an annual report to the pollution control authorities. The report should provide activity data, emission figures and information about the particular source, and it should address compliance with current environmental standards. The Norwegian Environment Agency supplies Statistics Norway with data from the Norwegian PRTR which are relevant for the preparation of the national emission inventory.

1.3.1 Pollutants included, data collection, processing and archiving

Statistics Norway collects the majority of data necessary to run the Norwegian emission model. These are as follows: activity levels, emission factors, aggregated results from the smaller, supplementary models and emission figures for point sources. Table 1.1 gives an overview of pollutants included in the emission inventory which are restricted by CLRTAP.

Table 1.1. Definition of pollutants in the Norwegian emission inventory which are restricted by CLRTAP

Class	Pollutant	Symbol	Definition
Acidifying gases	Sulphur dioxide	SO ₂	NO + NO ₂
	Nitrogen oxides	NO _x	
	Ammonia	NH ₃	
Heavy metals (HM)	Lead	Pb	
	Cadmium	Cd	
	Mercury	Hg	
	Arsenic	As	
	Chromium	Cr	
	Copper	Cu	
Persistent organic pollutants (POPs)	Polycyclic Aromatic Hydrocarbons	PAH-4	Emissions are calculated for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene.
	Dioxins	-	Dioxin emissions are given in the unit I-TEQ, which is required for reporting to CLRTAP. I-TEQ is based on the international model ("Nato-modell") and is the sum of PCDD/PCDF multiplied by the components toxicity equivalency factor (I-TEF). TEQ = sum (PCDD _i * TEF _i) + sum (PCDD _j * TEF _j).
	Hexachlorobenzene	HCB	
	Polychlorinated biphenyl	PCB	
Particulates	Total suspended particulates	TSP	Particulate matter with diameter less than 10µm Particulate matter with diameter less than 2.5µm
	-	PM ₁₀	
	-	PM _{2.5}	
	Black carbon	BC	
Other pollutants	Carbon monoxide	CO	
	Non-methane volatile organic compounds	NM VOC	

Source: Statistics Norway/Norwegian Environment Agency

The collected data are subjected to the Quality Assurance and Quality Control (QA/QC) routines described in section 1.6 as well as source specific routines as described under each source chapter. They are subsequently processed by Statistics Norway into a format appropriate to enter the emission models. The models are designed in a manner that accommodates both the estimation methodologies reflecting Norwegian conditions and those recommended internationally.

Input data used and the model output are all stored at Statistics Norway. Relevant information including dates and procedures followed are also recorded.

1.3.2 Definitions and structure

The structure of this documentation follows the nomenclature used for reporting to the Convention on Long-Range Transboundary Air Pollution (CLRTAP) as Nomenclature For Reporting (NFR14).

The main sectors here are:

NFR 1. Energy

NFR 2. Industrial processes and product use

NFR 3. Agriculture

NFR 6. Waste

The description of the pollutants included is given in Table 1.1.

Emissions of heavy metals, POPs and particulates are further described in the reports Finstad et al. (2001), Finstad et al. (2002b), Finstad and Rypdal (2003) and Finstad et al. (2003). Emissions of black carbon is described in Aasestad (2013), emissions of HCB in Aasestad (2014) and emissions of PCB in Aasestad (2016).

1.3.3 Archiving

The national emission inventory is a part of Statistics Norway's data archiving system. All input data to, and results from, the general Norwegian emission model from every publication cycle are stored and documented in this system.

Several input data are used in preliminary calculations before entering into the general Norwegian emission model. This includes supplementary models such as road traffic and air traffic, as well as a number of simpler calculations that do not fit into the framework of the general model. The preliminary calculations are not included in the central archiving system, which is not suited for such a diverse collection of data. For some supplementary models there is an established archiving routine where all input data and results from every calculation cycle are stored.

1.4 Methods and data sources

This chapter describes the general structure of the Norwegian emission model. The model was developed by Statistics Norway (Daasvatn et al. 1992), (Daasvatn et al. 1994). It was redesigned in 2003 in order to improve reporting to the UNFCCC and UNECE, and to improve QA/QC procedures.

The Norwegian emission model is organised around a general emission model called “Kuben” (“the Cube”). Several emission sources, e.g. road traffic, air traffic and solvents are covered by more detailed supplementary models. Aggregated results from the supplementary models are used as input to the general model. The supplementary models are presented in the appropriate sections of chapters 3-8. This chapter describes the general emission model.

1.4.1 Structure of the general emission model

The general emission model is based on equation (1.1).

$$(1.1) \quad \text{Emissions (E)} = \text{Activity level (A)} \cdot \text{Emission Factor (EF)}$$

For emissions from *combustion*, the activity data concern energy use. In the Norwegian energy balance, the use of different forms of energy is allocated to industries (economic sectors). In order to calculate emissions to air, energy use must also be allocated to technical sources (e.g. equipment). After energy use has been allocated in this way, the energy balance may be viewed as a cube in which the three axes are fuels, industries and sources.

The energy use data are combined with a corresponding matrix of emission factors. In principle, there should be one emission factor for each combination of fuel, industry, source and pollutant. Thus, the factors may be viewed as a four-dimensional cube with pollutants as the additional dimension. However, in a matrix with a cell for each combination, most of the cells would be empty (no consumption). In addition, the same emission factor would apply to many cells.

Emissions of some pollutants from major manufacturing plants (point sources) are available from measurements or other plant-specific calculations. When such measured data are available it is possible to replace the estimated values by the measured ones:

$$(1.2) \quad \text{Emissions (E)} = [(A - A_{PS}) \cdot EF] + E_{PS}$$

where A_{PS} and E_{PS} are the activity and the measured emissions at the point sources, respectively. Emissions from activity for which no point source estimate is available ($A - A_{PS}$) are still estimated with the regular emission factor.

Non-combustion emissions are generally calculated in the same way, by combining appropriate activity data with emission factors. Some emissions may be obtained from current reports and investigations, and some are measured directly as described in chapters 3-8. The emissions are fitted into the general model using the parameters industry, source, and pollutant. The fuel parameter is not relevant here. The source sector categories are based on EMEP/NFR categories, with further subdivisions where more detailed methods are available.

1.4.2 The four axes: Pollutants, industries, fuels, and sources

The *pollutants* currently included in the model are listed in Table 1.1. The model uses approximately 130 *industries* (economic sectors). The classification is common with the basis data in the energy balance/accounts, and is almost identical to that used in the national accounts, which is aggregated from the European NACE (rev. 2) classification. The large number of sectors is an advantage in dealing with important emissions from manufacturing industries. The disadvantage is an unnecessary disaggregation of sectors with very small emissions. To make the standard sectors more appropriate for calculation of emissions, a few changes have been made, e.g. "Private households" is defined as a sector.

The *fuels* and technical *sources* used for combustion with energy use (NFR source sector 1A) are shown in Table 1.2, Table 1.3 and Table 1.4.

Table 1.2. Energy commodities in the Norwegian emission inventory

Energy commodity	Aggregate fuel category in NFR
Coal	Solid Fuels
Coke	Solid Fuels
Petrol coke	Liquid Fuels
Wood	Biomass
Wood waste	Biomass
Black liquor	Biomass
Wood pellets	Biomass
Wood briquettes	Biomass
Charcoal	Biomass
Natural gas	Gaseous Fuels
Refinery gas	Liquid Fuels
CO gas	Solid Fuels
Landfill gas	Biomass
Biogas	Biomass
Fuel gas	Liquid Fuels
LPG	Liquid Fuels
Gasoline (road transport)	Liquid Fuels
Aviation gasoline	Liquid Fuels
Kerosene (heating)	Liquid Fuels
Jet kerosene	Liquid Fuels
Autodiesel	Liquid Fuels
Marine gas oil	Liquid Fuels
Light fuel oils	Liquid Fuels
Heavy distillate	Liquid Fuels
Heavy fuel oil	Liquid Fuels
Municipal waste	Other Fuels
Special waste*	Other Fuels

* Special waste was moved from *Liquid* to *Other* fuels in 2014.

Source: Statistics Norway/Norwegian Environment Agency

Table 1.3. Sources for energy combustion in the Norwegian emission inventory

Source	NFR
Stationary combustion	
Direct fired furnaces	1A1, 1A2
Gas turbines	1A1c, 1A4a
Boilers	1A1, 1A2, 1A4, 1A5
Small stoves	1A2, 1A4, 1A5
Flaring	1B2C, 6C
Mobile combustion*	
Passenger car	1A3b i, 1A5b
Light duty vehicles	1A3b ii, 1A5b
Heavy duty vehicles	1A3b iii, 1A5b
Motorcycle	1A3b iv
Moped	1A3b iv
Snowscooter	1A4b, c
Railway	1A3c
Aviation jet/turboprop (0-100 m)	1A3a ii (i), 1A5b
Aviation jet/turboprop (100-1000m)	1A3a ii (i), 1A5b
Aviation jet/turboprop (cruise)	1A3a ii (ii), 1A5b
Aviation helicopter (0-100 m)	1A3a ii (i)
Aviation helicopter (100-1000m)	1A3a ii (i)
Aviation helicopter (cruise)	1A3a ii (ii)
Aviation small craft (0-100 m)	1A3a ii (i)
Aviation small craft (100-1000m)	1A3a ii (i)
Aviation small craft (cruise)	1A3a ii (ii)
Ships	1A3d, 1A4c, 1A5b
Small boats 2 stroke	1A4b
Small boats 4 stroke	1A4b, c
Equipment 2 stroke	1A4b, c
Equipment 4 stroke, tractor	1A2g-vii, 1A4a, b, c,

* For road transport the source split is more detailed in the sub-model. See section 3.2.4.2.

Source: Statistics Norway/Norwegian Environment Agency

Table 1.4. Combinations of fuels and sources in use

	Direct fired furnaces	Gas turbines	Boilers	Small stoves	Flaring	Passenger car	Light duty vehicles	Heavy duty vehicles	Motorcycle	Moped	Snowscooter	Railway	Aviation jet/turboprop	Aviation helicopter	Aviation small craft	Ships	Small boats 2 stroke	Small boats 4 stroke	Equipment 2 stroke	Equipment 4 stroke, tractor
Coal	X	..	X	X
Coke	X	..	X	X
Petrol coke	X	..	X
Fuel wood	X
Wood waste	X
Black liquor	X
Wood pellets	X	X
Wood briquettes	X
Charcoal	X
Natural gas	X	X	X	..	X	X	..	X	X
Refinery gas	X	..	X	..	X
CO gas	X	..	X
Landfill gas	X	..	X
Biogas	..	X
Fuel gas	X	..	X
LPG	X	X	..	X
Motor gasoline	X	X	X	X	X	X	X	X	X	X
Aviation gasoline	X
Kerosene (heating)	X	X
Jet kerosene	X	X
Auto diesel	X	X	X	X	X	X	..	X
Marine gas oil/diesel	X	X	X	X
Light fuel oils	X	X	X	X
Heavy distillate	X	..	X	X
Heavy fuel oil	X	..	X	X
Municipal waste	X
Special waste	X	..	X

Source: Statistics Norway

The sources for non-combustion emissions and for combustion without energy use are based on EMEP/NFR categories, with further subdivisions where more detailed methods are available.

1.5 Key Categories

Information about key categories is given in Appendix A.

1.6 QA/QC and Verification methods

This chapter describes general QA/QC procedures. For source specific QA/QC, see each source sector for detailed descriptions.

The QA/QC work has several dimensions. In addition to accuracy, also timeliness is essential. As these two aspects may be in conflict, the QA/QC improvements in recent years have been focused on how to implement an effective QA/QC procedure and how to obtain a more efficient dataflow in the inventory system.

During the past years several quality assurance and quality control procedures for the preparation of the national emission inventory have been established in Norway. Statistics Norway made its first emission inventory for some gases in 1983 for the calculation year 1973. The emission estimation methodologies and the QA/QC procedures have been developed continuously since then. Norway has implemented a formal quality assurance/quality control plan, which covers the reporting of long-range transboundary air pollution as well as greenhouse gases. A detailed description of this is presented in Annex V in the National Inventory Report 2016 (NEA 2016).

The established QA/QC procedures include the following:

- The Norwegian Environment Agency is the national entity designated to be responsible for the reporting of the national inventory of greenhouse gases to the UNFCCC and the reporting of long-range transboundary air pollution to the UNECE. This includes coordination of the QA/QC procedures;
- Statistics Norway is responsible for the quality control system with regard to technical activities of the inventory preparation;
- General inventory level QC procedures, as listed in table 6.1 in chapter 6 of the 2006 IPCC Guidelines (IPCC 2006), is performed every year;
- Source category-specific QC procedures are performed for all key categories and some non-key categories; with regard to emission factors, activity data and uncertainty estimates.

1.6.1 QA Procedures

According to the IPCC Good practice guidance, good practice for QA procedures requires an objective review to assess the quality of the inventory and to identify areas where improvements could be made. Furthermore, it is good practice to use QA reviewers that have

not been involved in preparing the inventory. In Norway, the Norwegian Environment Agency is responsible for reviewing the inventory with regard to quality and areas for improvement. For most source categories it is a person within the Norwegian Environment Agency who has not been involved in the calculations and the quality controls who performs the QA for the particular source.

Norway has performed several studies comparing inventories from different countries (Haakonsen et al. 2000). Verification of emission data is another element to be assessed during the elaboration of a QA/QC and verification plan.

Both Statistics Norway and the Norwegian Environment Agency are responsible for archiving the data they collect and the estimates they calculate with associated methodology documentation and internal documentation on QA/QC. Due to the differences in the character of data collected, Norway has chosen to keep archiving systems in the separate institutions, which means that not all information is archived at a single location. These archiving systems are, however, consistent, and operate under the same rules. Although the data are archived separately, all can be accessed efficiently during a review.

1.6.2 General QC procedures

The Norwegian emission inventory is produced in several steps. Preliminary estimates are first produced by Statistics Norway 4-5 months after the end of the inventory year. These data are based on preliminary statistics and indicators and data that have been subjected to a less thorough quality control. The "final" update takes place about one year after the inventory year. At this stage, final statistics are available for all sources. Recalculations of the inventory are performed annually, as methodological changes and refinements are implemented. In itself, this stepwise procedure is a part of the QA/QC-procedure since all differences in data are recorded and verified.

For each of the steps described above, general quality control procedures are performed, but with different levels of detail and thoroughness as mentioned. The national emission model was revised in 2002 in order to facilitate the QC of the input data rather than the emission data only. Input data include emissions reported from large plants, activity data, emission factors and other estimation parameters.

In the following, the procedures listed in table 6.1 in chapter 6 of the 2006 IPCC Guidelines (IPCC 2006), are described, as well as how these checks are performed for the Norwegian emission inventory.

Check that assumptions and criteria for the selection of activity data and emissions factors are documented

Thorough checks of emission factors and activity data and their documentation have been performed for existing emission sources. When new sources appear (for example a new industrial plant) or existing sources for the first time are recognised as a source, the Norwegian Environment Agency delivers all relevant information to Statistics Norway. This information is

then thoroughly checked by the inventory team at Statistics Norway. All changes in methodologies or data are documented and kept up to date.

Check for transcription errors in data input and references

Activity data are often statistical data. Official statistical data undergo a systematic revision process, which may be manual or, increasingly frequently, computerised. The revision significantly reduces the number of errors in the statistics used as input to the inventory. Furthermore, all input data (reported emissions, emission factors and activity data) for the latest inventory year are routinely compared to those of the previous inventory year, using automated procedures. Large changes are automatically flagged for further, manual QC. In addition, implied emission factors are calculated for emissions from stationary combustion at point sources. The IEFs are subjected to the same comparison between the years t and $t-1$. The most thorough checks are made for the gases and categories with the largest contribution to total emissions.

Check that emissions are calculated correctly

When possible, estimates based on different methodologies are compared. The Norwegian Environment Agency and Statistics Norway control and verify emission data reported to the Norwegian Environment Agency by industrial enterprises, registered in the database *Forurensning*. First, the Norwegian Environment Agency checks the data received from these plants, and if errors are discovered, they may then ask the plants' responsible to submit new data. Subsequently, Statistics Norway makes, where possible, occasional comparable emission calculations based on activity data sampled in official statistics, and deviations are explained through contact with the plants.

Check that parameter and emission units are correctly recorded and that appropriate conversion factors are used

All parameter values are compared with values used in previous years and with any preliminary figures available. Whenever large deviations are detected, the value of the parameter in question is first checked for typing errors or unit errors. Changes in emissions from large plants are compared with changes in activity level. If necessary, the primary data suppliers (e.g. The Norwegian Petroleum Directorate, Norwegian Public Roads Administration, various plants etc.) are contacted for explanations and possible corrections.

Check the integrity of database files

Control checks of whether appropriate data processing steps and data relationships are correctly represented are made for each step of the process. Furthermore, it is verified that data fields are properly labelled and have correct design specifications and that adequate documentation of database and model structure and operation are archived.

Check that the movement for inventory data among processing steps is correct

Statistics Norway has established automated procedures to check that inventory data fed into the model does not deviate too much from the figures for earlier years, and that the calculations within the model are correctly made. Checks are also made that emissions data are correctly transcribed between different intermediate products. The model is constructed so that it gives

error messages if factors are lacking, which makes it quite robust to miscalculations.

Check that uncertainties in emissions and removals are estimated correctly

For long-range transboundary air pollutants the last uncertainty analysis was undertaken in 2001. See further information about the uncertainty analysis in section 1.7 and Appendix C.

Undertake review of internal documentation

For some sources, expert judgements dating some years back are employed with regard to activity data/emission factors. In most of the cases these judgements have not been reviewed since then, and may not be properly documented, which may be a weakness of the inventory. The procedures have improved the last few years, and the requirements for internal documentation to support estimates are now quite strict; all expert judgements and assumptions made by the Statistics Norway staff should be documented. This should increase reproducibility of emissions and uncertainty estimates. The model at Statistics Norway has improved the process of archiving inventory data, supporting data and inventory records, which does facilitate review. The model runs are stored and may be reconstructed, and all input data from the Norwegian Environment Agency as well as notes with explanations on changes in emissions are stored. This is a continuous process of improvement at Statistics Norway.

Check of changes due to recalculations

Emission time series are recalculated every year to ensure time series consistency. The recalculated emission data for a year is compared with the corresponding figures estimated the year before. The intention is to explain all major differences as far as possible. Changes may be due to revisions in energy data, new plants, correction of former errors and new emission methodologies.

Undertake completeness checks

Estimates are reported for all source categories and for all years as far as we know, apart from a few known data gaps, which are listed in section 1.8. There may, of course, exist sources which are not covered. However, we are quite certain that emissions from potentially additional sources are very small or negligible. During comparisons with previous emission estimates, any emission calculations that have been erroneously omitted during the most recent production cycle will be identified and included.

Compare estimates to previous estimates

Internal checks of time series for all emission sources are performed every year when an emission calculation for a new year is done. It is then examined whether any detected inconsistencies are due to data or/and methodology changes. For example, in 2016 Statistics Norway/the Norwegian Environment Agency calculated emission data for 2015 for the first time. These data were compared with the 2014 figures for detection of any considerable deviations. Some large deviations may be correct, caused for instance by the shutdown of large industrial plants or the launch of new ones.

1.6.3 Source category-specific QC procedures

Statistics Norway and the Norwegian Environment Agency have carried out several studies on specific emission sources, e.g. emissions from road, sea, and air transport. These projects are repeated in regular intervals when new information is available. During the studies, emission factors have been assessed and amended in order to represent the best estimates for national circumstances, and a rationale for the choice of emission factor is provided. The emission factors are often compared with default factors from the most recent EMEP/EEA air pollutant inventory guidebook and emission factors from other literature. Furthermore, activity data have been closely examined and quality controlled.

The QC procedures with regard to emission data, activity data and uncertainty estimates for the different emission sources are described in the QA/QC-chapters of the relevant source-categories. The source category-specific analyses have primarily been performed for key categories on a case-by-case basis, which is described as being good practice. The QA/QC process for many of the sources could be improved.

In the following is a more detailed description of QC of emission data reported from plants:

Plant emission data that are used in the European Emission trading system (EU-ETS) will undergo annual QC checks. Activity data and emission estimates from plants that are included in the EU-ETS undergoes annual third-party verification. The source-specific QC checks for other plants are performed as part of the controls of the reporting under the emission permits.

The plant specific data undergo further QC from the emission inventory team at the Norwegian Environment Agency before figures are sent to Statistics Norway for inclusion in the emission inventory. Statistics Norway is responsible for reporting the results of the key category analysis to the Norwegian Environment Agency, and the agency places special emphasis on plants that belong to key categories.

At some point since the inclusion in the inventory, each plant has been QC checked more thoroughly, including:

- An assessment of the internal QA/QC of the plants reporting data to the Norwegian Environment Agency
 - Their QA/QC system including archiving
 - Any changes to the QA/QC system
- An assessment and documentation of measurements and sampling
 - Measurement frequency
 - Sampling
 - Use of standards (e.g. ISO)
 - Documentation for archiving
- An assessment and explanation of changes in emissions over time (e.g. changes in technology, production level or fuels) (annual check)

- An assessment of time-series consistency back to 1990 (if plant emission data are missing for some years and estimates are made using aggregate activity data and emission factors)
- A comparison of plant emissions to production ratios with those of other plants, including explanations of differences
- A comparison of the production level and/or fuel consumption with independent statistics
- An assessment of reported uncertainties (including statistical and non-statistical errors) to the extent this has been included in the reporting

The QC checks are made in close cooperation with the emission reporting plants. For more details of QA/QC of specific source categories, see “source specific QA/QC” in relevant chapters.

1.6.4 Verification studies

In general, the final inventory data provided by Statistics Norway are checked and verified by the Norwegian Environment Agency.

In the following, some verification studies which have been performed are briefly described. Emission estimates for a source are often compared with estimates performed with a different methodology.

In 2004, the Nordic Council of Ministers initiated a new project that was finalised in 2006. This project focused on NMVOC, heavy metals and POPs. An unpublished, final report has been worked out, containing the following elements:

- comparisons of the emission estimation methodologies and emission factors used in each country (review)
- identification of gaps in knowledge
- identification of possible "burden sharings" with respect to research areas (research taking place in one country, but used in all countries)
- discussions of the particular Nordic aspects influencing the emissions
- discussions of the possible contributions from research in the Nordic countries
- proposals for research areas

In 2006, the Nordic Council of Ministers initiated a new project that was finalised in 2010. This project focused on emission of particulate matter. The final report contains the following elements:

- comparisons of the emission estimation methodologies and emission factors used in each country (review)
- identification of gaps in knowledge
- discussions of the particular Nordic aspects influencing the emissions
- discussions of the possible contributions from research in the Nordic countries
- proposals for research areas
- recommendations for further work

In 2015, a Nordic project started, financed by the Nordic Council of Ministers, with the aim to

improve the Nordic emission inventories on heavy metals and POP compounds. In the proposed programme the experts working with air pollutant inventories in Denmark, Finland, Iceland, Norway and Sweden will have a yearly meeting for knowledge exchange. The main focus is to compare emission factors and methodologies used, inform about national studies performed, and study and set up further plans to develop national air pollution inventories, especially for emission sources where studies and other Nordic information sources can be used in developing methodologies suitable for Nordic countries.

1.7 General uncertainty evaluation

1.7.1 Acidifying substances and NMVOC

The emission estimates for long-range air pollutants in the Norwegian emission model may be ranked roughly in order of increasing uncertainty as follows:

$$\text{SO}_2 < \text{NO}_x < \text{NH}_3 \approx \text{NMVOC}$$

The sources of uncertainty in the emission estimates include sampling errors, poor relevance of emission factors or activity data, and gross errors.

Evaluation of the uncertainty in the long-range air pollutants is given in the report Rypdal and Zhang (2001). Summary tables with the results are given in Appendix C.

1.7.2 Heavy metals and POPs

The uncertainty is generally higher for HM and POPs than for other components in the Norwegian emission model except for N_2O . There are various reasons for this high uncertainty. The most important reason is that there is limited information about emission factors, and it is not clear how usable the emission factors found in international literature are for Norwegian conditions. Emission factors for some HM and POPs components are insufficient for some sources, so emission factors for similar sources have then been used. In addition it is not certain that all emission sources are known or sufficiently mapped. The industrial reporting to the Norwegian Environment Agency has improved in recent years. The reported figures can, however, vary a great deal from one year to another. For earlier years they can be insufficient, and since HM and POPs are to be calculated from 1990, recalculations are necessary. These recalculations are based on a combination of assumptions and knowledge of the plants. Emission figures from the early 1990s are therefore more uncertain than figures produced today.

1.8 General Assessment of Completeness

Norway is requested to report emissions to UNECE for the pollutants restricted by CLRTAP (Convention on Long-Range Transboundary Air Pollution). Minimum reporting request each year includes the acidifying pollutants (NO_x , SO_2 , NH_3) and NMVOC, the heavy metals Pb, Cd and Hg,

particulate matter (TSP, PM₁₀ and PM_{2.5}), CO and the POPs dioxins, Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4), HCB and PCB. Norway also voluntarily reports the heavy metals As, Cr and Cu and BC.

In terms of spatial coverage, the calculated air emissions cover all activities within Norway's jurisdiction.

In the case of temporal coverage, emission figures for CO, SO₂, NO_x, NH₃ and NMVOC are produced and updated every year for all years from 1990. For HM, POPs, particles and BC, emission figures are also produced for all years from 1990.

With regard to sectoral coverage, sources with relevant emission amounts which are not covered in the inventory even if emissions can be expected, have been reported as Not Estimated (NE) in the reported NFR tables. The following sources are not covered, even if emissions might be expected:

Energy sector:

- NH₃ emissions from Energy Industries (1A1), from stationary combustion in manufacturing industries and construction: Non-ferrous metals (1A2b), Pulp, Paper and Print, Food processing(1A2d), beverages and tobacco (1A2e), from National Navigation (1A3dii), from stationary combustion: in Commercial/Institutional (1A4i), Agriculture/Forestry/Fishing (1A4ci), from Fishing (1A4ciii) and from Venting and Flaring (1B2c)
- NH₃ and PCB emissions from Civil aviation, domestic and international LTO (1A3aii (i) and 1A3ai (i))
- Emissions of particulate matters (PM) from clutch wear (1A3b)
- Emissions of PM from use of unpaved roads (1A3b)
- Emissions of PM from sand strewing (1A3b)
- Emissions of PM from Mopeds & motorcycles (1A3biv)
- Emissions of HCB from Railways (1A3c), from Other sectors: mobile sources (1A4)
- Fugitive emissions of NMVOC, PM, HM and PAH from solid fuel transformation (1B1b)
- Fugitive emissions of SO₂ from oil (exploration, production, transport) (1B2ai) and from natural gas (by land-based desulphurisation) (1B2b)

Industry sector:

- Emissions of NMVOC and PM from asphalt roofing (2D3c)
- Emissions of CO and NH₃ from ammonia production (2B1)

Agricultural sector:

- Emission of NO_x from Urine and dung deposited by grazing animals (3Da2)
- Emission of NH₃ from crop residues applied to soils (3Da4)
- Emission of NMVOC and NH₃ from cultivated crops (3De)

Waste sector:

- Emissions of NMVOC from composting (5B1) and NH₃ from biogas facilities (5B2)
- Emissions of NMVOC and NH₃ from waste incineration (5C1)
- Emissions from sewage sludge incineration (5C1biv) and open burning of waste (5C2)
- Emissions of NH₃ from waste-water handling (5D)
- Evaporation of Hg from landfills and emission of Pb by detonation of explosives (5E)
- Emissions of dioxins by smoking processes for preservation of meat and fish (5E)

The reasons for not including these emission sources are mainly lack of activity data, emission factors or known calculation methodology.

In each sector chapter more details about completeness is given.

2 Explanation of key trends

2.1 Acidifying substances and NMVOC

2.1.1 Total acidifying emission

Emissions of gases that transform into acid can be expressed in terms of acid equivalents. Total emissions of the three gases NO_x , SO_2 and NH_3 measured as acid equivalents have been reduced by 27 per cent since 1990, from 7 350 tonnes acid equivalents to 5 370 tonnes acid equivalents. SO_2 and NO_x emissions have been reduced by 69 and 24 per cent since 1990, respectively, while NH_3 emissions have increased by 12 per cent. In 1990, NO_x constituted 59 per cent of the acidifying emissions, NH_3 19 per cent and SO_2 22 per cent, while, in 2015, NO_x , NH_3 and SO_2 were responsible for 61, 29 and 10 per cent of these emissions, respectively. Norway has met the 2010 targets defined by the Gothenburg Protocol for SO_2 , NMVOC and NO_x while NH_3 emissions are still above the Gothenburg 2010 target. NO_x target has only been met since 2015.

2020 targets defined by the revised Gothenburg Protocol have already been met for SO_2 and NO_x .

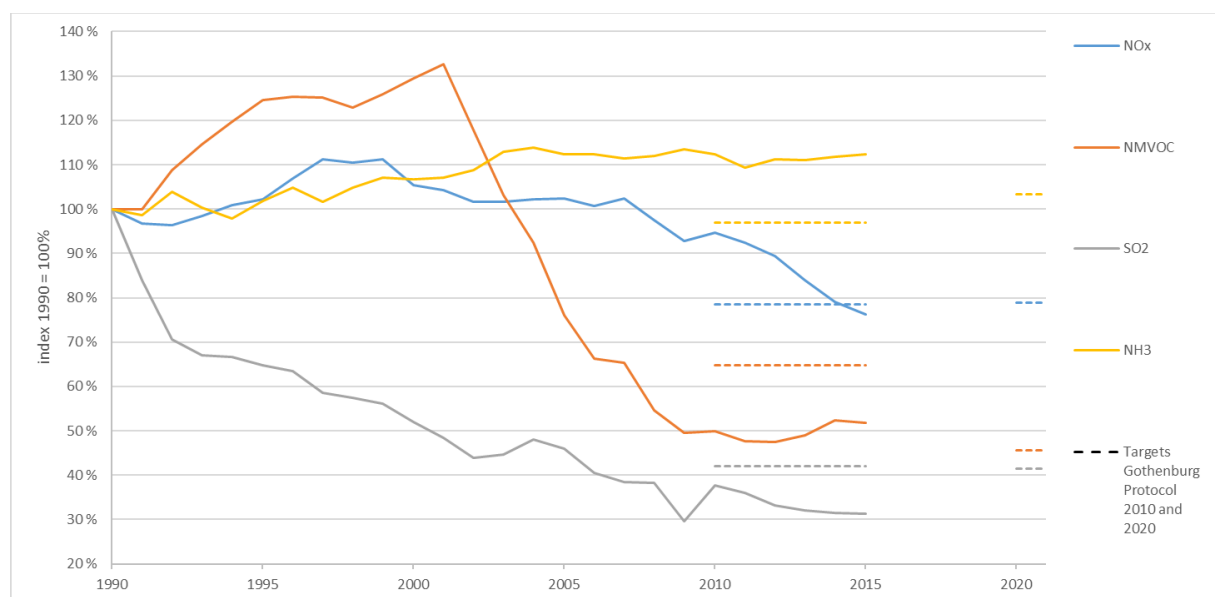


Figure 2.1. Trends in emissions for NO_x , SO_2 , NH_3 and NMVOC. 1990-2015. Index 1990 = 100%

Source: Statistics Norway/Norwegian Environment Agency

2.1.2 NO_x

- 2010 commitment of the Gothenburg Protocol: 156 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: a 23 per cent reduction compared to emissions in 2005, 156 700 tonnes.

Norway's NO_x emissions totaled 151 300 tonnes in 2015. The 2010 commitment of the

Gothenburg Protocol for NO_x emissions was reached in 2015 by approximately 5 500 tonnes. In 2015, emissions have also been lower than the 2020 commitment of the Gothenburg Protocol. NO_x emissions have been reduced by 24 per cent since 1990 and by 4 per cent since 2014. The biggest sources of NO_x emissions in 2015 were energy industries and transport, accounted for 33 and 34 per cent of the total emission, respectively. Emissions in the energy industries sector overall have increased by 79 per cent since 1990 while emissions from the transport sector overall have been reduced by 46 per cent since 1990.

In the energy industries sector, manufacturing of solid fuels and other energy industries accounts for 95 per cent, and have increased by 90 per cent since 1990. The transport reduction figure nonetheless hides significant changes within the transport sector: emissions from domestic and international aviation have increased by 134 per cent and 431 per cent respectively since 1990, whereas emissions from passenger cars have been reduced by 65 per cent.

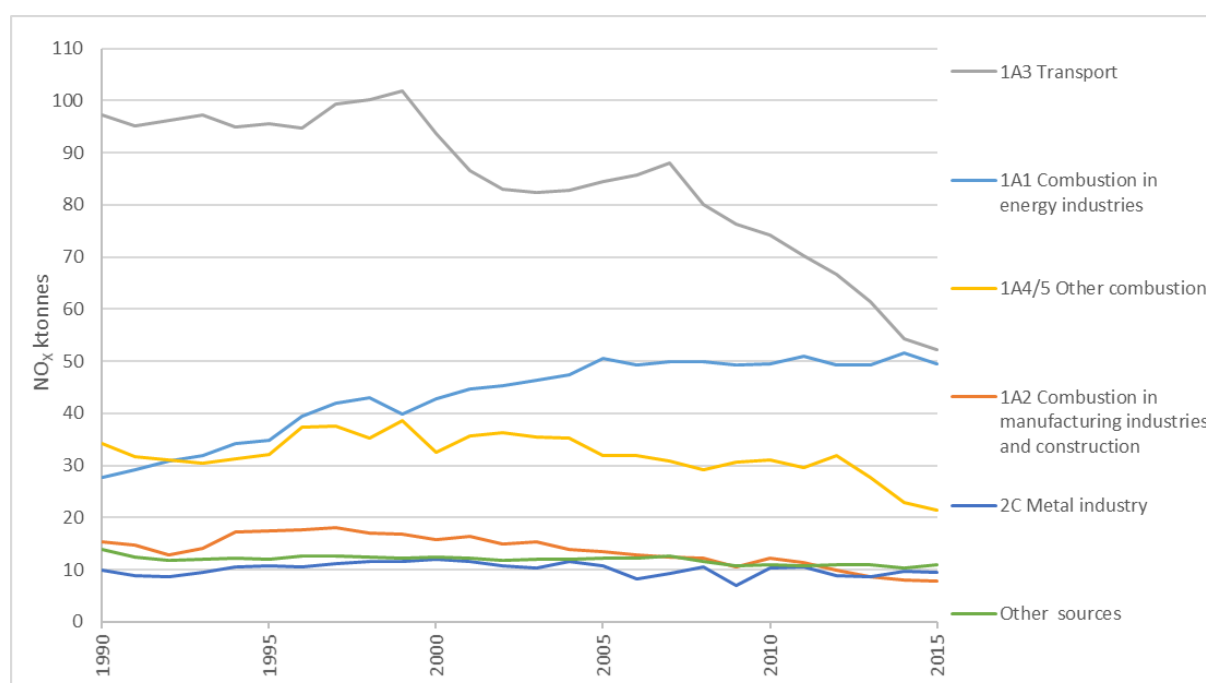


Figure 2.2. Trends in NO_x emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

Emissions from aviation have increased primarily due to increased traffic. Within road transport, traffic has also increased significantly from 1990 to 2015. NO_x emissions have however been reduced due to policy measures. Stricter emission requirements for new passenger cars, heavy duty vehicles and buses is the main cause of the reduction.

Emissions from national navigation have been reduced by 41 per cent since 1990. The reduced emissions in navigation are partly due to measures implemented by the NO_x Fund, which is financed by industry and businesses. The measures are not directly linked to fuel efficiency, and energy consumption in national navigation has not had the same decrease as the emissions. Passenger cars was the largest emission source within transport between 1990 and 1992. Since

1993, national navigation has been the largest emission source. In 2015, it is responsible of 34 per cent of emissions within the transport sector. Passenger cars and heavy duty vehicles followed, representing both 24 per cent.

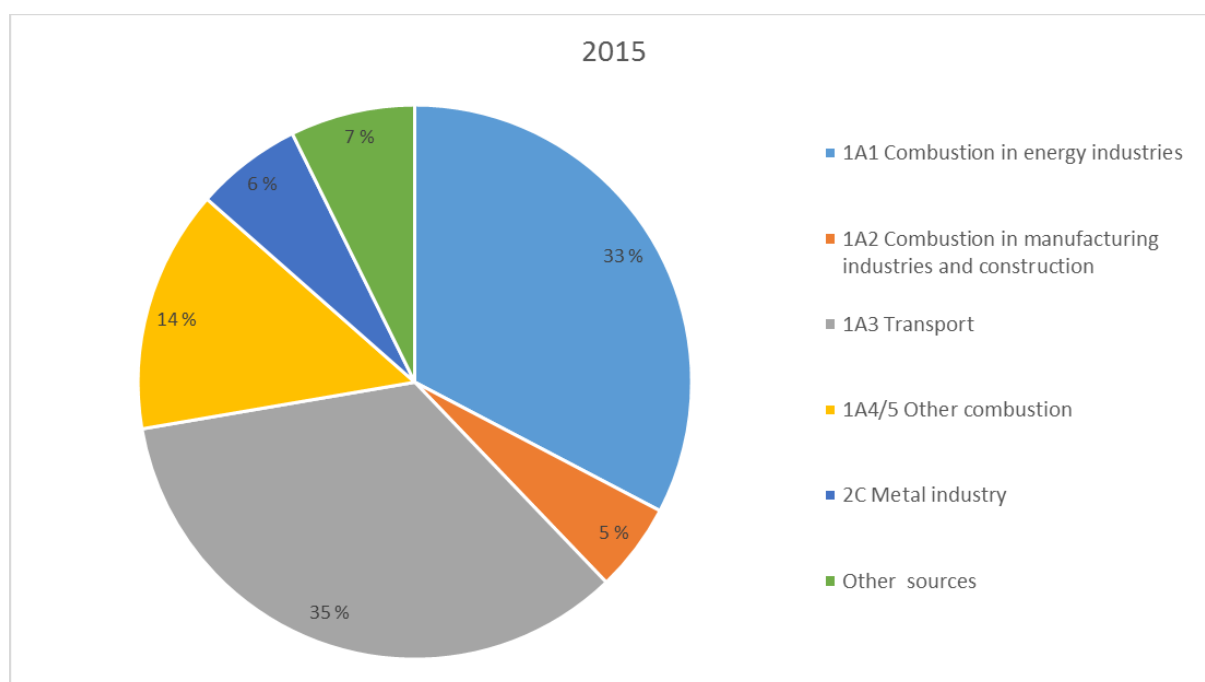


Figure 2.3. Distribution of NO_x emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

The third largest source of NO_x emissions in 2015 was “other combustion” (NFR 1A4 and 1A5), accounting for 14 per cent of the total NO_x emissions. National fishing is the largest source of emissions within the sector.

2.1.3 SO₂

- 2010 commitment of the Gothenburg Protocol: 22 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: a 10 per cent reduction compared to emissions in 2005, 21 600 tonnes.

Norway's SO₂ emissions totaled 16 300 tonnes in 2015. Both the 2010 and 2020 commitments of the Gothenburg Protocol for SO₂ emissions have been fulfilled since 2006.

The SO₂ emissions in Norway in 2015 have been reduced by 69 per cent since 1990. This has been achieved by pollution control, the closure of pollution-generating businesses, a reduction of sulfur content in petroleum products and a reduced consumption of petroleum products. Emissions have been reduced by 1 per cent between 2014 and 2015.

Metal industry was the largest source of SO₂ emissions in 2015 representing 48 per cent of total emissions. Emissions from this sector have been reduced by 54 per cent since 1990, primarily due to reductions in the ferroalloys production. Nonetheless, the production of ferroalloys

remains the most significant source of emissions within this category, being responsible for 71 per cent of emissions from the metal industry in 2015.

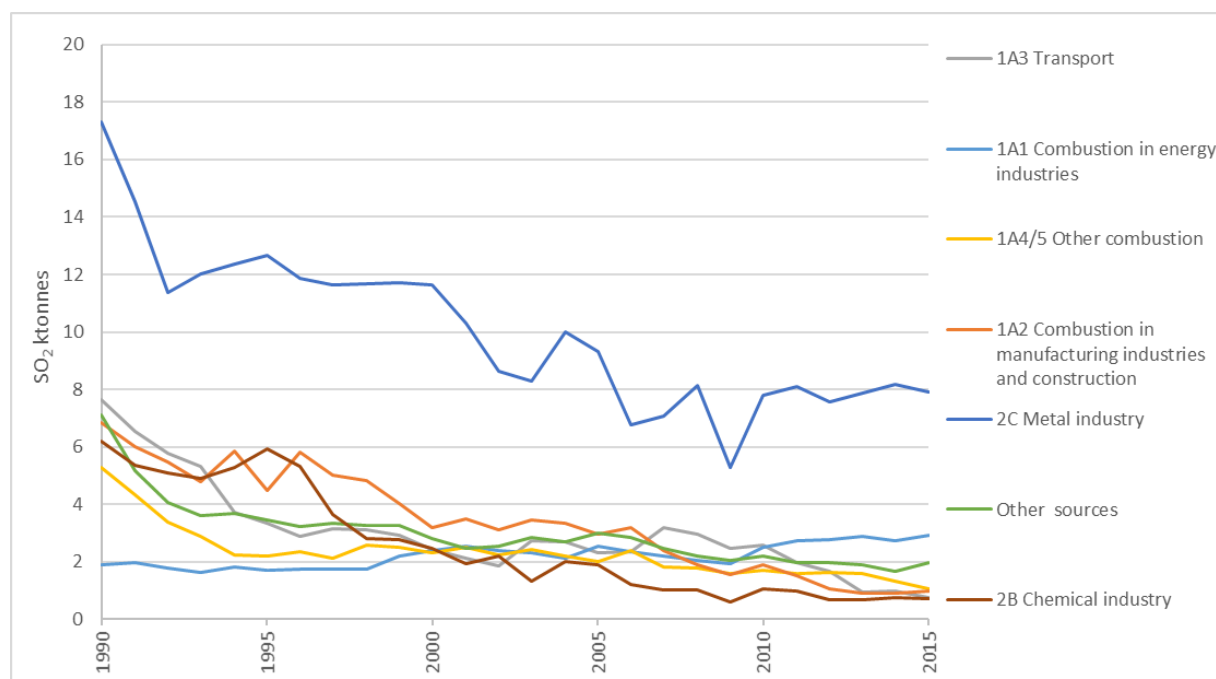


Figure 2.4. Trends in SO₂ emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

Transport constituted 5 per cent of total emissions of SO₂ in 2015. Emissions from transport have been reduced by 90 per cent since 1990, mainly due to less sulphur content in fuels. Most of the reduction took place at the beginning of the period. Indeed, in 1994, emissions were reduced by almost 60 per cent compared to 1990.

Emissions from all transport subcategories have been reduced. Notably, there are currently only significant emissions from national navigation (shipping) in this category, representing, in 2015, 84 per cent of the SO₂ emissions in the transport category.

Emissions from combustions in manufacturing industries and construction have decreased by 86 per cent since 1990, whilst emissions from chemical industries, including carbide production, have decreased with 88 per cent. The reduction is a result of lower production and closure of two plants. In 2015, these two categories contributed to 6 and 4 per cent of total SO₂ emissions, respectively.

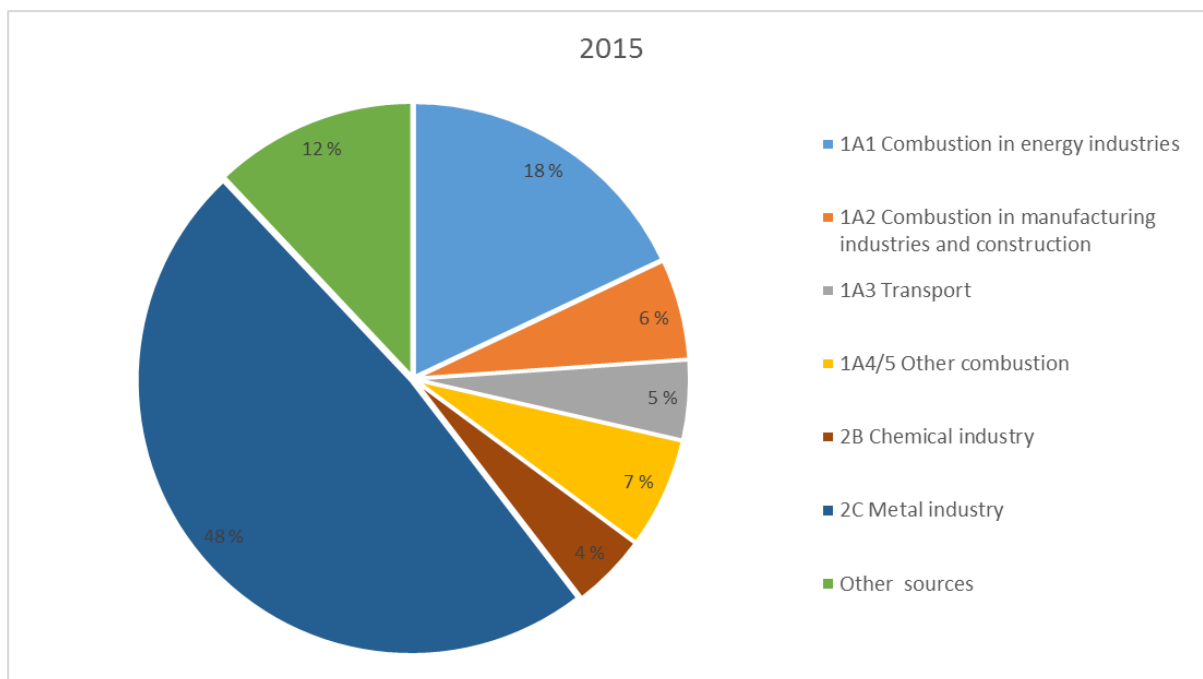


Figure 2.5. Distribution of SO₂ emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.1.4 NH₃

- 2010 commitment of the Gothenburg Protocol: 23 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: an 8 per cent reduction compared to emissions in 2005, 24 600 tonnes.

Norway's NH₃ emissions totaled 26 700 tonnes in 2015. The 2010 commitment of the Gothenburg Protocol for NH₃ emissions has not been fulfilled yet. Despite the fact that the revised commitment for 2020 is higher than the 2010 commitment, it will demand further reductions.

The Norwegian emissions of NH₃ increased by about 12 per cent from 1990 to 2015. From 2014 to 2015, the emissions decreased by 0.5 per cent. The most important reasons for the increase in emissions from 1990 are higher emissions from agriculture and higher emissions from private cars fitted with three-way catalytic converters. One important reason to the increased emissions from agriculture are higher protein content in cattle fodder, which gives higher nitrogen content in the cattle manure.

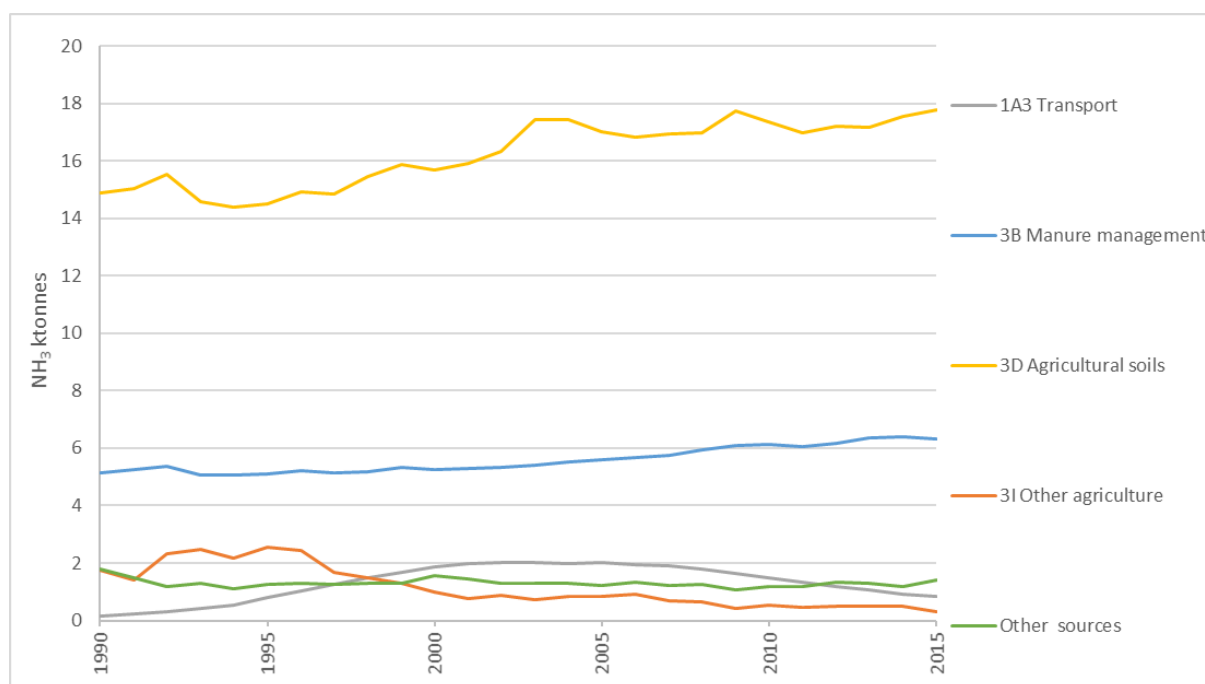


Figure 2.6. Trends in NH₃ emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

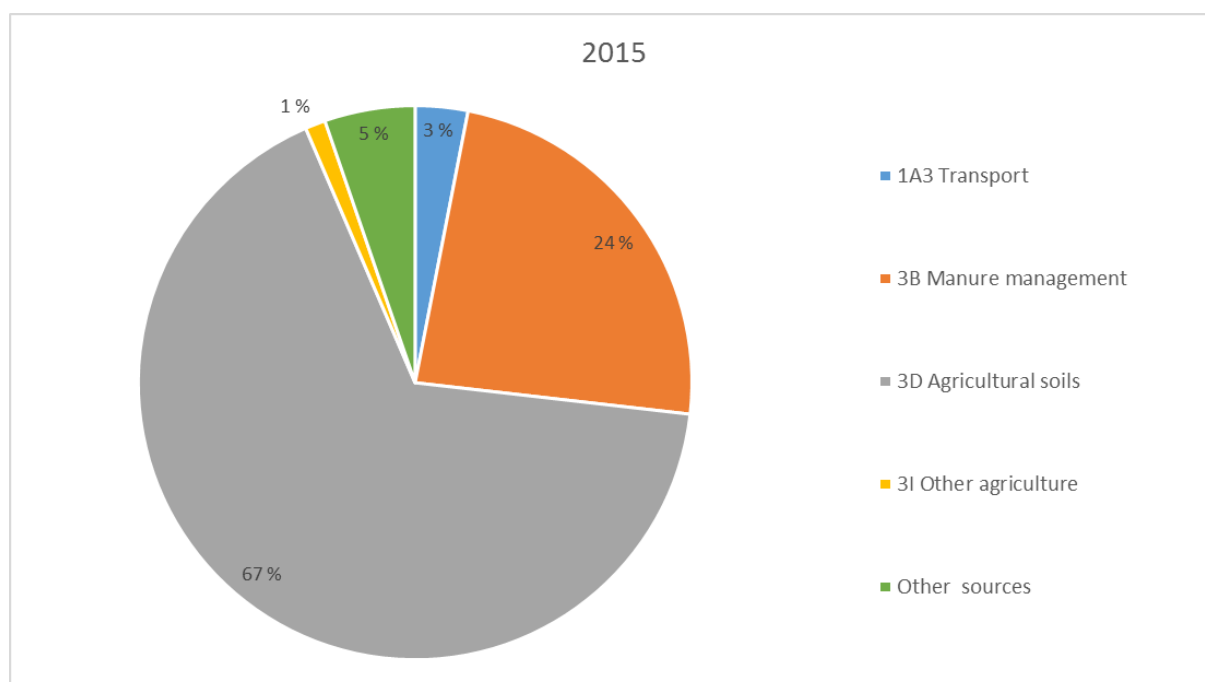


Figure 2.7. Distribution of NH₃ emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

In 2015, agriculture was the source of 92 per cent of the ammonia emissions in Norway. Animal manure is the most predominant source. While 67 per cent of the total Norwegian emissions of NH₃ originated from agricultural soils, 79 per cent of these emissions was the application of animal manure to soils. 24 per cent of total ammonia emissions originated from manure management. Within this category, cattle and swine have been the most important source of

emissions, swine, dairy cattle and non-dairy cattle representing respectively 18 per cent, 22 per cent and 18 per cent of the emissions from this category.

2.1.5 NMVOC

- 2010 commitment of the Gothenburg Protocol: 195 000 tonnes.
- 2020 commitment of the revised Gothenburg Protocol: a 40 per cent reduction compared to emissions in 2005, 137 450 tonnes.

Norway's NMVOC emissions totaled 155 910 tonnes in 2015. The 2010 commitment of the Gothenburg Protocol for NMVOC emissions has been fulfilled since 2008. Nevertheless, the revised 2020 commitment will demand further reductions.

NMVOC emissions have been reduced by 48 per cent since 1990, and by 61 per cent since the peak in 2001. Loading of crude oil offshore was the main reason for the increase in emissions from 1990 to 2001. Measures to prevent these emissions resulted in a 83 per cent decrease in fugitive emissions from fuels from 2001 to 2013. Since 2013, fugitive emissions from fuels have increased by 22 percent because new fields have been in operation. Fugitive emissions from fuels represent 38 per cent of total emissions of NMVOC in 2015.

30 per cent of the NMVOC emissions in 2015 originated from product use (solvents etc.), the most important subcategories being "other solvent use" (2D3i), 55 per cent, and domestic solvent use including fungicides, 29 per cent of emissions within this category. The category "other combustion" (NFR 1A4 and 1A5) was responsible for 13 per cent of total emissions of NMVOC in 2015. The two most significant sources of emissions are household and gardening (mobile) and stationary plants in the residential sector. These two subcategories were responsible for 92 per cent of emissions within this category. Emissions from household and gardening have remained stable since 1990, whilst emissions from residential, stationary plants show more interannual variations. NMVOC emissions from these two categories have been reduced by 1 per cent since 1990 and by 3 per cent since 2014.

NMVOC emissions from transport have decreased by 86 per cent since 1990, mainly due to reductions in emissions from passenger cars and gasoline evaporation. Stricter emission standards for petrol passenger cars were implemented in 1989, and these have led to reduced emissions. In addition, the increased share of diesel cars within the vehicle fleet has strengthened the downwards trend.

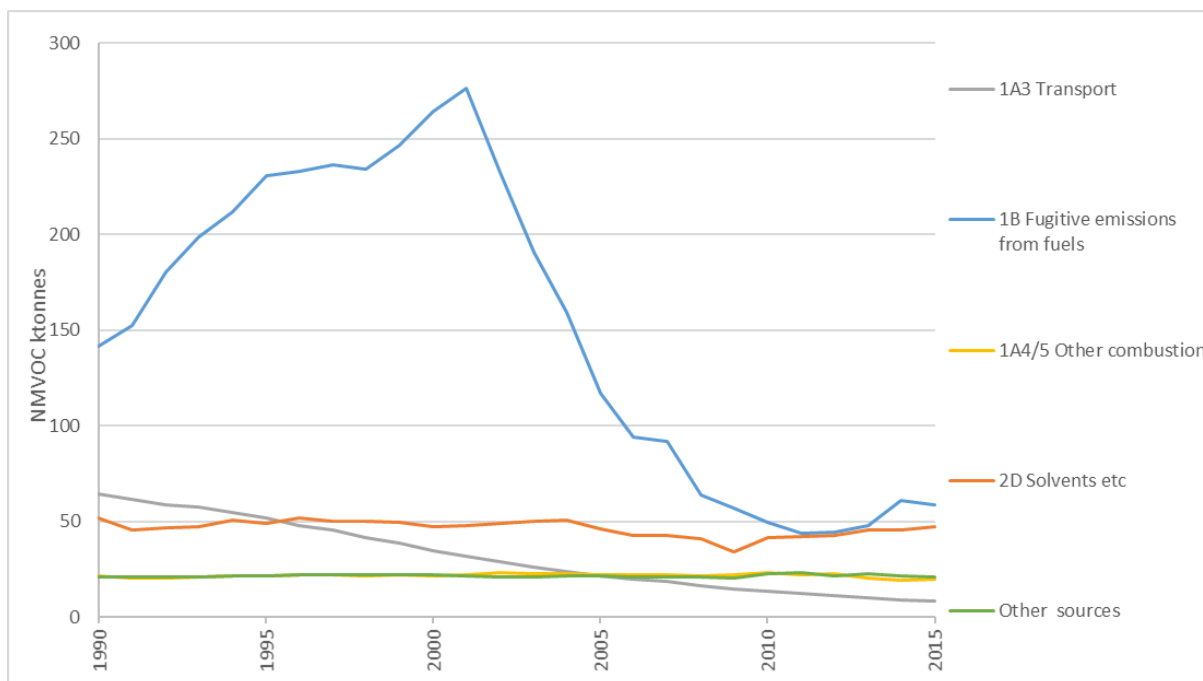


Figure 2.8. Trends in NMVOC emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

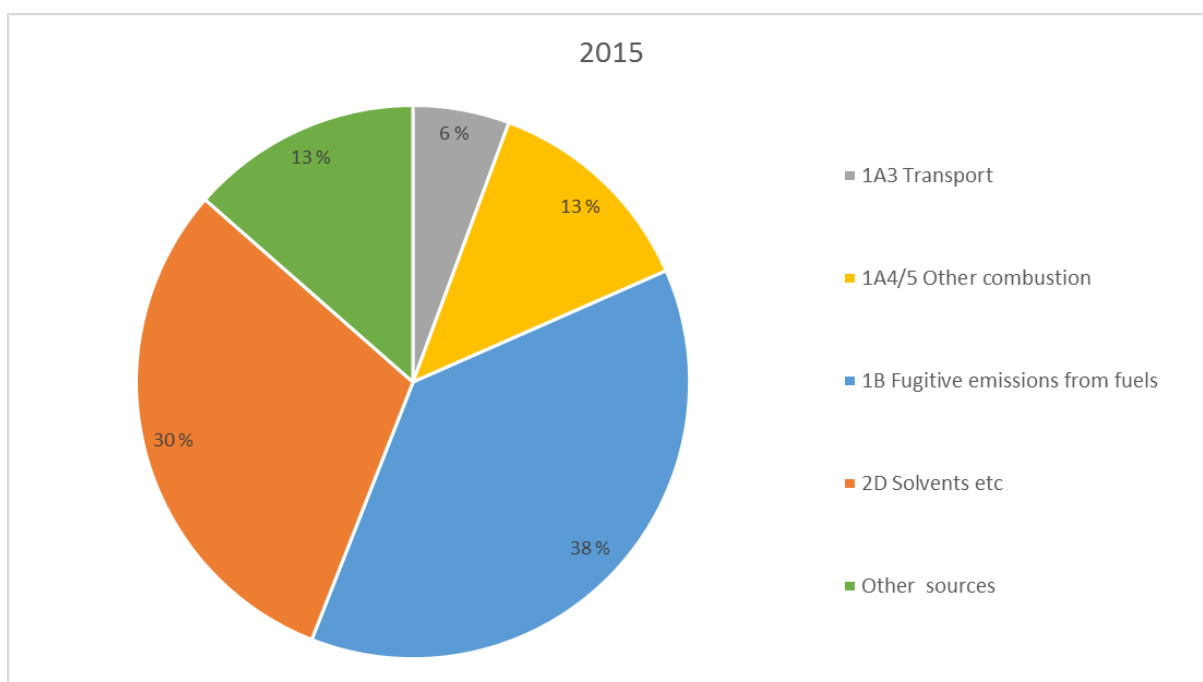


Figure 2.9. Distribution of NMVOC emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.2 CO

Emissions from carbon monoxide, CO, amounted to 382 500 tonnes in 2015. They have been reduced by 54 per cent since 1990 and by 4 per cent since 2014. This is mainly due to reductions in emissions from transport, which have been reduced by 88 per cent since 1990. Emissions from passenger cars represented almost 85 per cent of the transport category in 1990. They have been reduced by 91 per cent since 1990, primarily due to stricter emission standards. Emissions from light duty vehicles have also been significantly reduced (93 per cent since 1990) albeit from a lower absolute level.

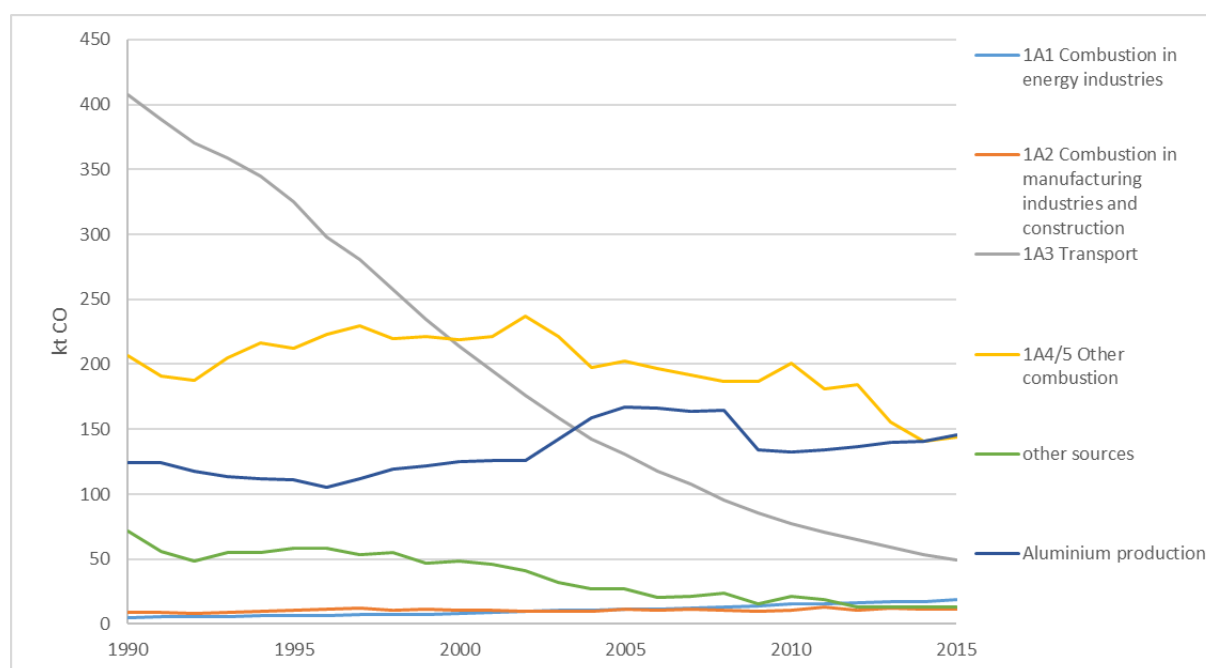


Figure 2.10. Trends in CO emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

The categories "other combustion" (NFR 1A4 and 1A5) and aluminium production were the largest sources of emissions in 2015, representing both 38 per cent of the CO emissions. Emissions from "other combustion" originated primarily from the NFR categories "Residential: Stationary plants" and "Residential: Household and gardening (mobile)", which accounted for 64 and 28 per cent of the total emissions of the category, respectively, in 2015. Emissions from "Residential: Stationary plants" were at a peak in 2002 but have decreased since. Emissions were 39 per cent lower in 2015 than in 1990. These emissions are mainly due to wood combustion for heating purposes. Emissions from "Residential: Household and gardening (mobile)" have remained stable since 1990. Emissions from aluminium production increased by 17 per cent since 1990 while production increased by 40 per cent.

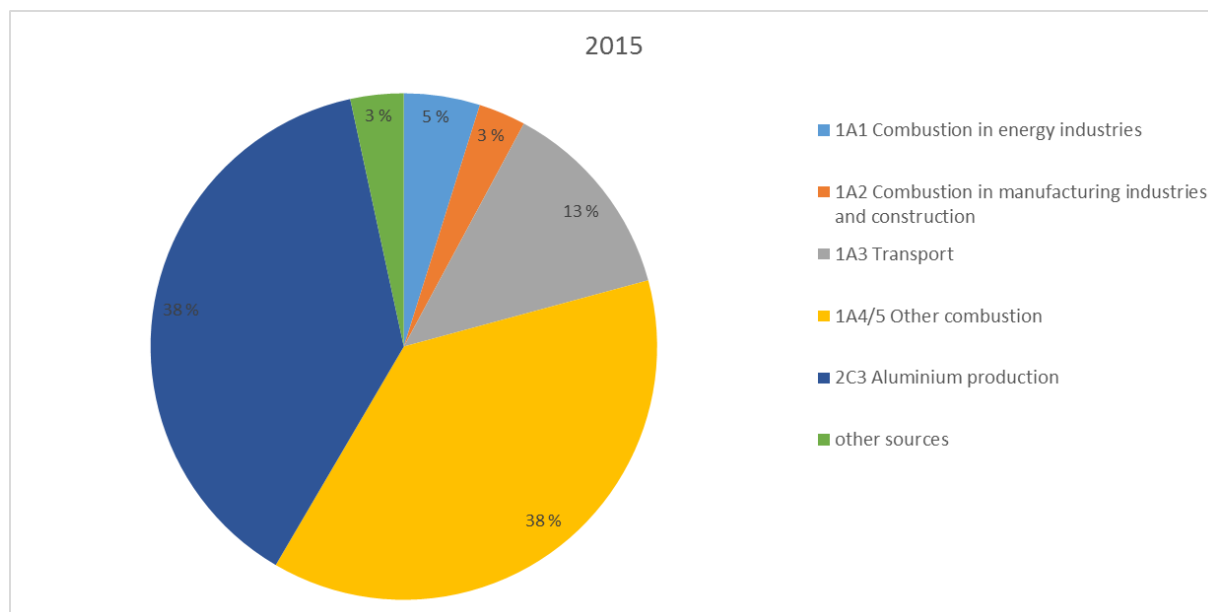


Figure 2.11. Distribution of CO emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3 PM, POPs and heavy metals

Emissions of all heavy metals except copper, PAH-4, dioxins, HCB and PCB, although fluctuating throughout the period, shows a general downward trend since 1990.

Since 2014, PCB emissions have decreased, HCB emissions have increased while PAH-4, dioxins and black carbon have remained stable.

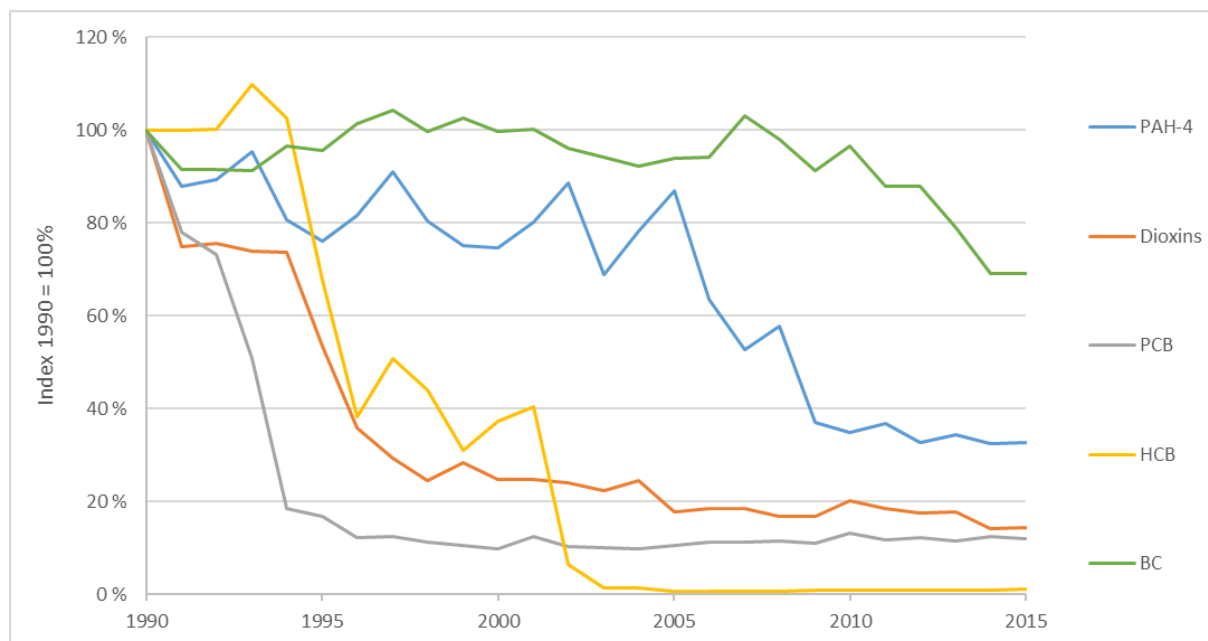


Figure 2.12. Trends in emissions for PAH-4, dioxins, PCB, HCB and BC. 1990-2015. Index 1990=100%.

Source: Statistics Norway/Norwegian Environment Agency

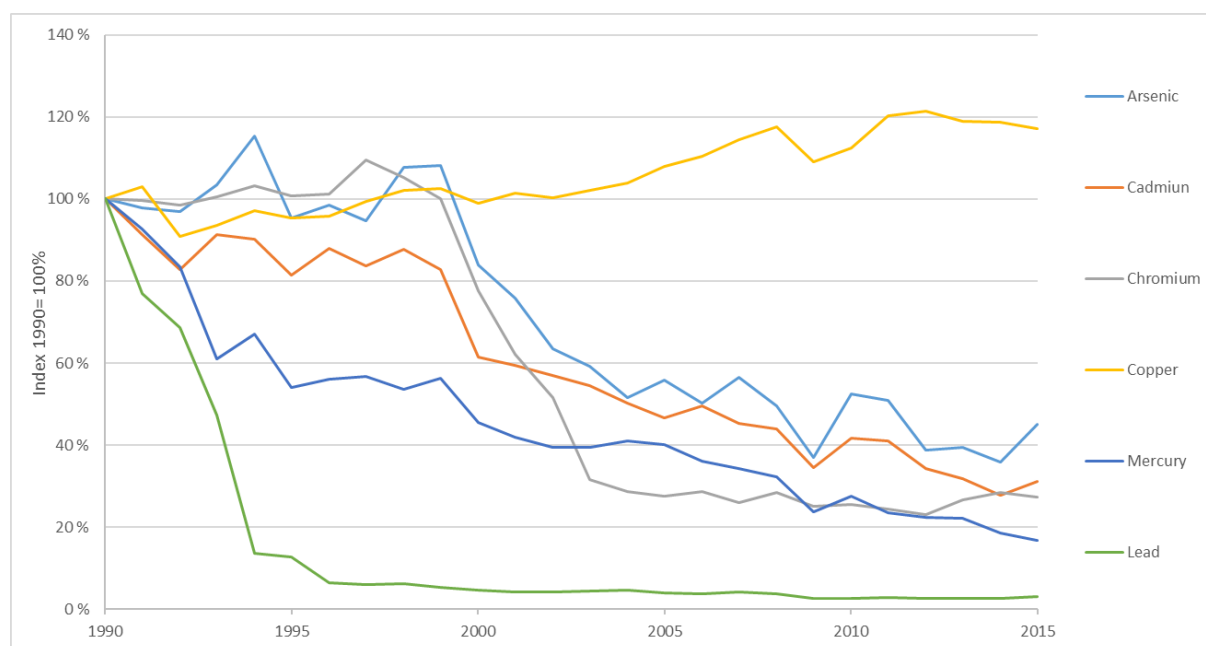


Figure 2.13. Trends in emissions for Heavy metals. 1990-2015. Index 1990=100%

Source: Statistics Norway/Norwegian Environment Agency

The emissions of several hazardous substances including PAHs (polycyclic aromatic hydrocarbons) have been considerably reduced since 1990. The cause of the decrease is primarily reduced emissions within manufacturing and mining. Cleaning measures, changes in production processes and lower activity in some industries have all resulted in reduced emissions. Norway has set a national target to reduce emissions compared with the emission levels in 1995. Internationally, Norway has an obligation to reduce emissions of selected hazardous substances compared to the level of emissions in 1990 through the Aarhus Protocol.

2.3.1 PM₁₀

Emissions of particulate matter (PM₁₀) totaled 37 100 tonnes in 2015. They have decreased by 29 per cent since 1990 and increased by 1 per cent since 2014.

The most important source of emissions is stationary residential plants (included in “other combustion” (NFR 1A4 and 1A5)), which amounted to 16 500 tonnes of PM₁₀ in 2015. Emissions have been reduced by 24 per cent since 1990 but increased by 6 per cent since 2014. Wood burning is the largest source of emissions from this subcategory. The wood consumption increased by 7 per cent from 2014 to 2015, mainly due to a colder winter. Since 2000, the reduction in particle emissions has been higher than reductions in wood consumption due to an increased share of new technology in wood burning appliances.

Process emissions from manufacturing and mining amounted to 8 200 tonnes in 2015. Within the process sector, the largest sources were mineral industry and metal industry, which accounted for 42 and 38 per cent of the process category, respectively, in 2015. Emissions from metal industry have decreased by 53 per cent since 1990 mainly due to reduced productions. Emissions from mineral industry have increased by 12 per cent since 1990. Emissions from this

category are mainly from sandpit and rock-crushing plants. It should be noted that the emissions from this source varies highly from year to year, and that the data are uncertain. Transport was responsible for 10 per cent of PM₁₀ emissions in 2015, of which 53 per cent originated from road abrasion and tyre wear.

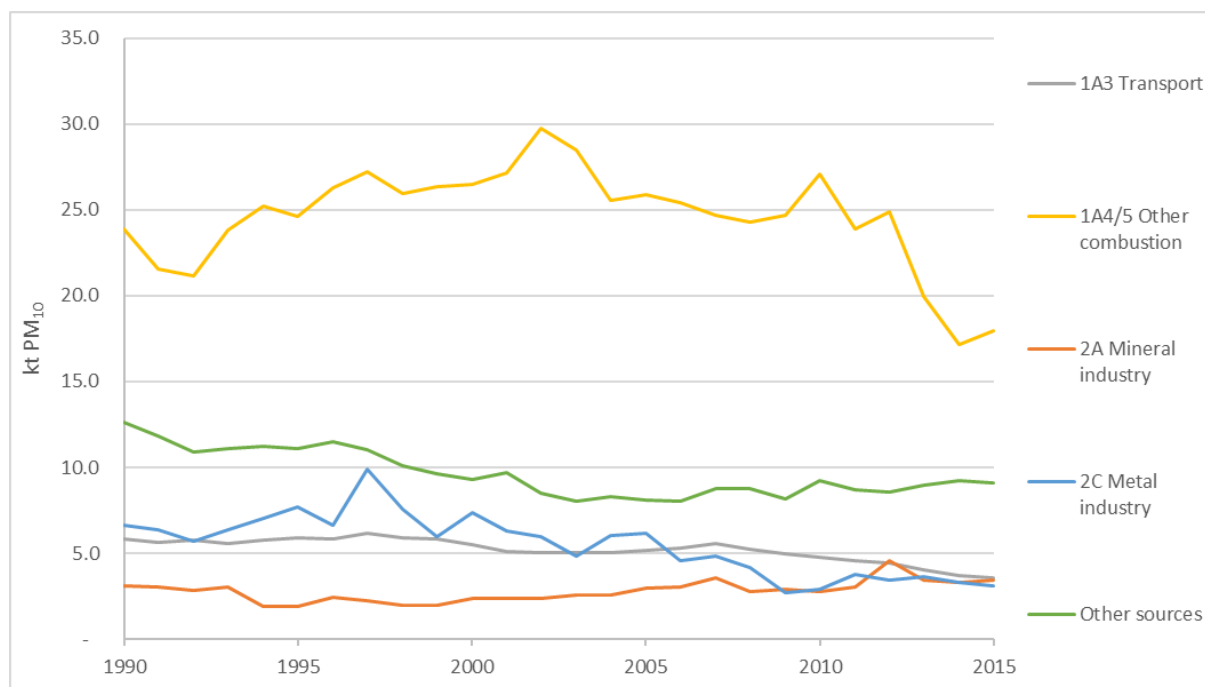


Figure 2.14. Trends in PM₁₀ emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

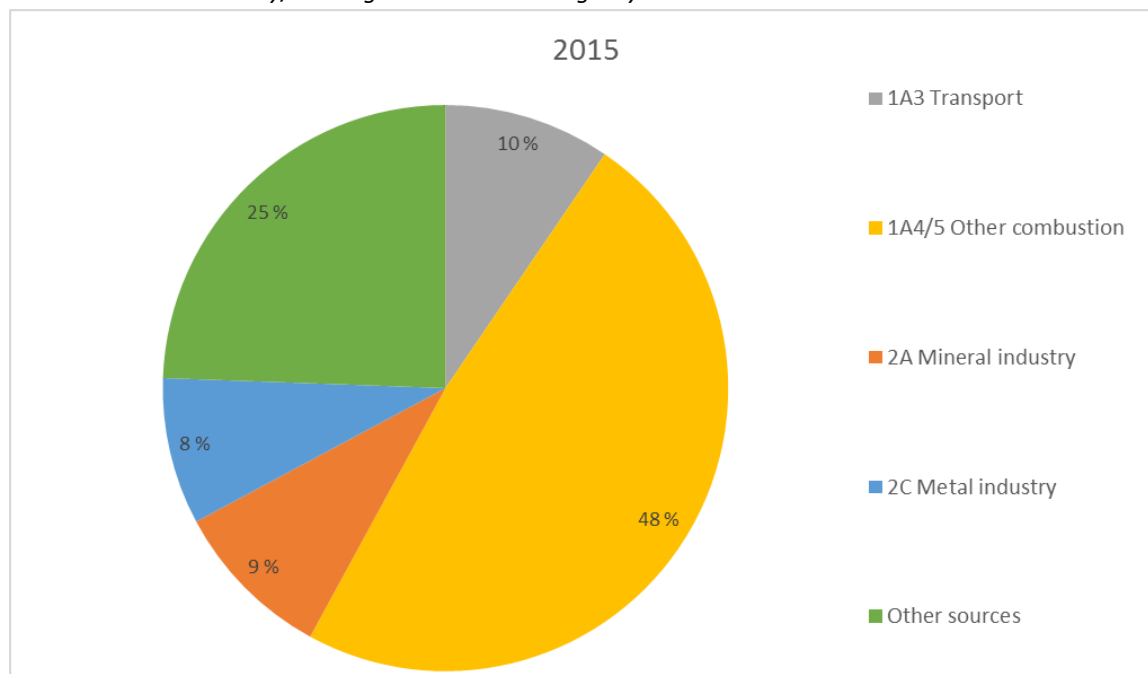


Figure 2.15. Distribution of PM₁₀ emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.2 PM_{2.5}

- 2020 commitment of the revised Gothenburg Protocol: a 30 per cent reduction compared to emissions in 2005, 27 300 tonnes.

Emissions of PM_{2.5} follows the same trend as PM₁₀ emissions. Norway's emissions totaled 28 100 tonnes in 2015. The 2020 commitment of the revised Gothenburg Protocol has almost been met.

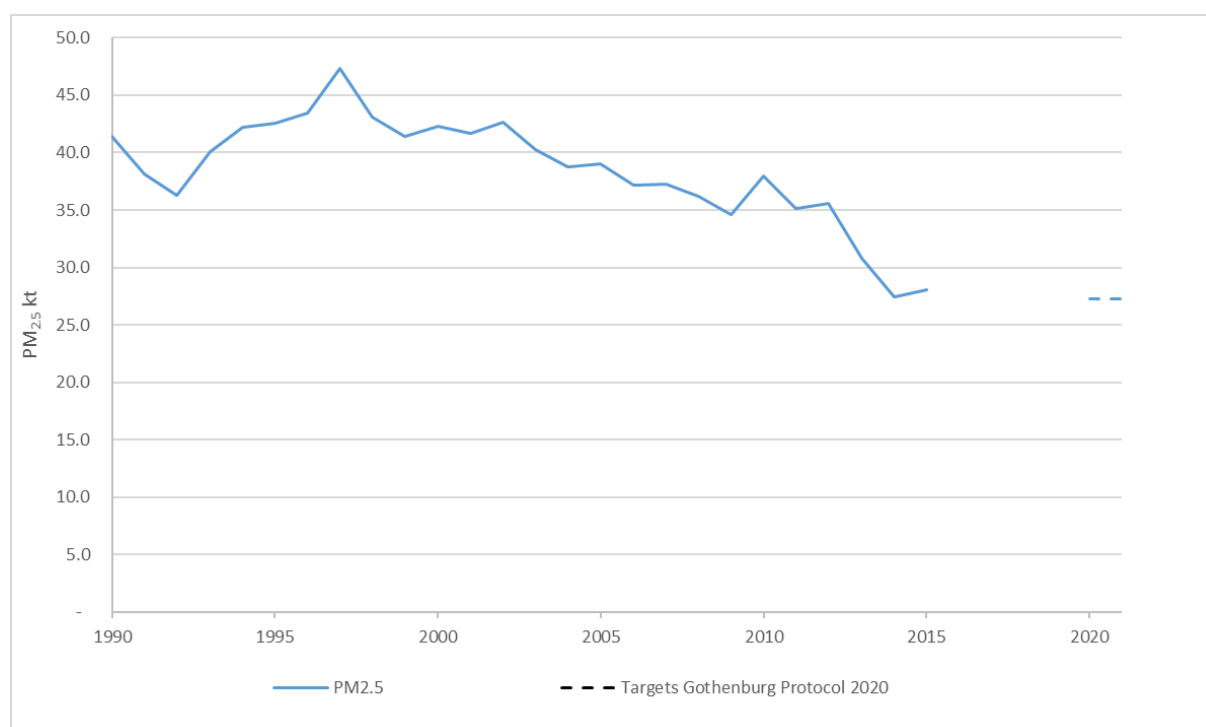


Figure 2.16. Trends in PM_{2.5} emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

2.3.3 Black carbon

The emissions of BC amounted to 3 300 tonnes in 2015, a total reduction of 31 per cent since 1990 and of 4 per cent since 2014.

In 2015, the most important source of emissions was “other combustion” (NFR 1A4 and 1A5), contributing to 40 per cent of the total emissions. From this category, 68 per cent of emissions originated in 2015 from residential stationary plants, primarily due to wood combustion in private households. Despite yearly variations, the emissions of BC from residential combustion increased between 1990 and 2010 and have only been reduced since 2010. Emissions increased by 35 per cent between 1990 and 2010 and decreased by 67 per cent since 2010.

In 2015, the second most important source of emissions was transport. It contributed to 28 per cent of the total BC emissions. The greatest share of emissions within the transport sector stem from shipping, passenger cars and light duty vehicles, contributing to 36, 24 and 19 per cent of the emissions in 2015, respectively. From 1990 to 2015, emissions from shipping have been

reduced by 36 per cent, while emissions from passenger cars have increased by 130 per cent. Emissions from light duty vehicles have been reduced by 29 per cent since 1990.

Combustion in energy industries, which, in 2015, comprised 11 per cent of the total BC emissions, have increased by 242 per cent since 1990. The most prominent source of emissions within this category is manufacture of solid fuels and other energy industries. It represented 64 per cent of the sector emissions in 2015. During the period from 1990 to 2015, there has been an increase of almost 150 per cent of BC emissions from this industry.

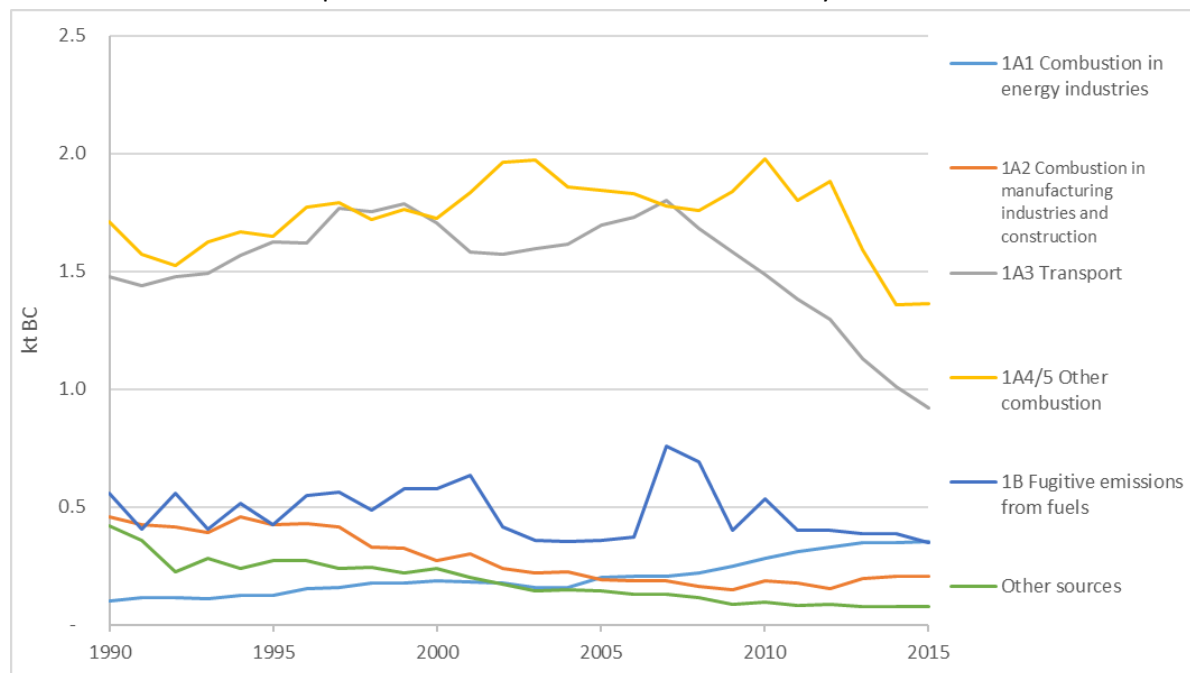


Figure 2.17. Trends in BC emissions, 1990-2015. 1000 tonnes

Source: Statistics Norway/Norwegian Environment Agency

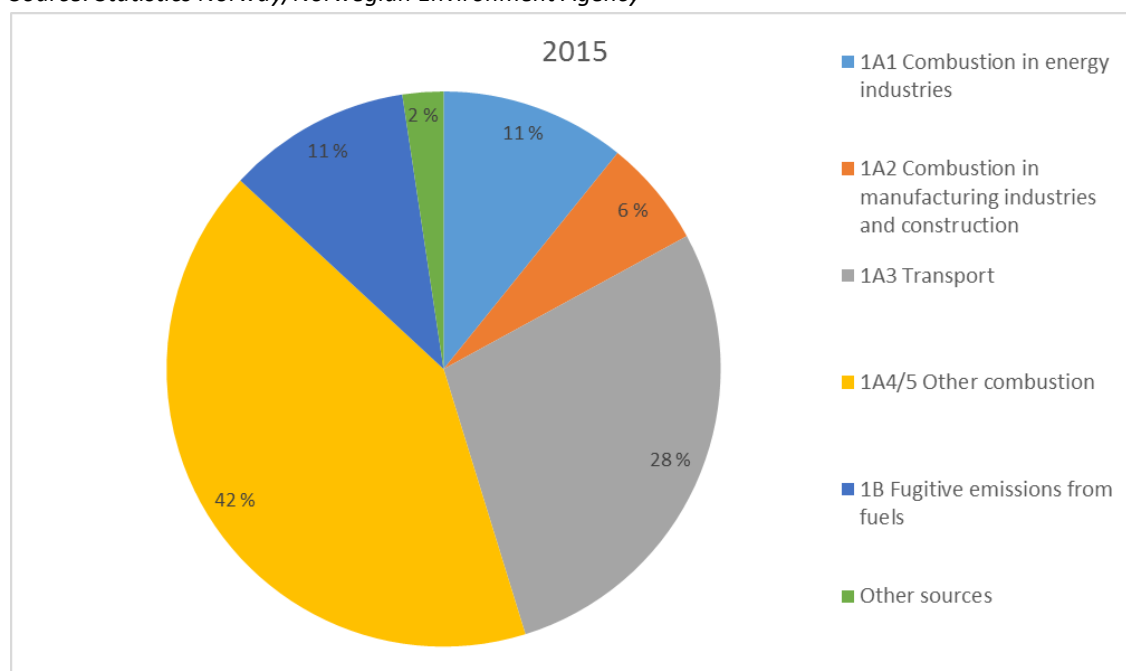


Figure 2.18. Distribution of BC emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.4 Dioxins

In 2015, 17.2 grams of dioxins were emitted. Since 1990, emissions of dioxins have decreased by 86 per cent, including a 19 per cent reduction since 2013. Emissions have been stable between 2014 and 2015. A large proportion of this reduction is due to the closure of industrial plants and mines. In addition, emissions from energy industries were reduced by 86 per cent from 1990 to 2005 due to the introduction of cleaning measures at waste incineration plants. Since 2005, the emissions have increased, mainly due to increased activity in the oil and gas production.

From 1990 to 1996, the largest source of dioxins emissions was the category other industrial processes (2H) due to an ore mine with high dioxins emissions which has been closed down in 1996. Since the closure, dioxins emissions from the source category 2H have been reduced to zero.

In 2015, the largest source of dioxins emissions was the category “other combustion” (NFR 1A4 and 1A5) contributing to 43 per cent. Combustion in private households contributed to 78 per cent to the total dioxins emissions of this category in 2015, primarily due to wood burning. National fishing, which is also included in this category, contributed to 18 per cent to the category.

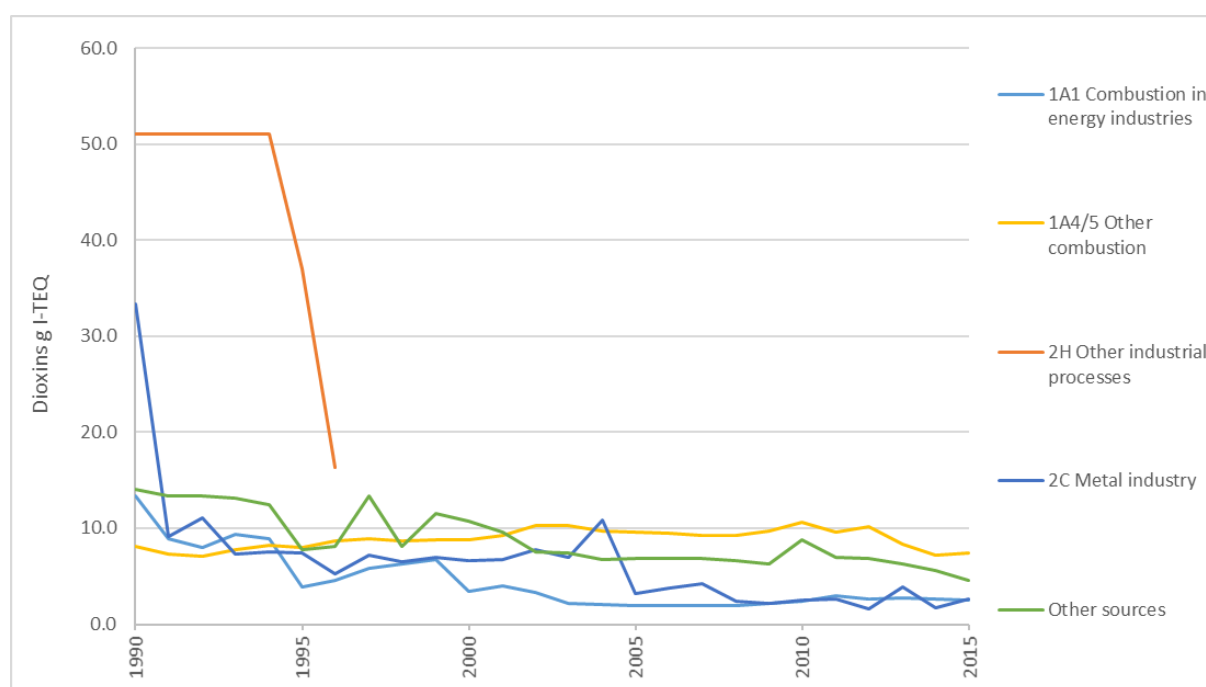


Figure 2.19. Trends in dioxins emissions, 1990-2015. Gram I-TEQ

Source: Statistics Norway/Norwegian Environment Agency

Dioxins emissions from combustion in energy industries were responsible for 15 per cent of total emissions of dioxins in 2015. There has been a significant decrease in emissions from public electricity and heat production; emissions in 2015 were 94 per cent lower than in 1990. Emissions from manufacture of solid fuels and other energy industries became the largest source of emissions within this category in 2004 and has remained so since. In 2015, it was

responsible for 67 per cent of dioxins emissions within the energy industries category. Emissions from manufacture of solid fuels and other energy industries have increased by 158 per cent since 1990.

Transport was responsible for 11 per cent of total dioxins emissions in 2015. National navigation (shipping) is by far the most significant source of emissions within this category, representing 82 per cent of the transport emissions. Emissions from passenger cars, which contributed to almost half of emissions within this category in 1990, declined rapidly from 1990 to 1994 and further to 1996. Since then, they have been amounted to approximately the same absolute level. In 2015, emissions from passengers cars were 91 per cent lower than in 1990.

Emissions from other waste, which accounted for 8 per cent of the total emissions in 2015, have been reduced by 31 percent since 1990 including a reduction of 37 per cent since 2014. Process emissions from metal production accounted 15 per cent of the total emissions of dioxins in 2015.

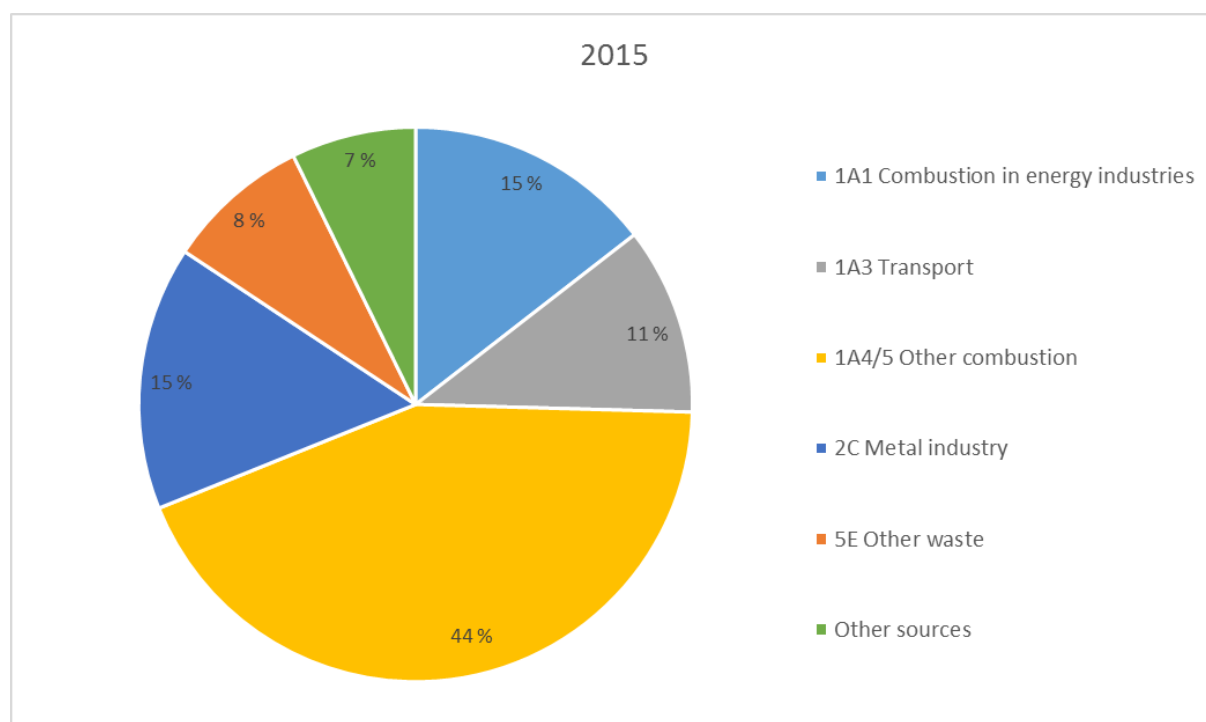


Figure 2.20. Distribution of dioxins emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.5 PAH-4

The present emission inventory for polycyclic aromatic hydrocarbons (PAH) includes four PAHs: benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene. The total emissions of PAH-4 in 2015 amounted to approximately 6.6 tonnes, which is a reduction of 67 per cent since 1990. Emissions increased by 1 per cent since 2014.

In 2015, benzo(b)fluoranthene contributed to 44 per cent of PAH-4 emissions while benzo(k)fluoranthene, benzo(a)pyrene and indeno(1,2,3-cd)pyrene contributed to 21, 20 and 14 per cent respectively.

per cent, respectively.

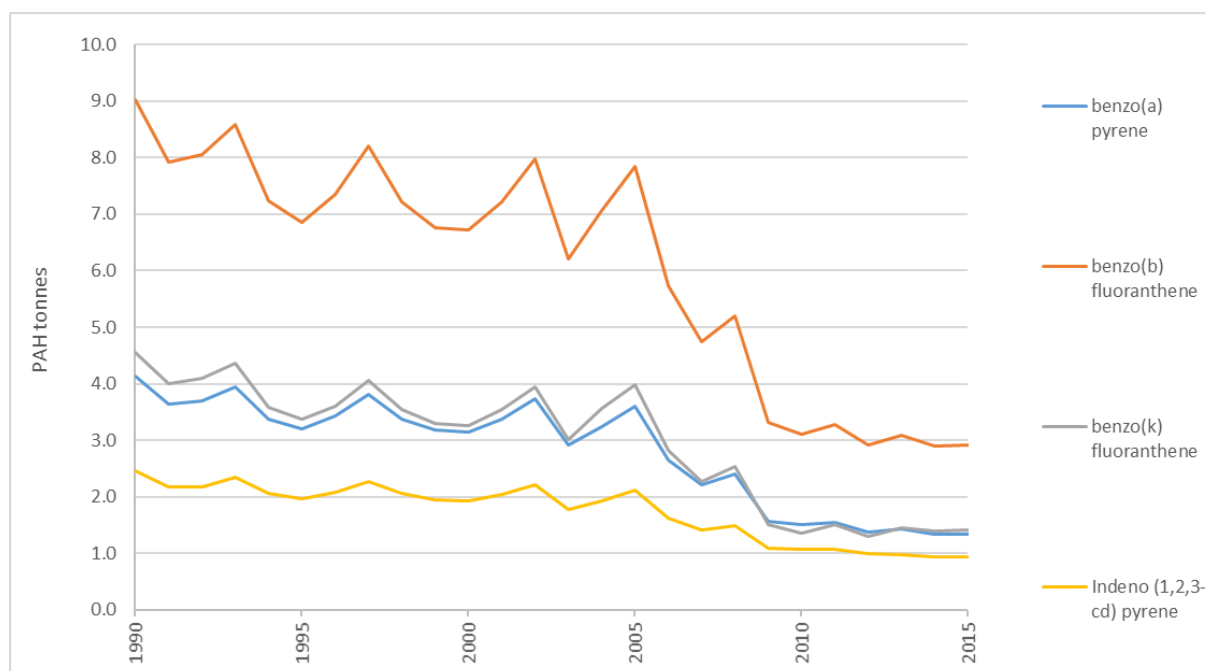


Figure 2.21. Trends in PAH emissions, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene, 1990-2015. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

Process emissions in aluminium production is the most dominant source of PAH-4 emissions. It contributed to 70 per cent to the total PAH-4 emissions in 1990 and with 45 per cent in 2015. The PAH-4 emissions decreased primarily because of the discontinuation of Soederberg technology in the aluminium production. Emissions from aluminium production have been reduced by 79 per cent since 1990.

Road traffic contributed to 22 per cent of the emissions in 2015. This includes both exhaust and tyre and brake wear. There is a general trend with increasing emissions from transport since 1990, emissions from passenger cars (without tyre and brake wear) have increased by 50 per cent, emissions from light duty vehicles have increased by 219 per cent whilst emissions from heavy duty vehicles have increased by 25 per cent.

Wood burning is by far the most significant source of emissions within the category “other combustion” (NFR 1A4 and 1A5). It contributed to 18 per cent of PAH emissions in 2015. Emissions from residential; stationary plants have been reduced by 58 per cent since 1990, due to warmer winters and the increasing share of new technology in wood burning appliances.

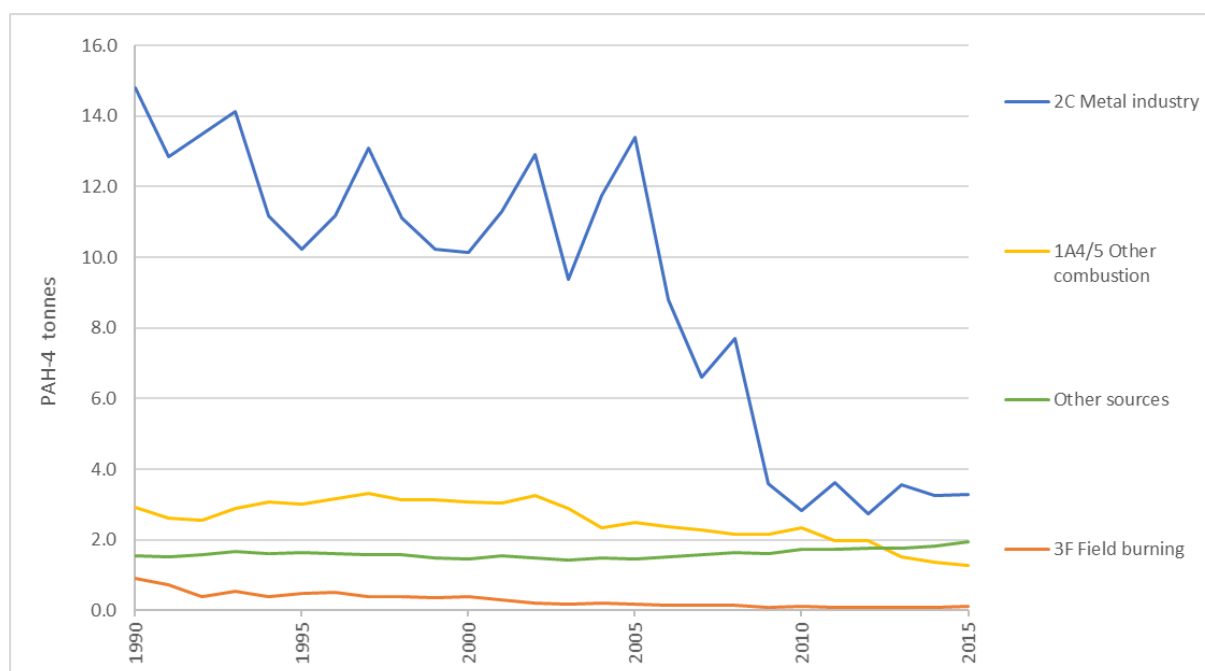


Figure 2.22. Trends in total PAH-4 emissions, 1990-2015. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

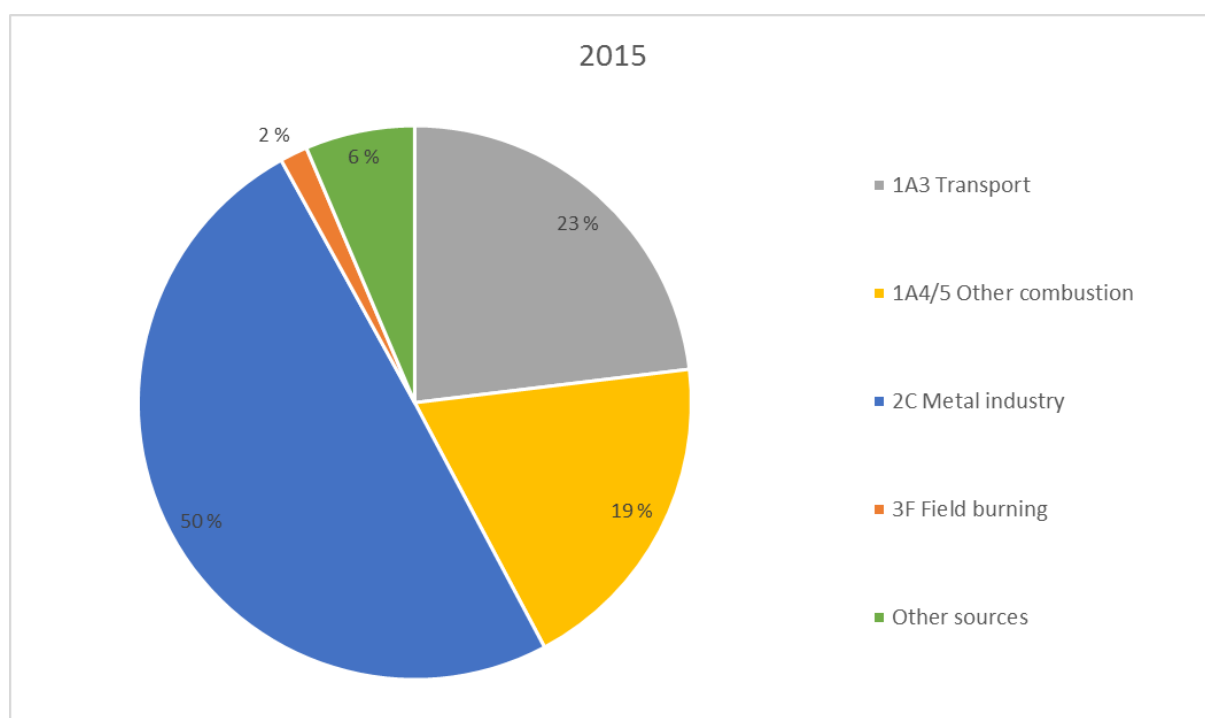


Figure 2.23. Distribution of total PAH-4 emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.6 HCB

Estimated HCB emissions in Norway amounted to 1.4 kilograms in 2015, and has decreased by 99 per cent since 1990. Emissions have increased by 49 per cent since 2005, but this amount to a small increase in absolute numbers after a decrease of 99 per cent between 1990 and 2005.

Emissions decreased mainly due to the closure of magnesium production which contributed to almost 99 per cent of total HCB emissions in 1990.

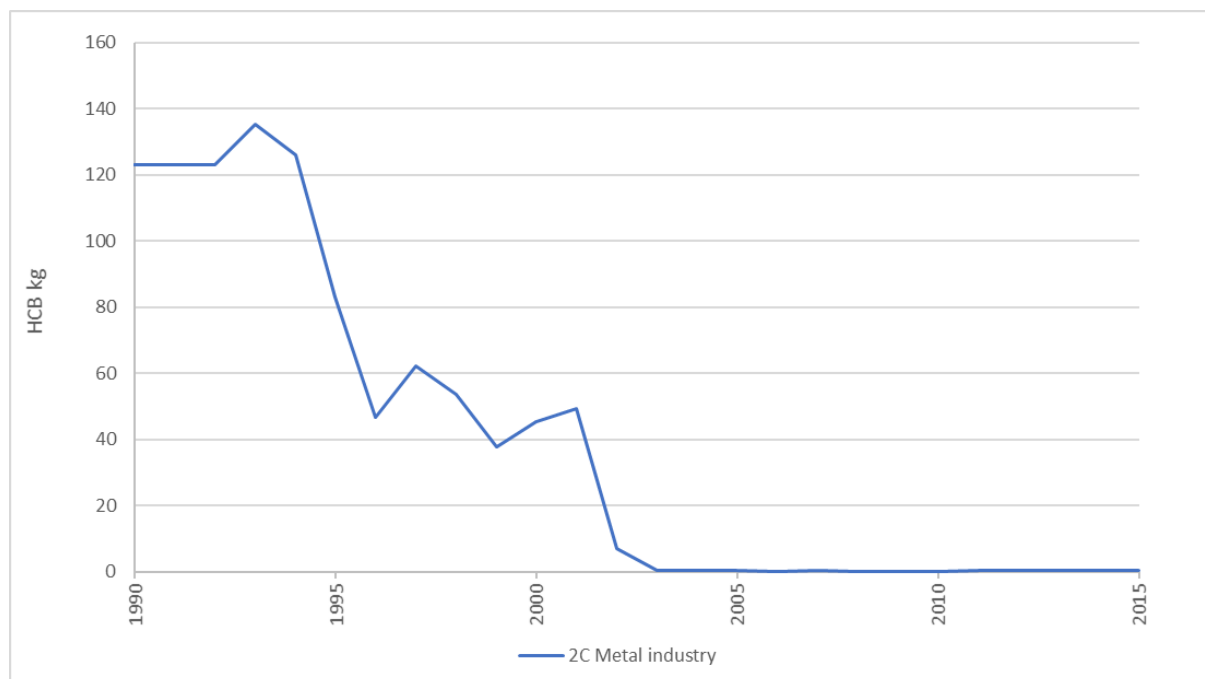


Figure 2.24. Metal industry, trends in HCB emissions, 1990-2015. Kilogram

Source: Statistics Norway/Norwegian Environment Agency

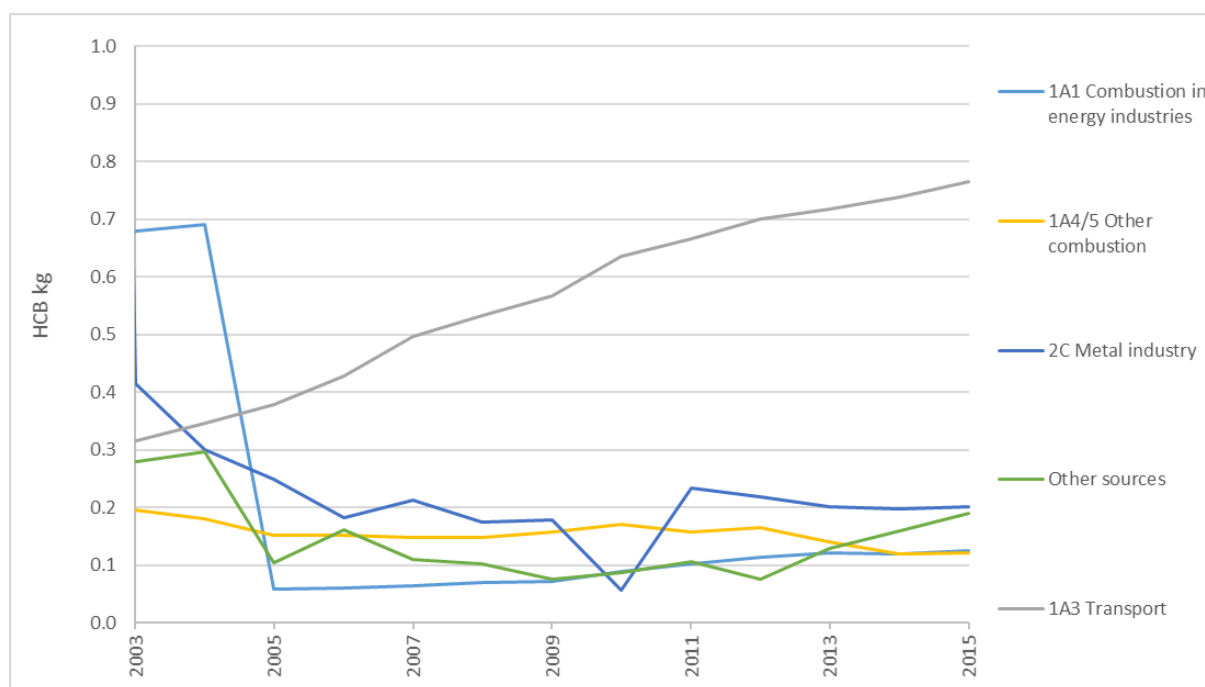


Figure 2.25. Trends in total HCB emissions, 2003-2015. Kilogram

Source: Statistics Norway/Norwegian Environment Agency

The most important source of emissions of HCB in 2015 was road transport, which contributed to 52 per cent to total emissions. Emissions from road transport have increased significantly

since 1990, mainly due to increased traffic activity. For instance, HCB emissions from passenger cars were more than fourteen times higher in 2015 than in 1990.

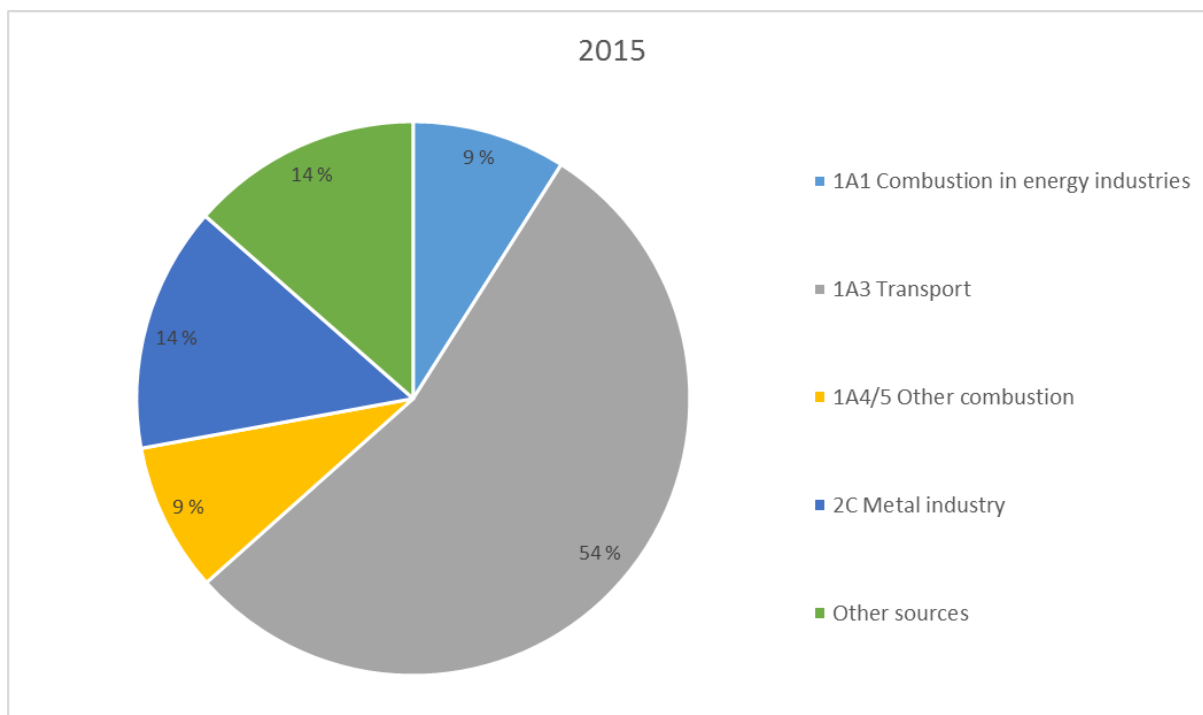


Figure 2.26. Distribution of HCB emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.7 PCB

Estimated PCB emissions in Norway amounted to 25 kilograms in 2015. Emissions have increased by 20 per cent since 2000, after a decrease of 90 per cent between 1990 and 2000. From 1990 to 1994, emissions from transport decreased by 86 per cent, due to fuel content regulations. Emissions from passenger cars, which accounted for 79 per cent of the total PCB emissions in 1990, decreased from 167 to 1 kilogram from 1990 to 2015.

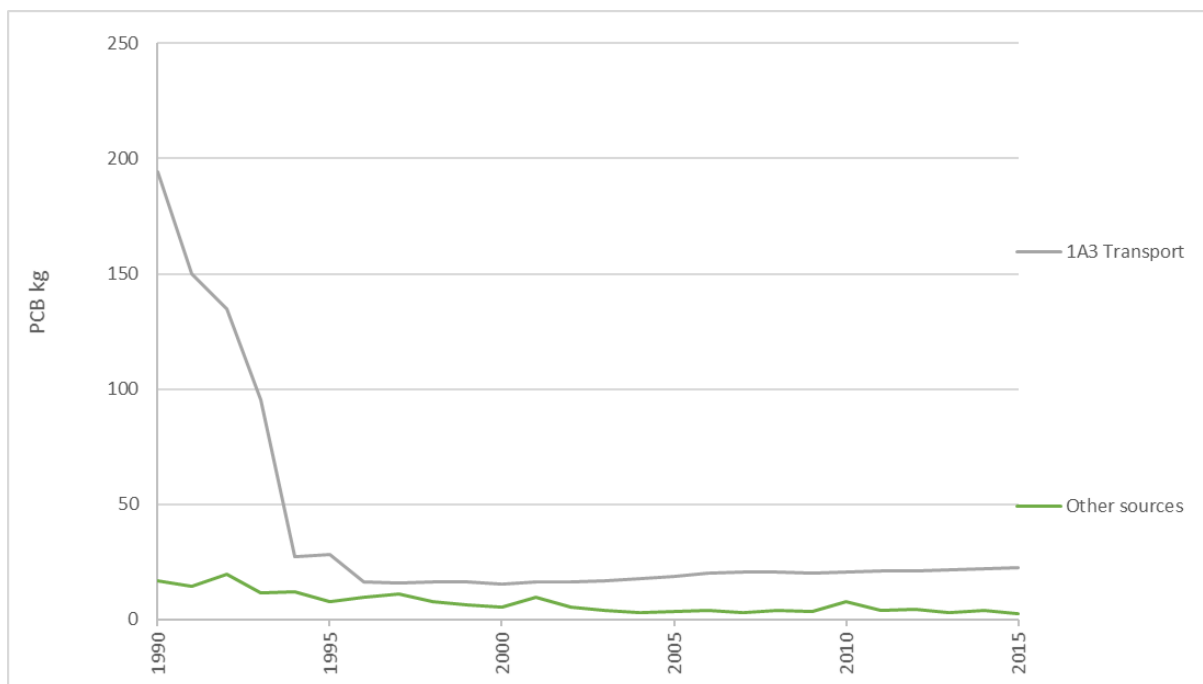


Figure 2.27. Trends in total PCB emissions, 1990-2015. Kilogram

Source: Statistics Norway/Norwegian Environment Agency

Despite large reductions, road transport stayed the most important source of emissions of PCB in 2015. It contributed to 89 per cent of total emissions.

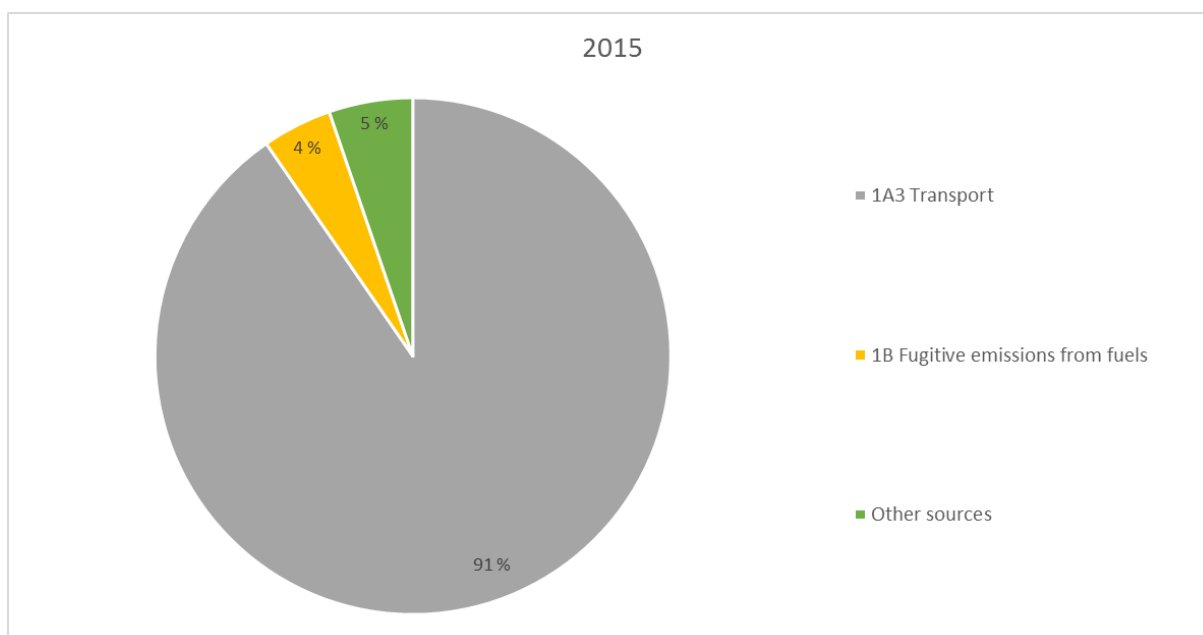


Figure 2.28. Distribution of PCB emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.8 Lead

Lead emissions totaled 5.6 tonnes in 2015 and have been reduced by 97 per cent since 1990. Regulations on lead content in fuels are the main reason for this reduction. Indeed, emissions from passenger cars constituted 82 per cent of the total in 1990, and only 2 per cent in 2015. Emissions of lead have been relatively constant in recent years.

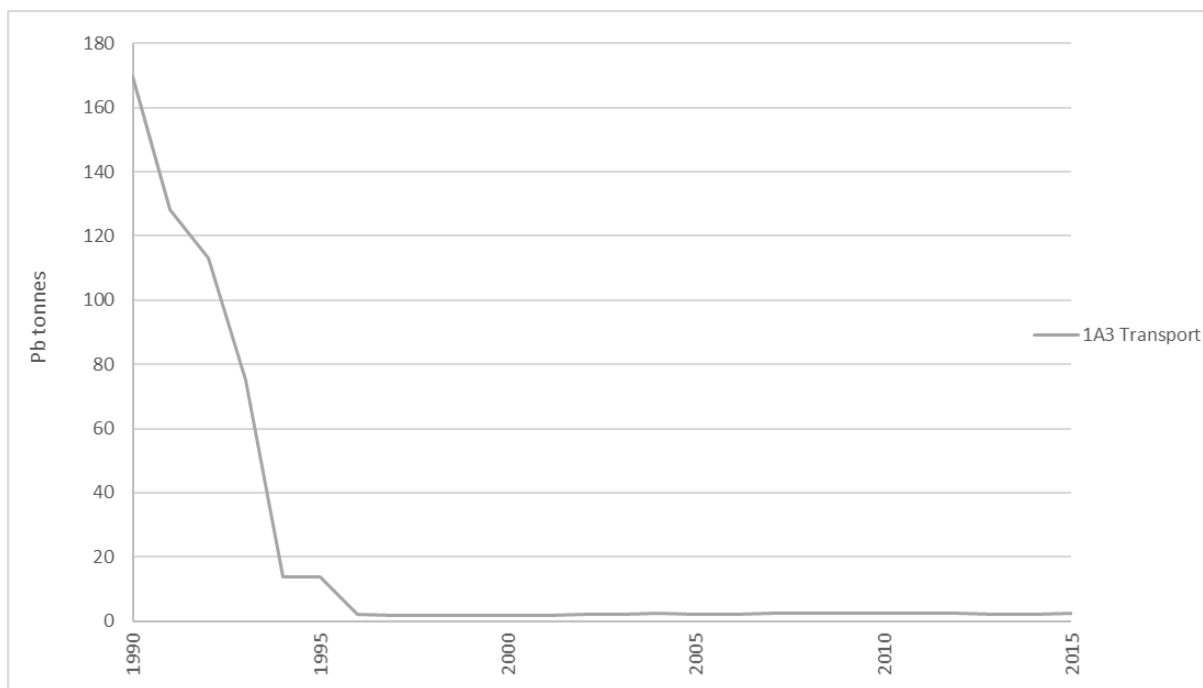


Figure 2.29. Transport, trends in lead emissions, 1990-2015. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

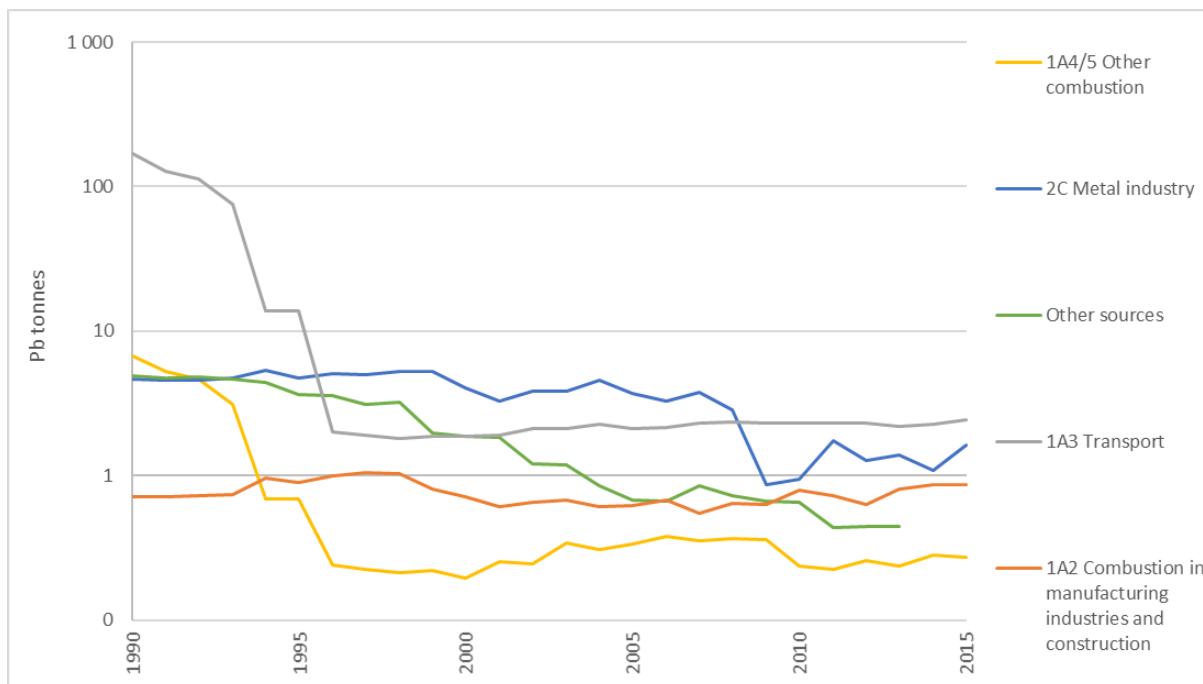


Figure 2.30. Trends in lead emissions, 1996-2015. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

Transport has remained the largest source of lead emissions, but since 1996 tyre and brake wear has been the most important source within the transport sector, being responsible for 30 per cent of lead emissions in 2015.

In 2015, process emissions from metal industry and combustion activities within manufacturing industries and construction emitted 29 and 15 per cent of the total lead, respectively.

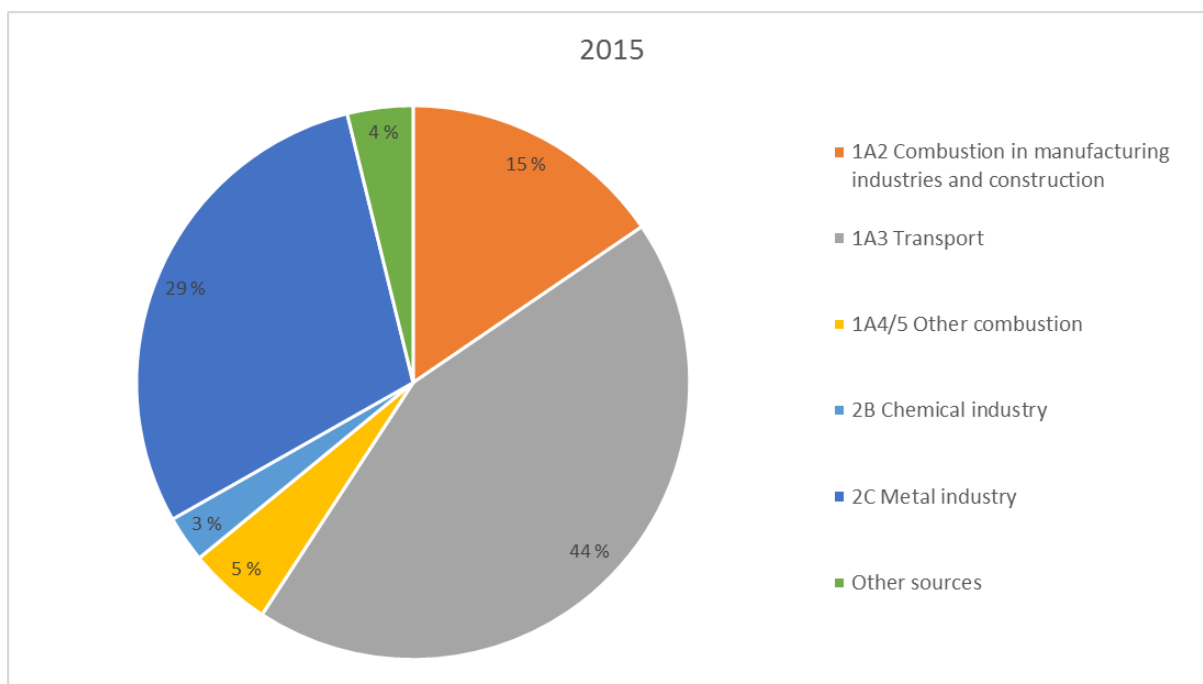


Figure 2.31. Distribution of lead emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.9 Cadmium

Emissions of cadmium totaled 0.5 tonnes in 2015, representing a 12 per cent increase from 2014, but a 69 per cent reduction from 1990. Emissions have primarily been reduced in manufacturing industries and field burning.

Process emissions from production of iron, steel and ferroalloys have been reduced due to reduction efforts and closing down of production plants. Metal industry were responsible for 15 per cent of cadmium emissions in 2015, compared to 36 per cent in 1990.

Cadmium emissions from field burning have been significantly reduced from 1990 to 2015. In 2015, it contributed to 9 per cent of total Norwegian cadmium emissions, compared to 24 per cent in 1990.

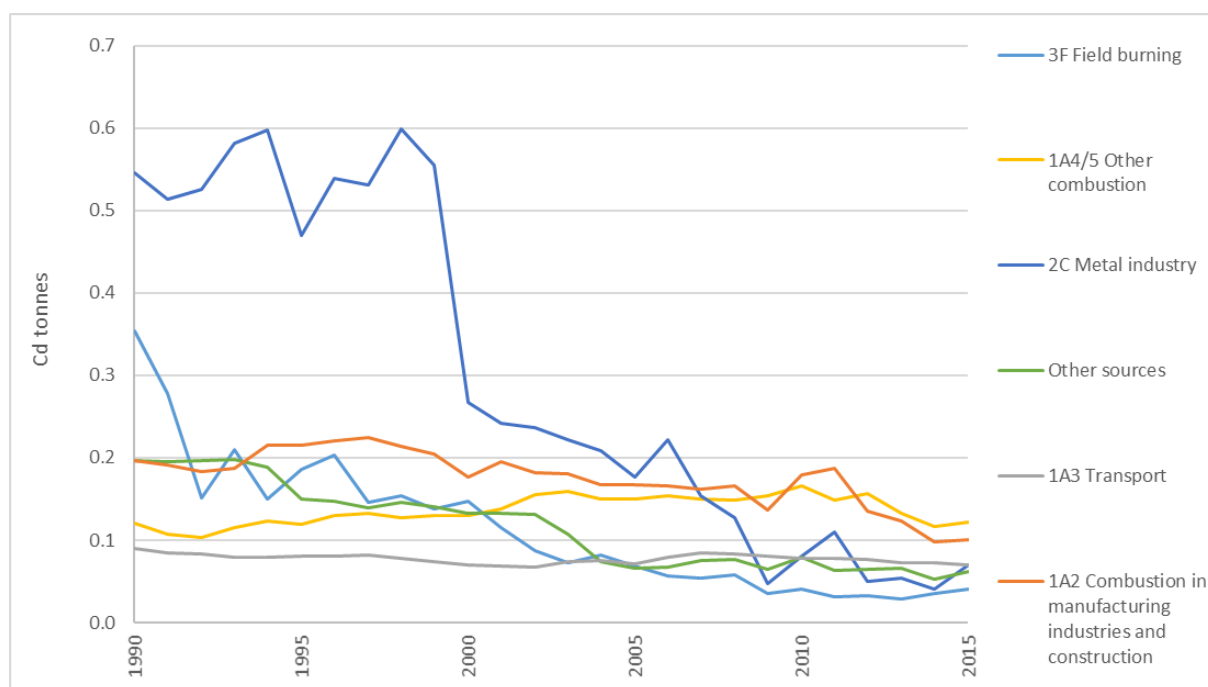


Figure 2.32. Trends in cadmium emissions, 1990-2015. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

Stationary combustion in private households was the most important source of cadmium emissions in 2015. It contributed to 21 per cent of the total emissions. Wood burning is the largest source within the private households sector.

Combustion in manufacturing industries and construction and transport are also large sources of cadmium emissions. In 2015, they contributed to 22 and 15 per cent of the total emissions, respectively.

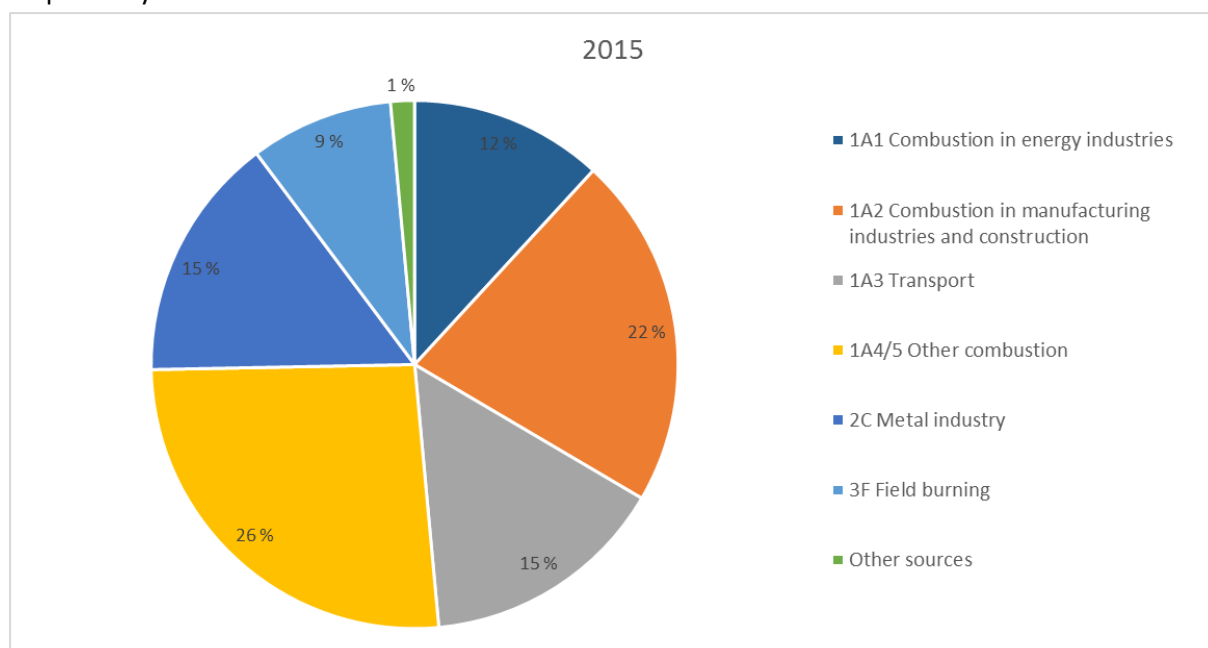


Figure 2.33. Distribution of cadmium emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.10 Mercury

Emissions of mercury amounted to 0.24 tonnes in 2015, which is a 83 per cent reduction from 1990 and a 10 per cent reduction from 2014. The decrease is mainly due to reductions within metal industry sector and use of tobacco. These sectors contributed, respectively, to 43 and 20 per cent of total mercury emissions in 1990, and have been reduced by 93 and 96 per cent, respectively, since then.

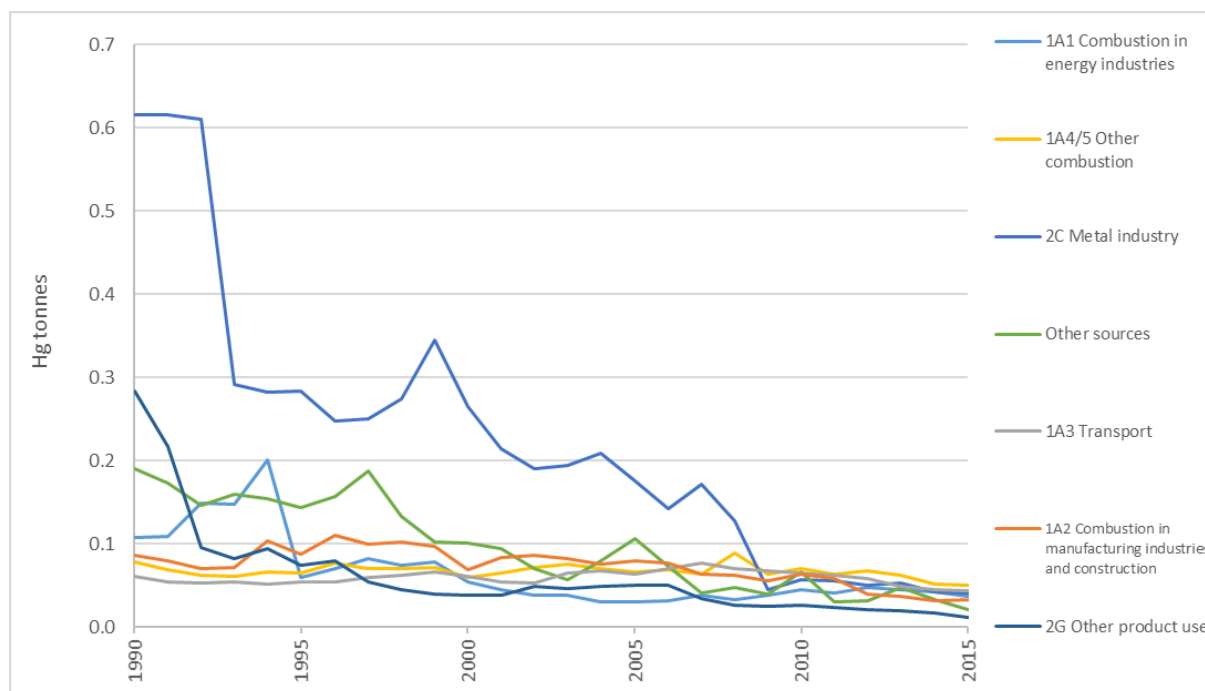


Figure 2.34. Trends in mercury emissions, 1990-2015. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

Mercury emissions originate from a wide range of sources. The most important sources of mercury emissions in 2015 were the category “other combustion” (NFR 1A4 and 1A5). The emissions from this source have decreased by 36 per cent since 1990.

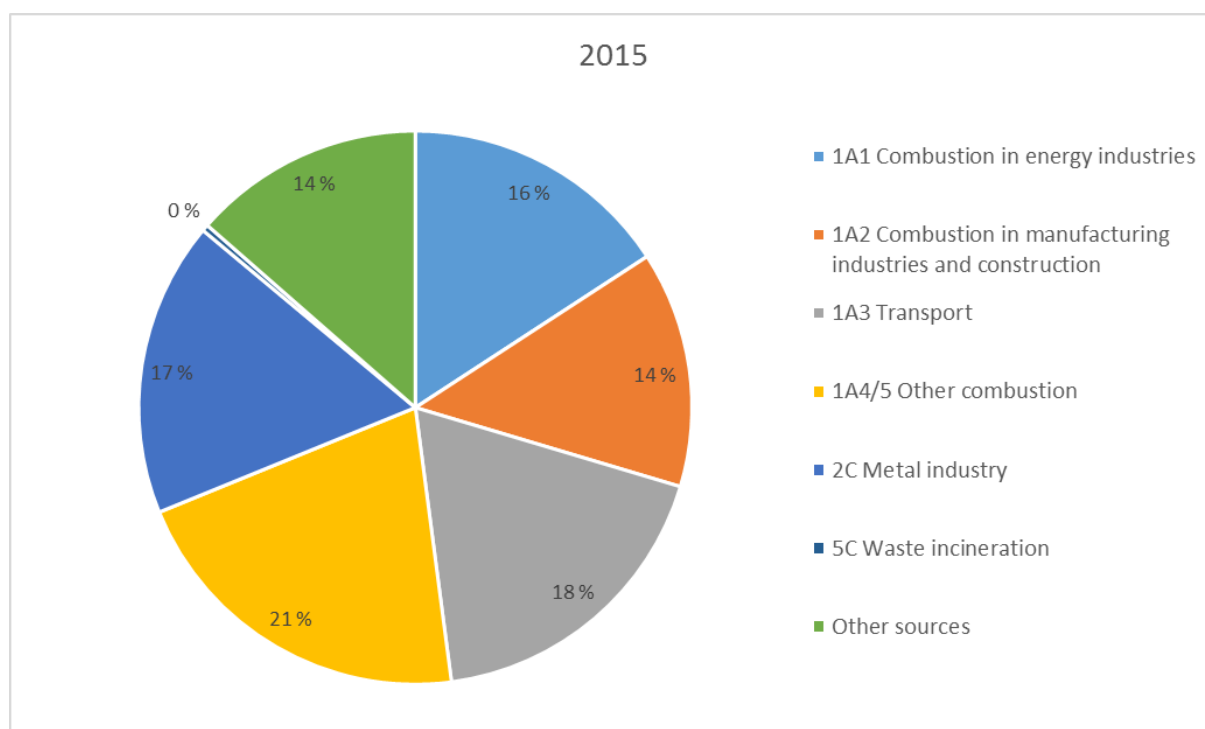


Figure 2.35. Distribution of mercury emissions between emission sources, 2015. Per cent

Source: Statistics Norway/Norwegian Environment Agency

2.3.11 Chromium, arsenic and copper

Emissions of chromium amounted to 3.1 tonnes in 2015 and have been relatively constant in recent years. Emissions have been reduced by 73 per cent since 1990.

Combustion in the chemical industry is the most dominant source of chromium emissions in 2015.

In 2015, 1.6 tonnes of arsenic were emitted, which was an increase of 26 per cent from 2014 and a reduction of 55 per cent from 1990. For the past few years, the variation in arsenic emissions has been due to varying arsenic content in raw materials and reducing agents used in metal production.

Emissions of copper were 28.1 tonnes in 2015, a reduction of 1 per cent since 2014 and an increase of 17 per cent since 1990. Brake wear is the dominant source of emissions of copper.

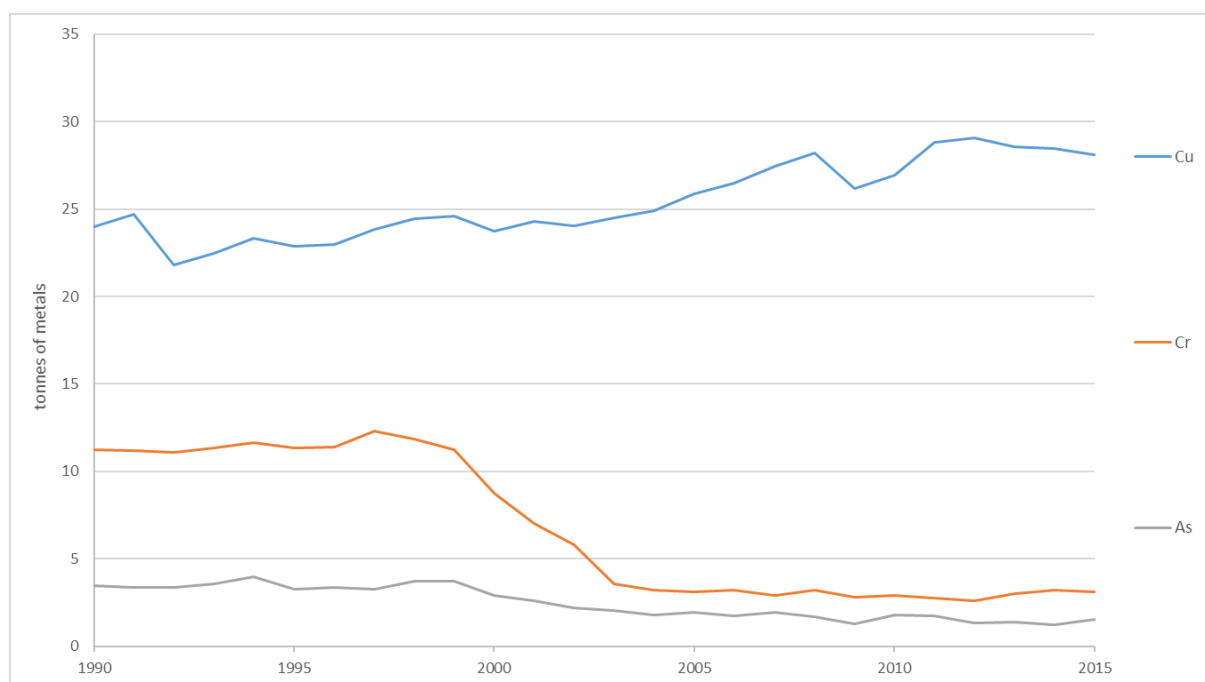


Figure 2.36. Trends in copper, chromium and arsenic emissions, 1990-2015. Tonnes

Source: Statistics Norway/Norwegian Environment Agency

3 ENERGY (NFR sector 1)

3.1 Overview

This chapter provides descriptions of methodologies employed to calculate emissions from the energy sector. The disposition of the chapter is following the NFR classifications of the emission sources. In section 3.2, emission estimations from energy combustion are described. This includes combustion emissions from energy industries, manufacturing industries and construction, transport and other combustion sources. Section 3.2 also includes memo items about international bunker fuels.

In section 3.3, a description is given for fugitive emissions from fuels. This includes fugitive emissions from coal mining and handling, and from oil and natural gas.

3.2 Energy combustion

NFR 1A

Last update: 10.03.2015

3.2.1 Overview

Combustion of fossil fuels and biomass leads to emissions of SO₂, NO_x, NMVOC, CO, particulate matter, heavy metals, PAH, dioxins and NH₃.

Table 3.1. Energy combustion emissions as per cent of total emissions, 2015

Pollutant	Per cent of emissions
SO ₂	35
NO _x	86
NMVOC	22
CO	58
NH ₃	3
PM ₁₀	68
BC	87

Source: Statistics Norway/Norwegian Environment Agency

The emissions of SO₂, NMVOC and CO have been significantly reduced since 1990. The reduction of SO₂ emissions has taken place in all sectors due to reduced sulphur content in fuels. NMVOC and CO emissions have been reduced mainly due to reductions in emissions from petrol passenger cars. Catalysts in petrol passenger cars cause NH₃ emissions.

Emissions of NO_x and particles have stayed stable in the 1990s. Particles have been reduced since 2002 while most of NO_x reductions in totals happened after 2007.

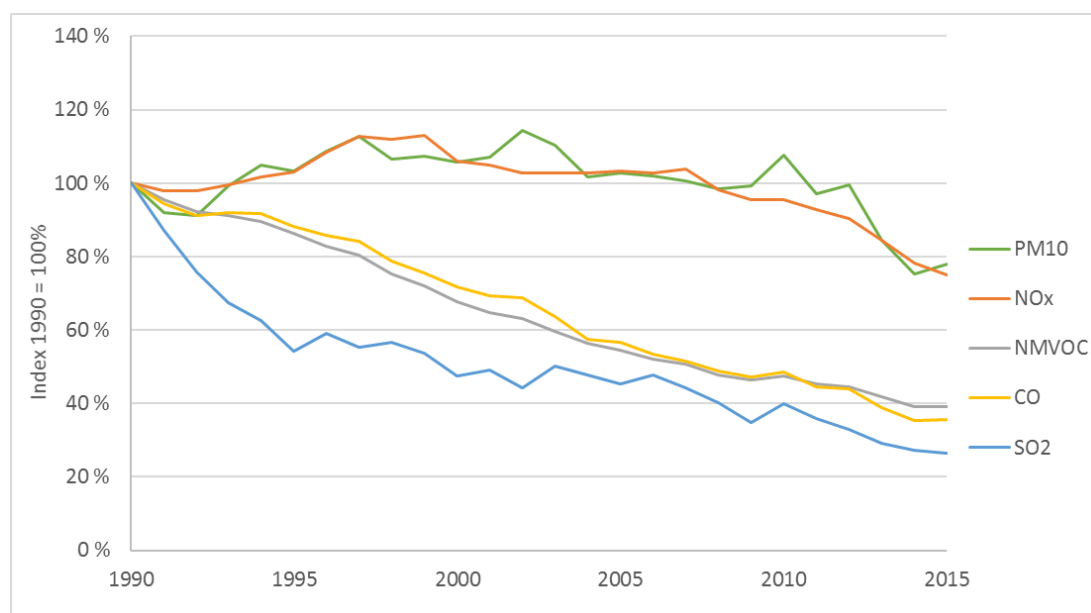


Figure 3.1. Trends for the emissions for most of the long-range transboundary air pollutants from energy combustion. Index 1990 = 100%

Source: Statistics Norway/Norwegian Environment Agency

Emissions from energy combustion include contributions from all sources addressed in the UNECE Guideline. Emissions from waste incineration at district heating plants are accounted for under the energy sector, as the energy is utilised. Emissions from flaring in the energy sectors are described in section 3.3 *Energy production*. Coal and coke used as reducing agents and gas used for production of ammonia (non-energy part) are accounted for under industrial processes. Flaring of natural gas and fuel gas in chemical industry is recorded in section 4.3. Other flaring outside the energy sectors is described in section 6.4. The same applies to emissions from accidental fires etc. Emissions from burning of crop residues and agricultural waste are accounted for and described in section 5.6. Emissions from tobacco are described in section 4.5.6.

The main source for calculation of emissions from energy combustion is the energy balance, which annually is prepared by Statistics Norway. The data used in the emission calculations are at a much more disaggregated level than the published energy balance. Many different sources are utilised in the preparation of the energy balance. E.g. energy use in extraction of oil and gas, which constitutes an important part of Norwegian energy use, is reported from the Norwegian Petroleum Directorate. Other energy producers, such as oil refineries and district heating plants, also report their own energy use to Statistics Norway.

For different oil products, the total frame for annual use is given by Statistics Norway's statistics on deliveries of petroleum products. These statistics are also used in the estimation of use in different economic sectors, together with other available information. The distribution between sectors is of varying quality – e.g. in some cases projections of previous surveys are used in this process. For manufacturing industries, however, Statistics Norway's annual survey on all types of energy use, based on reports from plants that are responsible for approximately 96 per cent of the energy use in these sectors, combined with estimations for the remaining plants, provides

figures of high quality.

3.2.1.1 Method

3.2.1.1.1 General

Emissions from energy combustion are estimated at the sectoral level in accordance with the IPCC sectoral approach Tier 2/Tier 3. Often total fuel consumption is better known than the sectoral consumption.

The general method to estimate emissions from fuel combustion is multiplication of fuel consumption by source and sector by an appropriate emission factor. Exceptions are road and air transport where more detailed estimation models are used, involving additional activity data (see section 3.2.4.2 and 3.2.4.1 respectively). Fuel consumption figures are taken from the Norwegian energy balance. The mean theoretical energy content of fuels and their density are listed in Table 3.2.

Table 3.2. Average energy content and density of fuels

Energy commodity	Theoretical energy content ¹	Density
Coal	28.1 GJ/tonne	..
Coal coke	28.5 GJ/tonne	..
Petrol coke	35.0 GJ/tonne	..
Crude oil	42.3 GJ/tonne = 36.0 GJ/m ³	0.85 tonne/m ³
Refinery gas	48.6 GJ/tonne	..
Natural gas (dry gas) ²	35.5 GJ/1000 Sm ³	0.74 kg/Sm ³ (domestic use)
Natural gas (rich gas) ²	40.3 GJ/1000 Sm ³	0.85 kg/Sm ³ (continental shelf)
Liquefied propane and butane (LPG)	46.1 GJ/tonne = 24.4 GJ/m ³	0.53 tonne/m ³
Fuel gas ³	50.0 GJ/tonne	..
Petrol	43.9 GJ/tonne = 32.5 GJ/m ³	0.74 tonne/m ³
Kerosene	43.1 GJ/tonne = 34.9 GJ/m ³	0.81 tonne/m ³
Diesel oil, gas oil and light fuel oil	43.1 GJ/tonne = 36.2 GJ/m ³	0.84 tonne/m ³
Heavy distillate	43.1 GJ/tonne = 37.9 GJ/m ³	0.88 tonne/m ³
Heavy fuel oil	40.6 GJ/tonne = 39.8 GJ/m ³	0.98 tonne/m ³
Methane ⁴	50.2 GJ/tonne	..
Wood	16.8 GJ/tonne = 8.4 GJ/solid m ³	0.5 tonne/solid m ³
Wood waste (dry wt)	16.25-18 GJ/tonne	..
Black liquor (dry wt)	7.2-9.2 GJ/tonne	..
Waste	10.5 GJ/tonne	..

¹ The theoretical energy content of a particular energy commodity may vary; Figures indicate mean values. All data are net calorific value (NCV).

² Sm³ = standard cubic metre (at 15 °C and 1 atmospheric pressure).

³ In this inventory, *fuel gas* is a hydrogen-rich excess gas from petrochemical industry.

⁴ Landfill gas and other types of biogas are reported as methane content in the energy balance.

Source: Statistics Norway/Norwegian Environment Agency

Handbook of Emission Factors (HBEFA; (INFRAS 2010)) describes methodologies used for road traffic. Several documentation reports have been published describing the methodologies used for road traffic (Holmengen & Fedoryshyn 2015) and navigation (Tornsjø 2001) and (Flugsrud et

al. 2010). The methodology for aviation is described in an internal document from Statistics Norway (Skullerud 2014).

3.2.1.1.2 *Delimitation towards industrial processes etc.*

The energy combustion sector borders to several other source categories. This section presents the demarcation with other sectors used in the inventory.

Energy consumption reported as activity data in the emission inventories is generally delimited in the same way as emissions. In cases where different substances are handled differently, the delimitation of energy consumption follows the delimitation of CO₂ emissions.

Flaring is not reported as energy use in 1A. Instead, flaring is reported in the following source categories:

- Flaring in refineries and in exploration/extraction is reported in 1B – Fugitive emissions.
- Flaring in manufacturing industries is reported in 2 – Industrial processes, particularly in 2B – Chemical industry. (In the energy balance, flaring in manufacturing is reported as "losses".)
- Flaring of landfill gas is reported in 6C – Waste incineration.

Combustion of *solid waste* and *hazardous waste* is reported in the energy section (district heating in 1A1a and in several manufacturing industries). No significant combustion of solid or hazardous waste occurs without energy recovery.

Combustion of *landfill gas* with energy recovery is reported in the energy section (mainly in 1A4a Commercial/Institutional). Flaring is reported in 6C waste incineration, as mentioned above.

Some special problems relating to allocation of reported total plant emissions are discussed in section 3.2.1.1.4.

3.2.1.1.3 *Emissions reported by plants: overview*

However, for some major manufacturing plants (in particular offshore activities, refineries, gas terminals, cement industry, production of plastics, ammonia production), emissions of one or more compounds, reported to the Norwegian Environment Agency from the plants, are used instead of figures calculated with general emission factors as described above. In these cases, the energy consumption at the plants in question is subtracted from the total energy use before the general method is used to calculate the remaining emissions of the compound in question, in order to prevent double counting.

Emissions are reported to the Norwegian Environment Agency under a number of different reporting obligations. Most CO₂ emissions are today reported as part of the Emissions Trading System (ETS).

In the general equation (2.2), *Emissions* (E) = $[(A - A_{PS}) \cdot EF] + E_{PS}$, E_{PS} represents the reported

emission data, while A_{PS} represents the energy consumption at the plants. Note that for most plants, reported emissions are used only for a limited number of substances. For the remaining substances in the inventory, the general method with standard emission factors is used. Reported figures are used for a relatively small number of plants, but as these contribute to a large share of the total energy use, a major part of the total emissions are based on such reported figures. For the source categories petroleum refining, manufacture of solid fuels and other energy industries and iron and steel, more than 90 per cent of the sector emissions are based on reported data from plants. The reports are from the mandatory reporting obligation that is a part of the plants' permits given by the authorities.

3.2.1.1.4 Emissions reported by plants: Energy data

Energy data for plants with reported emissions (A_{PS} in equation (2.2)) should be consistent both with the energy balance that is used for activity totals A and with the reported emission data. Consistency with emission data means that the energy data should correspond to the same activity as the reported emissions.

In most cases, figures on plant energy use in the inventory are based on data reported from the plants to Statistics Norway. This ensures consistency with the energy balance. However, for some plants, some of the energy data may differ between reports to Statistics Norway and data reported together with emissions to the Norwegian Environment Agency. This may lead to problems with consistency. This is most evident for CO_2 , where the range of implied emission factors is very narrowly delimited. However, it may also affect the LRTAP substances.

3.2.1.1.5 Emissions reported by plants: Allocation to combustion/ processes

In some cases, emissions are reported as a plant total which includes both combustion and process emissions. It is likely that most of the emissions are from processes. All emissions of particulates, heavy metals and POPs are entered into the inventory as process emissions. Emissions from combustion are set to 0 in order to avoid double counting.

3.2.1.1.6 Emissions reported by plants: Allocation to fuels

The following discussion is particularly relevant for CO_2 , but it also applies to the LRTAP substances.

For some plants and substances, emissions are reported by fuel, but in most cases reported combustion emissions are entered as a plant total. The emissions are then allocated to fuels bases on standard EFs using equation 3.1:

$$(3.1) \quad E_{PS,f} = E_{PS} \cdot A_{PS,f} \cdot EF_f / \sum_f (E_{PS} \cdot EF_f)$$

where the subscript f denotes fuel type.

This means that any deviations in data will be distributed across all fuels at the plant. Typical situations include:

- Plants with atypical fuels which differ from standard emission factors.
- Plants with errors or other inconsistencies in energy data.

In such cases, implied emission factors may deviate from the standard range also for other fuels than the one which is really affected.

Plants/substances which are entered by fuel currently include among others:

- Particulate matter from manufacturing of wood products.
- Heavy metal and POP emissions from combustion of municipal solid waste and special waste.

3.2.1.2 Activity data

The annual energy balance, compiled by Statistics Norway, forms the framework for the calculation of emissions from energy use. The energy balance defines the total energy consumption for which emissions are accounted. However, a large part of the total emissions are based on reports from plants that use much energy, i.e. offshore activities and energy-intensive industries on shore. Energy consumption in these plants is included in the energy balance, but this consumption is subtracted before the remaining emissions are calculated by the standard method of multiplying energy use by emission factors. Energy figures reported from the plants to Statistics Norway, which are used in the energy balance, sometimes deviate from the energy figures used to estimate reported emission figures, and this may cause inaccuracies in implied emission factors.

The energy balance surveys the flow of the different energy carriers within Norwegian territory. It includes energy carriers used as raw materials and reducing agents, but these are presented in a separate item and are not included in the data used to estimate emissions from combustion. Some emissions vary with the combustion technology; a distribution between different sources is thus required. Total use of the different oil products is based on the Norwegian sales statistics for petroleum products. For other energy carriers, the total use of each energy carrier is determined by summing up reported/estimated consumption in the different sectors. A short summary of the determination of amounts used of the main groups of energy carriers and the distribution between emission sources is given below.

Natural gas

Most of the combustion of natural gas is related to extraction of oil and gas on the Norwegian continental shelf. The amounts of gas combusted, distributed between gas turbines and flaring, are reported annually to Statistics Norway by the Norwegian Petroleum Directorate (NPD). These figures include natural gas combusted in gas turbines on the various oil and gas fields as well as on Norway's four gas terminals on shore. Statistics Norway's annual survey on energy use in manufacturing industries and sales figures from distributors give the remainder. Some manufacturing industries use natural gas in direct-fired furnaces; the rest is burned in boilers and, in some cases, flared.

LPG and other gases

Consumption of *LPG* in manufacturing industries is reported by the plants to Statistics Norway in the annual survey on energy use. Figures on use of *LPG* in households are based on sales figures, collected annually from the oil companies. Use in agriculture and construction is based on non-annual surveys; the figure for agriculture is interpolated for years not included in surveys, whereas the figure for construction is adjusted annually, based on employment figures.

Use of *refinery gas* is reported to Statistics Norway from the refineries. The distribution between the sources direct-fired furnaces, flaring and boilers is based on information collected from the refineries in the early 1990's. However, the total emissions from the refineries included in the inventory are equal to emissions reported from the plants and are regarded being of high quality.

At some industrial plants, excess gas from chemical and metallurgical industrial processes is burned, partly in direct-fired furnaces and partly in boilers. These amounts of gases are reported to Statistics Norway. A petrochemical plant generates *fuel gas* derived from ethane and *LPG*. Most of the gas is burned on-site, but *fuel gas* is also sold to several other plants. All use of *fuel gas* is reported as energy consumption in the inventory.

Several metallurgical plants generate *CO-rich excess gas* that is either burnt on-site or sold to adjacent plants. Two ferroalloy plants sell parts of their *CO* gas to some other plants (one producer of ammonia, a district heating plant, iron and steel producers and mineral industry), where it is combusted for energy purposes. These amounts are reported as energy consumption.

One sewage treatment plant utilizes biogas extracted at the plant, and reports quantities combusted (in turbines). By definition, no *CO₂* emissions arise from biogas, but other emissions are estimated by Statistics Norway, using the same emission factors as for combustion of natural gas in turbines.

Oil products

Total use of the different oil products is based on Statistics Norway's annual sales statistics for petroleum products. The data are generally considered reliable (with some reservations which are accounted for further down in this chapter), since all major oil companies selling oil products report to these statistics and have an interest in the quality of the data. The statistics are corrected for direct import by other importers or companies. The use of sales statistics provides a total for the use of oil products. The use in the different sectors must sum up to this total. This is not the case for the other energy carriers. The method used for oil products defines use as identical to sales; in practice, there will be annual changes in consumer stocks, which are not accounted for.

However, since the late 1990s the distribution in the sales statistics between different middle distillates has not been in accordance with the bottom-up estimated consumption of the

products. In particular, the registered sales of light fuel oil have generally been too low, and it is known that some auto diesel also is used for heating. In order to balance the accounts for the different products, it has since 1998 been necessary to transfer some amounts between products instead of using the sales figures directly. The most important transfer is from auto diesel to light fuel oil, but in addition some auto diesel has also been transferred to heavy distillate.

Due to inaccuracies in the reporting of sales of marine gas oil from approximately 2005, there is also some uncertainty connected to the distribution between domestic and international sea transport for the latest years.

Stationary use takes place in boilers and, in some manufacturing industries, in direct-fired furnaces. There is also some combustion in small ovens, mainly in private households. Mobile combustion is distributed between a number of different sources, described in more detail in chapter 3.2.4 Transport. In addition to oil products included in the sales statistics, figures on use of waste oil are given in Statistics Norway's statistics on energy use in the manufacturing industries. Statistics Norway also collects additional information directly from a few companies about the use of waste oil as a fuel source.

Generally, in Norway there is a continual shift between use of oil and hydroelectricity, corresponding to changes in prices. Between years, this may cause changes in use of oil products and corresponding emissions which can be considerable.

Coal, coke and petrol coke

Use of coal, coke and petrol coke in manufacturing industries is annually reported from the plants to Statistics Norway. The statistics cover all main consumers and are of high quality. Combustion takes place partly in direct-fired furnaces, partly in boilers. Figures on some minor quantities burned in small ovens in private households are based on sales figures. In addition, the figure on an insignificant use of coal in the agricultural sector was formerly collected from one farmer. Since 2002, there has been no use of coal in Norwegian agriculture.

Biofuels

Use of wood waste and black liquor in manufacturing industries is taken from Statistics Norway's annual survey on energy use in these sectors. Use of wood in households is based on figures on the amount of wood burned from the annual survey on consumer expenditure for the years before 2005 and for 2012. The statistics cover purchase in physical units and estimates for self-harvest. The survey figures refer to quantities *acquired*, which not necessarily correspond to *use*. The survey gathers monthly data that cover the preceding twelve months; the figure used in the emission calculations (taken from the energy balance), is the average of the survey figures from the year in question and the following year. For the years 2005-2011, the figures are based on responses to questions relating to wood burning in Statistics Norway's Travel and Holiday Survey. The figures in this survey refer to quantities of wood *used*. The survey quarterly gathers data that cover the preceding twelve months. The figure used in the emission calculations is the average of five quarterly surveys. Figures on some minor use in agriculture and in construction

are derived from earlier surveys for these sectors. Combustion takes place in boilers and in small ovens in private households. Consumption figures for wood pellets and wood briquettes are estimates, based on annual information from producers and distributors. Data on use of peat for energy purposes is not available, but according to the Energy Farm, the centre for Bioenergy in Norway, such use is very limited (Hohle 2005).

Waste

District heating plants and incineration plants annually report combusted amounts of waste (boilers) to Statistics Norway and the Norwegian Environment Agency. There is also some combustion in manufacturing industries, reported to Statistics Norway.

According to the Norwegian Pollution Act, each incineration plant has to report emission data for SO₂, NO_x, CO, NH₃, particles, heavy metals and dioxins, and the amount of waste incinerated to the county governor. The county governor then reports this information to the Norwegian Environment Agency. If emissions are not reported, the general method to estimate emissions from waste incineration is to multiply the amount of waste used by an appropriate emission factor. Normally a plant specific emission factor is made for the component in question. This factor is based on the ratio between previous emission figures and quantities of waste burned. This factor is then multiplied with the amount of waste incinerated that specific year.

Energy balance sheets vs. energy accounts

There are two different ways of presenting energy balances: Energy balance sheets (EBS) and energy accounts. The energy figures used in the emission calculations are mainly based on the energy balance sheets.

The energy accounts follow the energy consumption in Norwegian economic activity in the same way as the national accounts. All energy used by Norwegian enterprises and households is to be included. Energy used by Norwegian transport trades and tourists abroad is also included, while the energy used by foreign transport industries and tourists in Norway is excluded.

The energy balance sheet follows the flow of energy within Norway. This means that the figures only include energy sold in Norway, regardless of the users' nationality. This leads to deviations between the energy balance sheet and the energy accounts, especially for international shipping and aviation.

The energy balance sheet has a separate item for energy sources consumed for transportation purposes. The energy accounts place the consumption of all energy under the relevant consumer sector, regardless of whether the consumption refers to transportation, heating or processing.

Figures from the energy sources balance sheet are reported to international organisations such as the OECD and the UN. The energy balance sheet should therefore usually be comparable with international energy statistics.

Important differences between figures presented in the energy balance sheet (EBS) and figures used in the emission calculations (EC) are:

- *Air transport*: EC use only Norwegian domestic air traffic (excluding military), while EBS includes all energy sold in Norway for air transport, including military and energy used for international air transport.
- *Coal/coke for non-energy purposes*: This consumption is included in net domestic consumption in EBS, whereas EC include only energy used for combustion in the calculation of emissions from energy. Emissions from coal and coke used as reducing agents are accounted for in the IPPU sector.

3.2.1.3 Emission factors

Emission factors used for the energy sector are given in Appendix B. Emission factors for SO₂ are independent of combustion technology. In cases where technology for cleansing of SO₂ has been installed, this will be reflected in the emission figures reported from the respective plants. For the other emission components, further descriptions are also given for each source sector.

The emission factors of NO_x, CO, NMVOC and NH₃ for stationary combustion have been evaluated by Norsk Energi for the Norwegian Environment Agency. The evaluation is described in the report "Vurdering av utslippsfaktorer for beregning av NO_x-utslipp med mer fra stasjonær forbrenning i Norge" (Evaluation of NO_x emission factors etcetera from stationary combustion in Norway) (Norsk Energi 2003). The report focused mainly on NO_x, but also emission factors for CO, NMVOC and NH₃ were considered.

The conclusion in Norsk Energi (2003) was that there are significant discrepancies between the emission factors from literature and the factors used in the inventory. Some of the emission factors used in the national inventory are higher and some lower than the emission factors found in literature. To some extent the discrepancy is due to the fact that the emission factors from literature are not reflecting technology used in Norway and are therefore not valid for Norwegian conditions. In addition it is considered that some of the Norwegian emission factors are based on more reliable data than the factors from literature. However, Norsk Energi (2003) proposed to change some of the emission factors, due to the fact that the factors from literature were considered to be of better quality than those used in the Norwegian emission inventory. One of the factors was the NO_x emission factor for heavy fuel oil, see below. In general, for all other compounds the emission factors proposed in Norsk Energi (2003) were lower than the emission factors that are used in the Norwegian emission inventory. We consider that the effect on national totals of not replacing the emission factors with the proposed factors in Norsk Energi (2003) has led to overestimated emissions. However, Norway is continuously considering all aspects of the Norwegian emission inventory, including the emission factors, and with the updated EMEP Guidebook (EEA 2016) we now consider to evaluate the emission factors in our inventory.

NO_x

The NO_x emission factors used in the Norwegian emission inventory have, as mentioned above, been evaluated by Norsk Energi in Norsk Energi (2003) and also in "NO_x-utslipp i forbindelse med eventuell NO_x-avgift" (Evaluation of NO_x emissions in connection with implementing NO_x tax) (Norsk Energi 2006). The conclusion in both reports is that the NO_x emission factors used in

the inventory are within the intervals Norsk Energi found in their own measured data and from literature.

Norsk Energi (2003) concluded that the general emission factor for heavy fuel oil should be considered to be changed from 4.2 to 5 kg NO_x per tonne fuel and for chemical and metal industry from 5 to 6 kg NO_x per tonne heavy fuel oil. The consumption of heavy fuel oil in stationary combustion in Norway is very small and NO_x emissions in the Norwegian inventory from the largest consumers of heavy fuel oil in industry are based on plant specific data. Due to this, the proposed emission factors from Norsk Energi (2003; 2006) are not included in the Norwegian emission inventory.

SO₂

The emission factors for SO₂ from oil products change yearly, in accordance with variations in the sulphur content in the products. The presented factors refer to uncleansed emissions; in cases where the emissions are reduced through installed cleansing measures, this will be reflected in emission figures reported from the respective plants.

3.2.1.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C, as well as under the individual underlying source categories.

Generally, the total energy use is less uncertain than the energy use in each sector. For some sectors (e.g. the energy and manufacturing industries) the energy use is well known, while it is more uncertain in households and the service sectors. The energy use in the most uncertain sectors has been adjusted in the official energy statistics, so that the sum of the energy use in all sectors equals the total sales.

3.2.1.5 Completeness

All known combustion with energy utilization in different industries and private households is included.

3.2.1.6 QA/QC

The emission sources in the energy sector are subjected to the QA/QC procedures described in section 1.6.

3.2.2 Energy industries

NFR 1A1

Last update: 10.03.2015

3.2.2.1 Description

Energy industries include emissions from electricity and heat generation and distribution, extraction of oil and natural gas, coal production, gas terminals and oil refineries. Norway produces electricity mainly from hydropower, so emissions from electricity production are small compared to most other countries. Due to the large production of oil and gas, the emissions from combustion in energy production are high.

Emissions from drilling at moveable offshore installations are included here. Emissions from these installations while not in operation (during transport, etc.) are included with 1A3d Navigation.

3.2.2.2 Method

A general description of the method used for estimation of emissions from fuel combustion is given in section 3.2.1.1. For waste incineration, also a more detailed description of the methodology for some components is given in this section.

Waste incineration

N₂O and NO_x

Emissions of NO_x are reported from each plant to the Norwegian Environment Agency. An estimated amount of 2.5 per cent of this NO_x is subtracted and reported to UNFCCC as N₂O (SFT 1996). Accordingly, the net NO_x emissions constitute 97.5 per cent of the emissions reported by the plants. For some years, emissions of NO_x have not been reported for a number of plants. In these cases, specific emission factors for the plants have been made, based upon earlier emissions and amounts of waste incinerated. These new factors have been used to estimate the missing figures.

Particles

Emissions of particles from district heating plants are reported to the Norwegian Environment Agency. The different plants started to report particulate emissions at various points in time. Most of them started reporting from 1994. Emissions of particles in the years before reporting have been assumed to be the same as in the first year the plant reported. New control device systems (mainly wet scrubbers) were installed at the end of the 1980s at the largest plants. Around 1995 more control device systems were installed as a result of stricter emission requirements. Most plants today have fabric filter or electrofilter together with wet scrubbers. Only two plants do not have wet scrubbers.

The emission permits do not state which particle fraction that is going to be measured. It is common to measure total amount of particles. It is however presumed that the particles emitted are less than PM_{2.5}. TSP and PM₁₀ are therefore the same as PM_{2.5}.

Dioxins

Emissions of dioxins from waste burning at district heating plants are reported to the Norwegian Environment Agency. We have reported data for each plant from the period 1994/1995. Before 1994 we have only national totals. For estimating the emissions of dioxins for each plant before 1994 we derived an emission factor from total amount of waste burned together with the total dioxin estimate. The emissions of dioxins were estimated by multiplying the given emission factor of 20 µg/tonne waste by the amount of waste burned at each plant. This calculation was done for each of the missing years for plants that did not report emissions.

PCB

PCB emissions are not reported to the Norwegian Environment Agency. A country specific

emissions factor has been used to estimate PCB. To take into accounts emission reduction systems implemented in incinerators, this emission factor decreases during the period, following the trend of dioxins emission factor.

Heavy metals

The estimate of heavy metals from waste combustion at district heating plants is reported to the Norwegian Environment Agency. Before 1999, many emissions of heavy metals were reported together as one group. This made it difficult to use the data to estimate the emission of each component. From 1999, there are separate data for each component, but for As, Cr and Cu there are a few plants that have insufficient reporting. To calculate the emissions of heavy metals before 1999 we have estimated an emission factor for each plant with the aid of reported emission data and amount of waste burned at each plant. The emission factor derived has been used to calculate emissions for previous years by multiplying each specific emission factor with the amount burned for the corresponding year for each plant.

Every district heating plant had stricter emission requirements for particles from 1995. It is expected that the emissions of heavy metals, except for mercury, were reduced analogously. At the same time, the emission of mercury was regulated from 0.1 mg/Nm³ to 0.05 mg/Nm³. These regulations are considered while calculating emissions for previous years.

3.2.2.3 Activity data

Electricity and heat generation and distribution

The energy producers annually report their use of different energy carriers to Statistics Norway. There is only some minor use of oil products at plants producing electricity from hydropower. Combustion of coal at Norway's only dual-purpose power plant at Svalbard/Spitsbergen is of a somewhat larger size. The amount of waste combusted at district heating plants is reported annually both to Statistics Norway and the Norwegian Environment Agency. The data are considered to be of high quality.

Extraction of oil and natural gas

Production of oil and natural gas is the dominating sector for emissions from combustion in the energy industries in Norway. The Norwegian Petroleum Directorate annually reports the amounts of gas combusted in turbines and diesel burned in turbines and direct-fired furnaces on the oil and gas fields. The data are of high quality, due to the CO₂ tax on fuel combustion. These activity data are used for 1990-2002. From 2003 onwards, reported emission figures from the field operators are used.

Coal production

Norway's coal production takes place on Svalbard. The only coal producing company annually reports its coal consumption and some minor use of oil products. In addition to emissions related to Norway's own coal production, also emissions from Russian activities are included in the Norwegian emission inventory. Russian activity data are scarce, and emissions from an estimated quantity of coal combusted in Russian power plants are calculated. Since 1999 there has been only one such plant, in earlier years there were two of those.

Gas terminals

Natural gas from the Norwegian continental shelf is landed, treated and distributed at gas terminals on shore. There are four gas terminals in Norway. The eldest started up before 1990, one in 1996 and two in 2007. Annual figures on natural gas combusted in turbines and flared are reported to the Norwegian Petroleum Directorate (figures on flaring at one plant is reported to the Norwegian Environment Agency).

Gas power plants

Norway had in 2011 two major gas power plants and several minor ones. The first of the large plants was opened in 2007 and runs intermittently, depending on electricity and gas prices. The second large plant was opened in 2010. Several of the smaller plants are back-up plants that are run only in emergency situations. Thus, there will be large annual fluctuations in emissions.

Oil refineries

The oil refineries annually report their use of different energy carriers to Statistics Norway. Refinery gas is most important, but there is also some use of LPG and oil products. Burning of coke while regenerating the catalyst in cracker units is reported under 1B2A4 – Fugitive emissions – Refining/Storage.

3.2.2.4 Emission factors

Emission factors used for the energy sector are given in Appendix B. For some industries and components more information about the derivation of the emission factors are given in this section.

3.2.2.4.1 SO₂

Russian electricity and heat production

Emissions from combustion of coal for electricity production in the Russian settlements on Svalbard are included in the Norwegian emission inventory. Up to 1998 there were two Russian settlements with electricity and heat production: Barentsburg and Pyramiden. Since the coal production at Pyramiden was closed down in 1998, the settlement was abandoned, and all activity now takes place in Barentsburg. For SO₂, emission factors are based on information from Trust Arktikugol in Moscow. From 1999, the factor 70 kg/tonne is used, and for earlier years 16 kg/tonne. The factor-estimated figures are reduced by 60 per cent, due to the assumption that such an amount of the sulphur is bound in the ash.

3.2.2.4.2 NO_x

Offshore installations

NO_x emissions from diesel engines at offshore installations were revised in 2014 based on Karlsson and Finborud (2012). The recommended factors are shown in Table 3.3.

Table 3.3. Recommended emission factors for NO_x for different engine types

	Before 2000 kg NO _x /tonne fuel	After 2000 kg NO _x /tonne fuel	Previous default factor
200-1000 rpm: Medium speed	54	53	70
1000-1500 rpm: High speed, lower range	50	50	60
> 1500 rpm: High speed, higher range	45	44	55

Source: Karlsson and Finborud (2012)

From 2003, emissions at fixed installations and at moveable installations during drilling operations are taken from reports from operators. Some operators use default emissions factors, whereas an increasing fraction use plant-specific factors.

In the implementation of the factors from Karlsson and Finborud (2012), the following principles were used:

- Reported emissions with implied emission factors less than 1 per cent from the old default values (70/60/55 kg/t) were assumed to having used the default factors. These emissions were reduced to the new default values for engines from before 2000 (54/50/45 kg/t).
- Other reported emissions were assumed to having used plant-specific factors and were left unchanged.
- Emissions from other consumption in engines were calculated with a general factor of 54 kg NO_x/tonne. This applies to all emissions before 2003, and the remaining fraction of sales to the oil and gas industry from 2003 onwards.
- Emissions from use of marine gas oil for turbines have not been part of this revision.

3.2.2.4.3 TSP, PM₁₀ and PM_{2.5}

Electricity and heat generation

Emission factors for TSP, PM₁₀ and PM_{2.5} are based on emission data given in EPA (2002). EPA (2002) gives emission data based on measurements made from various boilers using different control device systems. The Norwegian power plant at Svalbard is equipped with a multicyclone, and emission factors derived from measurements from boilers controlled with multicyclone device systems are used.

Waste incineration

Emissions of particles from district heating plants are reported to the Climate and Pollution Agency.

3.2.2.4.4 BC

For energy industries, BC emissions are estimated as fractions of PM_{2.5}-emissions. The share of BC has been given by IIASA (Kupiainen & Klimont 2007) as it is used in the GAINS model.

Emission factors used for the energy sector are given in Appendix B.

Waste incineration

The share of BC among TSP proposed in Kupiainen and Klimont (2004) has been used to estimate BC emissions. It gives a mass share of 0.9 percent of BC in PM_{2.5}.

For incineration of special waste, the same BC share as for heavy fuel oil combustion in residential, commercial, services and agriculture have been used. The share is given in Appendix B.

3.2.2.4.5 Dioxins and PAH

Electricity and heat generation

Dioxin emissions from coal combustion at the power plants at Svalbard are derived from emission factors found in literature. The emission factor used is the emission factor recommended in Bremmer et al. (1994). The same emission factor is also used in Parma *et al.* (1995) and Hansen (2000). Burning of coal at power plants is also expected to give particle-bound dioxin emissions, but because of the effective control device using multicyclone collector, the emissions are expected to be low.

Emission factors used for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are the emission factor recommended in Guidebook 2016. Chapter 5C1, table 3-1. Is used for the years 1990-1995, and table 3-2 is used for the years after 1995.

PAH emissions from waste incineration are calculated by emission factors and amount of waste burned. We have no plant or country specific emission profile of PAH from waste incineration at district heating plants in Norway.

3.2.2.4.6 PCB

So as to take into account emissions regulation in waste incineration, which was first implemented in 1995 and reinforced in 2006, PCB emissions factor decreases along the period. From 1990 to 1994, the value given in the EMEP/EEA guidebook (2007) has been used to estimate PCB emissions from municipal waste and special waste.

In 2010, a PCB emission factor for municipal waste was estimated from emission measurements of waste incinerators. This emission factor has been considered for the estimations of PCB emissions after 2006. For the period 1995-2005, an emission factor has been derived from the dioxins trend observed during the same period.

As no measurements have been done for incineration of special waste, emission factors for the years after 1995 has been estimated using the same trend as for municipal waste. Table 3.4 presents PCB emissions factors.

Table 3.4. Emission factors for PCB from waste incineration plants, mg/tonn of waste

	1990-1994	1995-2005	2006 ->
Municipal waste	0.82	0.00064	0.000032
Special waste	5	0.0039	0.0002

Source: Statistics Norway

3.2.2.4.7 Heavy metals

Electricity and heat generation

The emission factors for heavy metals used for calculating emissions from coal fired power plants are from EEA (2001). The factors are, however, not specific for coal fired power plants but standard factors recommended for calculating emissions from coal combustion in energy and transformation industries.

3.2.2.4.8 HCB

For energy industries, HCB emissions have been estimated for the use of coal, light fuel oil, wood waste, black liquor, municipal and other waste. Emission factors used for the energy sector are given in Appendix B.

Waste incineration

HCB emissions are not reported by waste incineration plants. So as to take into account emissions regulation in waste incineration, which was first implemented in 1995 and reinforced in 2006, HCB emission factor decreases along the period. Most of installations have anticipated 2006 regulation. Therefore, the emission factor from waste incineration in Denmark (Nielsen et al. 2010) has been considered since the year 2005. For 1995, emission factor has been estimated using the reduction trend of dioxins emission factor observed between 1995 and 2004. This emission factor has been used for the period 1995-2004. For the period 1990-1994, an emission factor from a former guidebook (EEA 2007) has been considered. Emission factors are presented in the Table 3.5.

Table 3.5. Emission factors for HCB from waste incineration plants, mg/tonn of waste

	1990-1994	1995-2004	2005 ->
Municipal waste	2	0.9	0.045
Other waste	10	4.5	0.2

Source: Statistics Norway, Nielsen et al. (2010), EEA (2007)

Extraction of oil and natural gas

HCB emissions from the use of marine gas oil in offshore platform have been estimated using the same emission factor as for the use of marine gas oil in ship. 0.08 mg HCB/tonn of gas oil has been used.

3.2.2.5 Uncertainties

Uncertainty estimates for long-range air pollutants are given in Appendix C. Since the energy use is well known for the energy industries, the uncertainty in the activity data is considered to be minor.

The uncertainty in the activity data is ± 3 per cent of the mean for oil, ± 4 per cent for gas and ± 5 per cent of the mean for coal/coke and waste.

3.2.2.6 Completeness

Major missing emission sources are not likely.

3.2.2.7 Source specific QA/QC

The energy industries are subjected to the general QA/QC procedures described in section 1.6. Some source specific QA/QC activities were conducted in the following industries:

Heat generation in district heating plants

Emissions of heavy metals and POPs from waste incineration have been subject to detailed control. The estimates are based on measurements, but the values are uncertain due to high variability. Reported emission values can vary by orders of magnitude from year to year. Each historical value has been checked in the QA/QC process, and some data have been rejected and replaced by calculated values.

Extraction of oil and natural gas

For emissions of NO_x from turbines offshore, time series over the emissions calculated with field specific emission factors have been compared with the emissions given, using the earlier used average emission factor.

From 2003 onwards, field specific emission figures reported from the companies are used directly in the emission model. These figures are compared with emissions calculated on the basis of field specific activity data and emission factors.

3.2.3 Manufacturing industries and construction

NFR 1A2

Last update: 20.05.09

3.2.3.1 Description

Emissions from the sector of manufacturing industries and construction include industrial emissions that to a large extent originating from the production of raw materials and semi-manufactured goods (e.g. metals, petrochemicals, pulp and paper and mineral products). These emissions are related to fuel combustion only, that is, emissions from use of fuel for heating purposes. Consumption of coal as feedstock and reduction medium is not included in this sector, but is accounted for under the industrial processes sector (chapter 4).

The interannual variation in implied emission factors for heavy metals and dioxins in 1A2 is in some cases considerable. Examples are emissions of Pb, Hg and dioxin in 1A2d. These emission estimates are based on a combination of reported figures from the plants to the Norwegian Environment Agency and emissions based on activity data multiplied with emission factors given in Appendix B. Energy use from the same plants is reported to Statistics Norway. Whenever emissions are reported these figures are used in the inventory.

Emissions of Pb in 1A2d have increased since 2000 due to increased combustion of special waste at the plants. The EF used for burning of special waste is 14 g Pb/tonne while for instance the EF for burning of heavy fuel oil is 1 g Pb/tonne. Emissions have increased, while the total energy consumption (activity data) has been more or less at the same level since 1994, with a minor decrease the latest years. This has led to increased IEFs. For some glass production plants the

reported emissions of Pb is not splitted into emissions from process and combustion. In these cases, the whole emission is placed on the source that is thought to be the most prominent.

For Hg the emissions in 1A2d follow the same trend as the amount of liquid fuels used. The IEFs increases the years the consumption of liquid fuels increases and decreases the years the consumption of liquid fuels decreases.

For dioxins, IEFs vary due to variations in reported figures for one plant. The plant burnt various waste fractions in addition to regular fuel. The plant was closed down during 2001.

3.2.3.2 Activity data

Most of the emission figures are calculated on the basis of activity data and emission factors. For some large plants, varying emission figures are based on reported figures from the plants.

Statistics Norway carries out annual sample surveys on energy use in manufacturing industries, which supply most of the data material for the calculation of combustion emissions in these sectors in cases when reported emission figures not are used. The energy use survey is assumed to cover approximately 96 per cent of the energy use in this sector. For the remaining companies, figures are estimated based on data from the sample, together with data on economic turnover, taking into account use of different energy carriers in the same industries and size groups. A change in methodology from 1998 has had minor consequences for the time series, since the energy use is mainly concentrated to a few major plants within the industry, from which data were collected in both the present and the earlier method. The data on energy use in manufacturing industries are considered to be of high quality. Information on use of waste oil and other hazardous waste is also collected through the energy use statistics.

For the construction industry, the figures on use of the different energy carriers are partly taken from the annual sales statistics for petroleum products and partly projected from earlier surveys; the energy data are considered rather uncertain. In some sectors, autodiesel is mainly used in machinery and off-road vehicles, particularly in mining and construction. This amount of fuel is based on reported consumption of duty-free autodiesel in the manufacturing industries and on reported sales of duty-free autodiesel to construction. The methods for calculating emissions from motorized equipment are discussed in section 3.2.4.7. Emissions from off-road machinery in manufacturing industries and construction are reported in NFR category 1A2g vii.

3.2.3.3 Emission factors

Emission factors used for the energy sector are given in Appendix B.

3.2.3.4 Uncertainties

Uncertainty estimates for long-range air pollutants are given in Appendix C. The energy use is considered well known for the manufacturing industries.

3.2.3.5 Completeness

Major missing emission sources are not likely.

3.2.3.6 Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4 Transport

NFR 1A3

3.2.4.1 Aviation

NFR 1A3a

Last update: 04.03.2016

3.2.4.1.1 Method

The calculation methodology applied is described in Skullerud (2014). According to the IPCC Good Practice Guidance the methodology used is Tier 2 based on the detailed methodology in EEA (2001). This methodology allows estimation of emissions and fuel consumption for different types of aircraft according to the average flying distance and numbers of landings and take-offs (LTO). All movements below 1000 m are included in the "Landing Take Off" (LTO) cycle. Movements over 1000 m are included in the cruise phase. All emissions from international aviation are excluded from national totals, and are reported separately (see section 3.2.6.3).

3.2.4.1.2 Activity data

Statistics Norway annually collects data on use of fuel from the air traffic companies. These data include specifications on domestic use and amounts bought in Norway and abroad. The types of fuel used in aircrafts are both jet fuel (kerosene) and aviation petrol. The latter is used in small aircrafts only. Emissions from the consumption of jet kerosene in domestic air traffic are based directly on these reported figures. Domestic consumption of jet kerosene has been reported to Statistics Norway by the airlines since 1993. The survey is annual, but data from the surveys for 1993 and 1994 have not been used here, as one of the largest airlines in Norway was not included. Domestic consumption prior to 1995 is estimated by extrapolation on the basis of domestic kilometres flown and is more uncertain (Finstad et al. 2002a). Sales figures are used for the minor use of aviation petrol. The time series for liquid fuels used in aviation is given in Table 3.6.

Table 3.6. Liquid fuels in aviation, 1990-2015. TJ

	1 A 3 a ii (i) Domestic aviation (LTO)	1 A 3 a i (i) International aviation (LTO)
1990	3 234.4	1 379.2
1995	4 127.3	1 511.0
2000	5 008.3	2 222.3
2005	4 407.1	2 318.0
2006	4 586.2	2 925.6
2007	4 695.8	3 005.2
2008	5 053.6	2 982.1
2009	5 019.1	3 042.7
2010	5 305.9	3 970.4
2011	5 605.7	3 955.8
2012	5 739.0	4 647.4
2013	5 789.1	5 505.9
2014	5 878.7	5 685.0
2015	5 908.7	6 001.7

Source: Statistics Norway

3.2.4.1.3 Emission factors

Emission factors used are given in Appendix B, table B1 and B3, and tables B5-B7.

The Norwegian Petroleum Industry Association provides emission factors for SO₂ for the combustion of jet fuel and gasoline (Finstad et al. 2002a). The emission factor for SO₂ varies annually depending on the sulphur content of the fuel used.

New aircraft and flight phase specific emission factors for NO_x, CO, VOC and particles are given in EEA (2013). All particles are found to be less than PM_{2.5} (Finstad et al. (2001)). The detailed emission factors are combined with the specific fuel consumption for each aircraft and flight phase (EEA 2013), flight data by aircraft type and route from Avinor and the airports (Air transport statistics (Statistics Norway (Annually-a); background data 2013) and route distances to give weighted emission factors on an aggregated level. Separate factors for LTO and the cruise phase are elaborated. Emission factors for particles are from Brock *et al.* (1999) and Döpelheuer and Lecht (1998), and all particles are found to be less than PM_{2.5} (Finstad et al. 2002a).

BC emissions are estimated as fractions of PM_{2.5} emissions. Emission factors are given in Appendix B.

The weighted emission factors are combined with the activity data (fuel consumption) to estimate emissions from civil aircraft, except helicopters.

The new emission factors for civil aircraft except helicopters have been used in the inventory back to 1990. Aggregated emission factors were calculated with the new detailed factors combined with activity data for 1989, 1995, 2000 and 2012. Factors for the years 1990-1994, 1996-1999 and 2001-2011 were interpolated. Factors after 2012 are kept constant. Emission factors for helicopters and military aircraft were kept unchanged (EEA 2001; Finstad et al.

2002a).

3.2.4.1.4 Uncertainties

Activity data

The uncertainty in the activity data for civil aviation is estimated to be ± 20 per cent of the mean, primarily due to the difficulty in separating domestic emissions from emissions from fuel used in international transport (Rypdal & Zhang 2000). Fuel consumption in 2012 was also estimated bottom-up based on the fuel consumption factors from EEA (2013) and compared to the reported figures according to the method described in Finstad et al. (2002a), see also section 3.2.4.1.6). The estimated and reported data differed by about 25 per cent. However, the reported data are considered most accurate and were used in the calculation. As described above, data before 1995 are more uncertain than for later years. This may also to a certain degree affect the time series consistency.

3.2.4.1.5 Completeness

Major missing emission sources are not likely.

3.2.4.1.6 Source specific QA/QC

In 2002, a methodology improvement was made in the emission calculations for civil aviation (Finstad et al. 2002a). According to the IPCC Good Practice Guidance the methodology used is Tier 2 based on the detailed methodology in EEA (2001). This methodology allows estimation of emissions and fuel consumption for different types of aircrafts according to the average flying distance and numbers of landings and take-offs (LTO). The estimation is provided with new emission and fuel consumption factors for civil aircraft, except helicopters, from EEA (2013).

3.2.4.2 Road transport

NFR 1A3b i-v

Last update: 04.03.2016

3.2.4.2.1 Description

Emissions from this source includes combustion emissions from vehicles driven on roads, i.e. the categories passenger cars, light duty vehicles, heavy duty vehicles (including buses and coaches) and mopeds and motorcycles, as well as NMVOC emissions from gasoline evaporation. The methodology used for calculating emissions is described in more detail i Holmengen and Fedoryshyn (2015). The methodology corresponds to a Tier 3 methodology from the EMEP/EEA 2016 Guidebook (EEA 2016), using detailed information on vehicle fleet composition and driving patterns.

For passenger cars and light duty vehicles there has been a marked shift from petrol to diesel vehicles. In the 1990's petrol consumption within road transport far exceeded auto diesel. Distance driven by petrol passenger cars equally exceeded distances for auto diesel passenger cars. From 1.1.2007, there was a change in the registration tax for new passenger cars, and CO₂ became one parameter in calculating the level in addition to curb weight and engine power. This led to an increase in the sale of new diesel passenger cars. In 2006, the share of diesel vehicles within new passenger cars was 48 per cent. In 2007, the same share had increased to 74 per

cent and was steady on that level until 2011. Since 2012, a NO_x – component was added to the passenger cars taxation. The share of diesel cars sold has then been reduced. The share amounted to 49 per cent of diesel cars in 2014 and 41 per cent in 2015.

3.2.4.2.2 *Method*

The consumption of gasoline for road traffic is estimated as total sales minus consumption for other uses, i.e. a top-down approach. Other uses for gasoline are e.g. small boats, snow scooters and motorized equipment. For auto diesel, the total consumption in road traffic is all auto diesel charged with auto diesel tax, with two per cent addition for assumed tax free auto diesel used in road traffic. For the years prior to 1997, the auto diesel taxation was incomplete, and the consumption of auto diesel in road traffic was calculated as for gasoline, by subtracting the consumption for other uses. Other uses of auto diesel are e.g. motorized equipment in agriculture and construction. CNG and LPG are estimated by bottom-up approaches. The total consumption of each fuel is attributed to different vehicle classes based on results from the emission model of the Handbook of Emission Factors (HBEFA; (INFRAS 2009)).

Pollutants other than CO₂ are estimated by the emission model of the Handbook of Emission Factors (HBEFA; (INFRAS 2009)). The model uses a mileage approach: Emissions = mileage * emission per km. The model results are used directly without any adjustment for discrepancies between estimated and registered fuel consumption.

The HBEFA model provides emission factors and possibilities for calculating emissions for segments and sub-segments for six vehicle classes: passenger cars, light commercial vehicles, heavy commercial vehicles, urban buses, coaches and motorcycles (including mopeds). The segments are based on engine volume for passenger cars and motorcycles, total weight for heavy commercial vehicles, urban buses and coaches, and tare weight for light commercial vehicles. The segments are further disaggregated to subsegments based on fuel type and technology type (e.g. Euro-1 – Euro-5). The segments used for Norway in the HBEFA model are given in Table 3.7.

Table 3.7. Segments used for Norway in the HBEFA

Vehicle class	Segment	Fuel type	Segment split based on	Engine volume/weight class
Passenger car	PC petrol <1,4L	Petrol	Engine volume	< 1.4 litres
	PC petrol 1,4-<2L	Petrol	Engine volume	1.4- 2.0 litres
	PC petrol >=2L	Petrol	Engine volume	>= 2.0 litres
	PC diesel <1,4L	Diesel	Engine volume	< 1.4 litres
	PC diesel 1,4-<2L	Diesel	Engine volume	1.4- 2.0 litres
	PC diesel >=2L	Diesel	Engine volume	>= 2.0 litres
	PC LPG	LPG	-	All engine volumes
Light commercial vehicles	LCV petrol M+N1-I	Petrol	Tare weight	< 1305 kilos
	LCV petrol N1-II	Petrol	Tare weight	>= 1305-1760 kilos
	LCV petrol N1-III	Petrol	Tare weight	>= 1760-3859 kilos
	LCV diesel M+N1-I	Diesel	Tare weight	< 1305 kilos
	LCV diesel N1-II	Diesel	Tare weight	>= 1305-1760 kilos
	LCV diesel N1-III	Diesel	Tare weight	>= 1760-3859 kilos
Heavy goods vehicles	RT petrol	Petrol	-	All gross weights
	RigidTruck <7,5t	Diesel	Gross weight	<= 7.5 tonnes
	RigidTruck 7,5-12t	Diesel	Gross weight	> 7.5 - 12 tonnes
	RigidTruck >12-14t	Diesel	Gross weight	> 12 - 14 tonnes
	RigidTruck >14-20t	Diesel	Gross weight	> 14 - 20 tonnes
	RigidTruck >20-26t	Diesel	Gross weight	> 20 - 26 tonnes
	RigidTruck >26-28t	Diesel	Gross weight	> 26 - 28 tonnes
	RigidTruck >28-32t	Diesel	Gross weight	> 28 - 32 tonnes
	RigidTruck >32t	Diesel	Gross weight	> 32 tonnes
	Tractor for AT <=7,5t	Diesel	Gross weight	<= 7.5 tonnes
	Tractor for AT>7,5-14t	Diesel	Gross weight	> 7,5 - 14 tonnes
	Tractor for AT>14-20t	Diesel	Gross weight	> 14 - 20 tonnes
	Tractor for AT>20-28t	Diesel	Gross weight	> 20 - 28 tonnes
	Tractor for AT >34-40t	Diesel	Gross weight	> 34 - 40 tonnes
	Tractor for AT >40-50t	Diesel	Gross weight	> 40 - 50 tonnes
	Tractor for AT >50-60t	Diesel	Gross weight	> 50 - 60 tonnes
Coach	Coach Std <=18t	Diesel	Gross weight	<= 18 tonnes
	Coach 3-Axes >18t	Diesel	Gross weight	> 18 tonnes
Urban bus	Ubus Midi <=15t	Diesel	Gross weight	<= 15 tonnes
	Ubus Std >15-18t	Diesel	Gross weight	>15 - 18 tonnes
	Ubus Artic >18t	Diesel	Gross weight	> 18 tonnes
	Ubus Std >15-18t CNG	CNG	Gross weight	>15 - 18 tonnes
	Ubus Artic >18t CNG	CNG	Gross weight	> 18 tonnes
Motorcycles and mopeds	Moped <=50cc (v<50kmh)	Petrol	Engine volume	<= 50 cc
	MC 2S <=150cc	Petrol	Engine volume	<= 150 cc
	MC 2S >150cc	Petrol	Engine volume	>150 cc
	MC 4S <=150cc	Petrol	Engine volume	<= 150 cc
	MC 4S 151-250cc	Petrol	Engine volume	151-250 cc
	MC 4S 251-750cc	Petrol	Engine volume	251-750 cc
	MC 4S >750cc	Petrol	Engine volume	> 750 cc

Source: Statistics Norway/Norwegian Environment Agency

The model combines the number of vehicles within each segment with driving lengths for the same segments to produce annual national mileage per subsegment. For heavy goods vehicles, the vehicle number is corrected for vehicles driving with trailers, and the driving is split into three load classes (empty, half loaded and fully loaded).

The annual national mileage is split between shares driven in different traffic situations. The traffic situations are a combination of area (urban/rural), road type (e.g. trunk road and access road), speed limit and level of service (freeflow, heavy, saturated, and stop and go). The traffic situations are further disaggregated by gradients, where the amount of driving on roads with slopes ranging from -6 per cent to 6 per cent is specified for each traffic situation.

Hot emission factors are provided on the disaggregated level of subsegments and traffic situations with different gradients, and the emissions are estimated after these steps of disaggregation.

The HBEFA model provides emission factors for cold emissions and evaporative emissions (soak, running losses and diurnal), in addition to hot emission factors. In order to calculate cold and evaporative emissions, information on diurnal variation in curves of traffic, trip length distributions, parking time distributions and driving behaviour distributions must be provided, in addition to variations in mean air temperature and humidity.

3.2.4.2.3 Activity data

All activity data are, as far as possible, updated for every year of the inventory. Data are taken primarily from official registers, public statistics and surveys. However, some of the data are based on assumptions. Many of the data sources are less comprehensive for the earliest years in the inventory. The sources of activity data are listed below:

- *Total fuel consumption*: the total amounts of fuels consumed are corrected for off-road use (in boats, snow scooters, motorized equipment, etc.). These corrections are estimated either from assumptions about the number of units, annual operation time and specific fuel consumption, or from assumptions about and investigations of the fraction of consumption used off-road in each sector. Statistics Norway's sales statistics for petroleum products supply the data for total fuel consumption (Statistics Norway Annually-b). Fuel consumption in road transport is given in Table 3.8.
- *Number of vehicles*: the number of vehicles in the various categories and age groups is taken from the statistics on registered vehicles, which receives data from the official register of the Norwegian Directorate of Public Roads. The model input is number of vehicles per vehicle class for each inventory year, and the share of vehicles for any given combination of segment and fuel type. These data are combined with information on the introduction of technology classes to provide number of vehicles within each subsegment. The information on introduction of technology classes are for recent years based on information from the official register of the Norwegian Directorate of Public Roads, and on legislation for the years in which the information in the register is insufficient.
- The HBEFA model distinguishes between two types of buses: urban buses, mainly used for urban driving, and coaches, mainly used for rural and motorway driving. Due to lack of specific information to make this split in the national vehicle register, the distinction between urban buses and coaches are based on a methodology used in Sweden (Swedish environmental protection agency 2011), where the split is made based on the ratio p/w . Here, p is equal to the maximum allowed number of passengers (number of seats plus number of

allowed standing passengers), and w is equal to the gross vehicle weight. These data are available from the national vehicle register. Buses with a p/w -value above 3.75 are classified as urban buses, whereas buses with a p/w -value below 3.75 are classified as coaches.

- *Average annual mileage*: Mileages for passenger cars, light commercial vehicles, heavy goods vehicles, coaches and urban buses are from 2005 onwards based on odometer readings taken during annual or biannual roadworthiness tests. The readings are collected by the Directorate of Public Roads and further processed by Statistics Norway (Statistics Norway 2010a). For earlier years, most figures are determined from surveys by Statistics Norway or the Institute of Transport Economics. In some instances assumptions are needed.
- The statistics on number of vehicles depict the vehicle fleet per December 31st of the inventory year, while the statistics on mileages represents annual driving for the entire year, including vehicles that have been scrapped or in other ways been in the vehicle fleet for only parts of the inventory year. To adjust for this discrepancy for the years 2005-2011, mean annual driving lengths for each vehicle category have been adjusted upwards in such a way that the totals correspond to the total annual traffic activity from the statistics on annual driving lengths.
- The average annual mileages vary as a function of age, with older vehicles generally driving shorter annual distances than newer vehicles. The correction of driving as a function of vehicle age is based on odometer readings taken during the roadworthiness test. The functions are calculated as the mean of the years 2005-2011, and the same correction curve is used for all years.
- Motorcycles and mopeds are not subject to roadworthiness tests in Norway. Average annual mileages are taken from a report on transport volumes in Norway (Vågane & Rideng 2010). Due to lack of data, corrections of annual mileage as a function of age for motor cycles and mopeds are taken from a Swedish survey (Bjørketun & Nilsson 2007) under the assumption that annual mileages as a function of age are comparable in Norway and Sweden.
- *Load data* are taken from the Road goods transport survey (Statistics Norway 2010b).
- *Transformation patterns* are calculated using information from Statistics Norway' Road goods transport survey on use of trailers and trailer size (Statistics Norway 2010b).
- *Traffic situations*: The Directorate of Public Roads has data on the annual number of vehicle-kilometres driven on national and county roads. The data are allocated by speed limits, road type, area type (urban/rural), and vehicle size (small/ large). Traffic on municipal roads (approx. 15 per cent) is estimated by Statistics Norway based on road lengths, detailed population data, traffic on adjoining roads, etc. The HBEFA model has emission factors for different situations of traffic flow (freeflow, heavy traffic, saturated traffic, and stop and go). Assumptions have been made as to this distribution for the different combinations of area type, road type and speed limits for Norway. Effects of road gradients are included, based primarily on Swiss data supplied to the HBEFA.
- *Ambient conditions* (air temperature and humidity) are included in the model to calculate cold and evaporative emissions. An average of five larger Norwegian cities has been used for spring, summer, autumn and winter separately. The data are based on measurements from the Norwegian Meteorological Institute.
- *Trip length and parking time distributions* are calculated from the Norwegian travel survey (Institute of transport economics 1993). The distributions are given on an hourly basis.

Table 3.8. Fuel consumption in road transport, 1990-2015. TJ

Year	Petrol	Diesel	LPG	Gaseous fuels	Biofuels
1990	75 488	30 565	-	-	-
1995	70 325	41 741	-	-	-
2000	68 255	47 431	-	-	-
2005	66 608	65 388	-	-	126
2006	63 983	72 009	-	-	259
2007	60 512	78 380	161	108	1 424
2008	56 490	80 374	161	136	3 800
2009	52 810	81 720	161	144	4 479
2010	49 556	87 915	161	185	5 538
2011	45 013	91 754	161	212	5 557
2012	41 789	95 041	161	485	6 413
2013	39 126	97 829	161	589	5 934
2014	36 908	101 646	161	659	6 213
2015	34 633	104 451	161	433	7 169

Source: Statistics Norway

3.2.4.2.4 Emission factors

Emission factors are taken from the Handbook of Emission Factors (HBEFA). Factors are given as emission per vehicle kilometres for detailed combinations of subsegments and traffic situations.

BC is estimated as a fraction of PM_{2.5} emissions. BC emission factors depend on the vehicle category, the type of fuel and the existence of a treatment before exhaust. Shares of BC in PM_{2.5} were given by IIASA (Kupiainen & Klimont 2004).

HCB emissions have been estimated using an emission factor extracted from a former version of the EMEP guidebook (EEA 2007).

PCB emissions factors from Andrijewski et al. (2004) has been used to estimate emission from road transport. Table 3.9 presents PCB emissions factors.

Table 3.9. PCB emissions factors for gasoline and diesel combustion

Leaded gasoline	106	mg/tonne
Unleaded gasoline	0.02	mg/tonne
Diesel light vehicles	0.00000005	mg/tonne
Diesel, heavy vehicles	0.00000539	mg/tonne

Source: Andrijewski et al. (2004)

It has been assumed that PCB emissions vary with the gasoline lead content. Therefore, PCB emissions factor for combustion of gasoline in cars varies from 1990 to 1997 as it is presented in Table 3.10.

Table 3.10. PCB emissions factors for gasoline combustion for the period 1990-1997. Mg PCB/ tonne

	1990	1991	1992	1993	1994	1995	1996	1997
mg/tonne	106	82	74	49	8	8	0.2	0.02

Source: Statistics Norway

Biofuels for transport are not handled as separate fuels. The consumption is included with gasoline and autodiesel.

Average factors are listed in Appendix B.

3.2.4.2.5 Uncertainties

The uncertainty estimates are given in Appendix C.

The comparison of bottom up estimates of fuel consumption from HBEFA with total sales (source specific QA/QC) reveals a discrepancy of 5-15 per cent. This is deemed to be a reasonable difference. This discrepancy is handled differently for different emission components. The total consumption of each type of fuel is the most important parameter in relation to the reporting requirements of the UNFCCC, as this forms the basis for the calculation of CO₂ from road traffic (IPCC 2006). One kilogram of gasoline or autodiesel yields a fixed amount of CO₂ irrespective of vehicle type.

Guidelines for greenhouse gas reporting, the IPCC guidelines (IPCC 2006), states that CO₂ emissions should be calculated using fuel consumption, and that sold amount of fuel should form the basis. Calculations of emissions of CH₄, N₂O and many of the components of emissions reported to CLRTAP (e.g. NO_x and particulates), however, depends on more detailed information about vehicle types and driving patterns, and here a more detailed model (for example HBEFA) should be applied. The relationship between emissions and fuel consumption must be considered differently for the emission components that are directly dependent on the composition and quantity of fuel (CO₂, SO_x and heavy metals) and those who to a larger extent depend on the type of vehicle and driving mode (e.g. NO_x, CH₄, N₂O, NH₃, CO, particles).

Fuel consumption is not an input to HBEFA, where emissions are calculated based on mileage and number of vehicles in each subsegment of vehicle classes, as well as other data sets, such as cold start and age distribution of mileage. Fuel consumption is however calculated in the model similarly to emission calculations. The estimated fuel consumption for the country as a whole can be compared with sold fuel sales statistics for petroleum products and energy balance. The comparison shows that the fuel consumption calculated in HBEFA is systematically lower than the fuel in the energy balance, and that the difference is greater for auto diesel than for petrol. The difference has been between approximately 1 and 10 per cent for gasoline, and 4 and 15 per cent for diesel in the period 1990-2011. Exceptions are 1990 and 1991 for auto diesel when the difference was very small, and 1993, when the difference was almost 30 percent. There is no increasing or decreasing trend in the deviations, but there seems to be a correlation between the deviation of petrol and diesel.

It is not known why there is a discrepancy between the consumption in the energy balance and

bottom-up calculation in HBEFA, but there are several possible explanations as to why fuel sold does not match the fuel consumption calculated from the road transport emission model:

Fuel purchased by foreign vehicles. Foreign vehicles are not included in the vehicle register statistics, even though they drive on Norwegian roads. Similarly, no fuel bought by Norwegian vehicles abroad is sampled. It is likely that there is no systematic "fuel tourism" across the Norwegian border, as there are no significant price differences between fuel prices in Norway and Sweden. The current calculations are based on the assumption that driving in Norway by foreign vehicles equals the driving of Norwegian vehicles abroad.

1. *Vehicles drive longer in reality than what the model calculates.* Seeing that the Technical Inspection of vehicles is a new data source for mileage, it is hard to imagine that mileages in the model are systematically underestimated. Motorcycles do not have such a Technical Inspection. They can however not explain the discrepancy between the calculated use and the amount of fuel sold. For example, they mostly run on gasoline, while the largest deviation is within auto diesel.
2. *Driving patterns.* There may be elements in the driving patterns that causes fuel consumption per kilometer per vehicle to be higher than what the model calculates. One possible reason here is that the fuel consumptions stated in the vehicle type approvals are used as part of the input to the model, and there is an ongoing discussion about whether these systematically underestimate consumption. These data are however available only for the latter part of the series, and can not explain the discrepancies in the 1990s.

Whether the emission calculations should be corrected for differences in fuel consumption depends on the pollutants in question. For those components that are directly dependent on the amount of fuel (CO₂, SO₂, heavy metals) it will always be appropriate to use the fuel consumption from the energy balance as a basis for calculation. For the other emission components, the decision on whether to correct for total fuel consumption or not will depend on what is causing the discrepancy between fuel consumption calculated in the model and fuel consumption in the energy balance. If the reason is that the total mileage is underestimated in the model, and that the energy balance represents a "truer" picture of the consumption of fuels, emissions should be corrected. If the discrepancy, however, is due to an underestimation of the fuel consumption per kilometer, the emission estimates should not be corrected unless one finds a clear correlation between changes in consumption per kilometer and emissions per kilometer for the relevant emission components. As long as we do not know the reason for the discrepancy, an assessment of data quality in the various input data is crucial to determining whether emissions should be reconciled against fuel sales or not.

In the previous model (SFT 1993; SFT 1999b), the emissions of all substances were corrected to account for the discrepancy between the energy balance and the model calculations, because the energy balance was considered the most secure data source. When HBEFA was introduced as the computational model, a new data source was also introduced, namely the mileage

statistics from Statistics Norway. These statistics are based on data from periodical technical inspections, and goes back to 2005. This important new data source is considered to be of good quality, and it has changed the assessment of whether the emissions shall be corrected for the consumption in the energy balance or not. There is no reason to believe that the total run lengths are underestimated, and we consider it likely that the reason for the discrepancy lies in the estimates of fuel consumption per kilometer. The energy balance is based on the assessment that Norwegian purchases abroad correspond to foreign purchases in Norway, and the same assessment is applied to the emission calculations. We have not found any reason to believe that the reasons for the discrepancies in fuel consumption are directly correlated with driving behavior. It has therefore been assessed that HBEFA estimated emissions of pollutants that are not directly related to fuel consumption should not be reconciled with fuel consumption.

There are currently no comprehensive statistics on foreign vehicles driving in Norway. One possible explanation for the discrepancy between the calculated fuel consumption in HBEFA and sold quantity of fuel is that foreign driving in Norway exceeds Norwegian driving abroad. There has been an issue that the proportion of heavy vehicles with foreign vehicles increases. However, we see no clear increasing trend in the difference between the model results and sales. Better data related to foreign driving in Norway and the Norwegian driving vehicles abroad would strengthen or refute the current assumption that these two balance each other out.

3.2.4.2.6 Completeness

Major missing emission sources are not likely.

3.2.4.2.7 Source specific QA/QC

Top-down and bottom-up data on fuel consumption are compared for gasoline and diesel vehicles on an annual basis. The consumption of gasoline and auto diesel for road traffic is estimated as total sales minus consumption for other uses, i.e. a top-down approach. The HBEFA emission model also makes bottom-up estimates of consumption, which can be compared with the top-down data. The estimated fuel consumption from HBEFA deviates from the top-down estimate by approximately 5-15 per cent per year, with the higher value for auto diesel. The causes are on one hand uncertainties in the amount of non-road use and on the other hand uncertainties in mileage and specific consumption in road transport.

However, the total consumption of auto diesel is well known. The uncertainty concerns the allocation between road and non-road use, connected to illegal use of diesel without road tax in road traffic. The total emissions may be sensitive to this allocation, due to different emission calculation methodologies. When inspected, taxed and tax free diesel can be identified by colour. The fine for illegal use of tax free diesel is currently from NOK 20 000,- and upwards. There is no reason to believe that this is a major problem.

3.2.4.3 Railways

NFR 1A3c

Last update: 04.03.2016

3.2.4.3.1 Description

Railway traffic in Norway uses mainly electricity. Auto diesel is used at a small number of lines, for shunting etc. There is also a minor consumption of coal in museum railways.

3.2.4.3.2 Method

General estimation methodology for calculating combustion emissions from consumption figures and emission factors is used.

3.2.4.3.3 Activity data

Consumption figures for auto diesel used in locomotives are collected annually from the Norwegian State Railways. Consumption of coal is estimated based on information from different museum railways; the same figure is used for all years from 1990. There has been a continuous electrification of the railroads in Norway, and the diesel consumption is more than halved from the level in the early 1990's.

Table 3.11. Fuel consumption in railways, 1990-2015. TJ

	Liquid fuels	Solid fuels
1990	1306	3.9
1995	1405	3.9
2000	663	3.9
2005	581	3.9
2006	564	3.9
2007	593	3.9
2008	617	3.9
2009	612	3.9
2010	535	3.9
2011	514	3.9
2012	573	3.9
2013	634	3.9
2014	588	3.9
2015	619	3.9

Source: Statistics Norway

3.2.4.3.4 Emission factors

Emission factors for NO_x, HC, CO, and PM₁₀ were estimated by Bang (1993) based on a literature survey and data on Norwegian usage profiles. The HC factor of 4 g/kg was used directly for NMVOC.

The other emission factors are the same as for diesel machinery in mining and quarrying (see section 3.2.4.7.4), with the following exception:

- NH₃: 0.007 g/kg vs 0.005 g/kg for machinery (EEA 2013)
- BC: emission factors as share of PM_{2.5} from IIASA (Kupiainen & Klimont 2004) have been used. They are presented in Appendix B.

General emission factors for coal are used in the calculations.

3.2.4.3.5 *Uncertainties*

The consumption data are of high quality. Their uncertainty is estimated to be ± 5 per cent of the mean.

3.2.4.3.6 *Completeness*

Major missing emission compounds are not likely.

3.2.4.3.7 *Source specific QA/QC*

Consumption data from the Norwegian State Railways have previously been compared with sales to railways according to the Petroleum statistics. However, the latter includes some consumption by buses operated by the Norwegian State Railways. From 1998, the reported sales of "tax-free" auto diesel to railways have been higher than the consumption data from the Norwegian State Railways, although there was only a minor difference in 2009. The reason for this discrepancy has not been checked. "Tax-free" auto diesel is only for non-road use, so consumption by buses should not be the cause.

3.2.4.4 **Electric railway conductions**

NFR 1A3c

Last update: 01.09.05

3.2.4.4.1 *Method*

Electric railway conductions contain copper that is emitted in contact with trains. In the inventory, copper emissions are calculated by emission factors and activity data.

3.2.4.4.2 *Activity data*

The activity data used for calculating emissions of copper from electric wires are annual train kilometers given by the Norwegian State Railways (NSB).

3.2.4.4.3 *Emission factors*

According to Norwegian State Railways (Rypdal & Mykkelbost 1997) the weight of a contact wire is 0.91 kg/meters. The weight is reduced by 20 per cent after 3 million train passes. This gives an emission factor of 0.06 g/train kilometers. It is, however, uncertain how much of this is emitted to air. In the inventory it is assumed that 50 per cent is emitted to air. This gives an emission factor of 0.03 g/ train kilometer.

Table 3.12. Emission factor for electric railway conductions. g/km

Emission factor (g/train kilometers)	
Cu	0.03

Source: Norwegian Environment Agency

3.2.4.4.4 *Uncertainties*

The emission factor used is uncertain. First, there is an uncertainty connected to the reduction of 20 per cent after 3 million train passes. Secondly, there is uncertainty regarding the assumption that 50 per cent are emissions to air (Finstad & Rypdal 2003).

3.2.4.4.5 *Completeness*

No major components are assumed missing.

3.2.4.4.6 Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.5 Navigation

NFR 1A3d

Last update: 07.03.17

3.2.4.5.1 Description

According to CLRTAP Norwegian national sea traffic is defined as ships moving between two Norwegian ports. In this connection, installations at the Norwegian part of the continental shelf are defined as ports.

Fishing is described in section 3.2.5

3.2.4.5.2 Method

Emissions from navigation are estimated according to the Tier 2 IPCC methodology. Emissions from moveable installations used in oil and gas exploration and extraction are split between 1A1 – energy industries (section 3.2.2) and navigation: Emissions from drilling are reported under 1A1, while emissions from transport and other activities are reported under navigation. Emissions from international marine bunkers are excluded from the national totals and are reported separately (section 3.2.6.2), in accordance with the IPCC Good Practice Guidance.

Annual emissions are estimated from sales of fuel to domestic shipping, using average emission factors in the calculations.

For 1993, 1998, 2004 and 2007 emissions have also been estimated based on a bottom-up approach. Fuel consumption data were collected for all categories of ships (based on the full population of Norwegian ships in domestic transport); freight vessels (bulk and tank), oil loading vessels, supply/standby ships, tug boats, passenger vessels, fishing vessels, military ships and other ships. Emissions were estimated from ship specific emission factors and fuel use. From this information, average emission factors were estimated for application in the annual update based on fuel sales. This approach is unfortunately too resource demanding to perform annually.

3.2.4.5.3 Activity data

The annual sales statistics for petroleum products gives figures on the use of marine gas oil, heavy distillates and heavy fuel oil in domestic navigation. Information on fuel used in the ship categories in the bottom-up analysis is mainly given by data from the Business Sector's NO_x fund for 2007 and by earlier SSB analyses for 1993 and 1998 (Tornsjø 2001), and 2004. Data on fuel consumed by public road ferries are available from the Directorate of Public Roads.

Information on fuel use at mobile drilling rigs is taken from the sales statistics, but information

on use (whether it is used for drilling, stationary combustion etc.) is taken from Environmental Web/EPIM Environment Hub (EEH) (reported from oil companies to the Norwegian Environment Agency and the Norwegian Petroleum Directorate). Consumption during drilling activities are reported under "Energy industries" (CRF 1A1c). Only the remaining part of sales, assumed to be for drilling rigs during transit etc., is included with Navigation.

For marine gas oil, the amount used for navigation is equal to total sales figures except bunkers, after the deduction of estimated stationary use, mainly in oil and gas extraction, but also some minor use in manufacturing industries and construction. Due to inaccuracies in the reporting of distribution of marine gas oil between domestic and international shipping from approximately 2005, there is some uncertainty connected to the figures for the latest years.

Use of natural gas in navigation, which was introduced in 2000 and has increased considerably from 2007, is based on sales figures reported to Statistics Norway from the distributors. Fuel consumption in national navigation is given in Table 3.13.

Table 3.13. Fuel consumption in national navigation, 1990-2015. TJ

	Liquid fuels	Gaseous fuels
1990	22 817	-
1995	25 606	-
2000	30 117	46
2005	26 459	332
2006	27 641	329
2007	29 143	1 577
2008	24 699	1 856
2009	25 300	1 980
2010	26 691	2 157
2011	25 320	2 465
2012	23 180	3 174
2013	19 680	3 736
2014	15 843	4 222
2015	16 331	4 417

Source: Statistics Norway

3.2.4.5.4 Emission factors

Emission factors used for navigation are given in Appendix B, table B1, table B4 and tables B9-B12.

SO₂

The emission factors are determined from the sulphur content of the fuel.

NO_x

NO_x factors for different engine types (slow, medium and high speed) have been estimated by Marintek based on data from a comprehensive measure programme for NO_x emissions from ships, which has been implemented under the leadership of the Business Sector's NO_x fund. The

new basis factors from Marintek apply to emissions from different engine types built before and after emission restrictions were implemented in 2000 (Bremnes Nielsen & Stenersen 2009).

Table 3.14. Recommended emission factors for NO_x for different engine types

	Engine building year	
	Before 2000 kg NO _x /tonne fuel	After 2000 kg NO _x /tonne fuel
Slow speed NO _x factor	82	78
Medium speed NO _x factor	54	53
High speed NO _x factor	47	41

Source: Bremnes Nielsen and Stenersen (2009)

The factors were weighted in two steps: First, by engine type distribution within ship categories (passenger, general cargo, offshore, fishing, etc.). Secondly, by estimated fuel consumption among categories. The fuel consumption weights were calculated based on data for 1993, 1998, 2004 and 2007, which are years with good availability of activity data. Average factors for other years were interpolated. In the interpolation of the average factors over the time series, a peak in the use of shuttle tankers has been taken into consideration. The fact that we have reported data for public road ferries for some years, and a gradual change to new engines with lower emissions starting in 2000 due to new restrictions, has also been taken into consideration. The factors from Marintek are valid for engines with no particular NO_x reduction measures. The NO_x factors used in the inventory are documented in Flugsrud et al. (2010).

The method outlined above is used for the years up to 2007. From 2008 onwards, a large number of NO_x reducing technologies have been installed, funded through the NO_x fund and certified by emission measurements. Annual emissions are reported by companies to the NO_x fund and/or to the Norwegian Tax Administration as part of a national NO_x tax system.

In 2016, data on NO_x emissions and consumption of fuel in the third quarter of 2016 were collected from companies participating in the NO_x fund with ships operating in Norwegian coastal traffic. The data were made available to Statistics Norway for inventory preparation. The NO_x emissions in these data are mainly based on ship-specific measurements. For ships without measurements, a slightly adjusted version of the Marintek factors were used³. Based on these data, an average NO_x factor for domestic navigation in 2016 was calculated. Emission factors for 2008-2015 were obtained by linear interpolation.

For gas engines the NO_x factor 5.6 kg NO_x/tonne LNG is established based on the mass of LNG consumed (Bremnes Nielsen & Stenersen 2010).

For offshore drilling rigs, the factor 54 kg NO_x/tonne is used (Karlsson & Finborud 2012). See

³ Medium speed engines 200-1000 rpm and high speed engines >1500 rpm used factors from table 3.14. Slow speed engines <200 rpm used 100 kg NO_x/tonne. Medium/high speed engines 1000-1500 rpm used 50 kg NO_x/tonne. (Norwegian Tax Administration, <http://www.skatteetaten.no/globalassets/saravgifter/avgiftsrundskriv/2016-nox.pdf>)

further discussion on NO_x from offshore installations in the section on stationary combustion.

Average NO_x factors for fishing and for general shipping are given in Appendix B, table B11.

NH₃

Emissions of NH₃ from navigation are reported as "Not Estimated". The EMEP/EEA Guidebook (EEA 2013) has no emission factors, and in table 2-2 over "Contributions to total emissions" NH₃ is stated as "No emissions reported".

Particles

Factors for particulate matter are based on measurements performed by MARINTEK and literature sources. The factors are presented in Table 3.15.

Table 3.15. Particulate matter emission factors for oil and gas operated vessels

Fuel	Emission factor	
	PM _{2.5}	PM ₁₀ , TSP
Marine gas oil, light fuel oils (kg/tonne)	1.5	1.6
Heavy fuel oil, heavy distillate (kg/tonne)	5.1	5.4
LNG (kg/1000 Sm ³)	0.032	0.032

Source: Bremnes Nielsen and Stenersen (2010).and Bremnes Nielsen (pers.comm.4)

For oil based fuels it is assumed that all particles are included in PM₁₀ and 95 per cent of the particles are included in PM_{2.5} (Finstad et al. 2003).

Emission factors for particle emissions from gas operated vessels are based on measurements made by MARINTEK (Bremnes Nielsen, *pers.comm*), which show 95-99 per cent emission reduction compared to marine gas oil.

BC

BC emissions are estimated using shares of PM_{2.5} as emission factors. Factors from the IIASA (Kupiainen & Klimont 2007) have been used.

Table 3.16. BC emission factors for oil operated vessels

Fuel	Emission factor	
	PM _{2.5} (kg/t)	BC
Marine gas oil, light fuel oils	1.5	40%
Heavy fuel oil, heavy distillate	5.1	43%

Source: GAINS, IIASA (2010)

As no share for BC was found in the literature for the use of natural gas in navigation, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

⁴ Bremnes Nielsen, J. (2010): Personal information, email from Jørgen Bremnes Nielsen, 11 Nov. 2010, Marintek.

HCB

HCB emissions from the use of heavy fuel oil and marine gas oil have been estimated using the EMEP-EEA guidebook (2013).

Table 3.17. HCB emission factors for oil operated vessels

Fuel	Emission factor HCB (mg/t)
Marine gas oil	0.08
Heavy fuel oil	0.14

Source: EEA guidebook (2013).

PCB

PCB emissions from the use of heavy fuel oil and marine gas oil are considered higher in the navigation sector due to the presence of chlorine in the air. Emission factors determined by Cooper (2004) have been used to estimate PCB emissions.

Table 3.18. HCB emission factors for oil operated vessels

Fuel	Emission factor PCB (mg/t)
Marine gas oil	0.36
Heavy fuel oil	0.60

Source: Cooper (2004)

3.2.4.5.5 Uncertainties

The estimation of fuel used by fishing vessels is assumed to be rather uncertain. There is also uncertainty connected to the fuel use for other domestic sea traffic due to uncertainty in the sales statistics for petroleum products. Particularly, the delimitation between sales of marine gas oil for national use and bunkers has become more uncertain from approximately 2005, due to new and less accurate reporting routines in some oil companies.

Some uncertainty is also connected to the emission factors.

The uncertainty in the activity data for navigation is assessed to be ± 20 per cent. The uncertainty in the NO_x factors depends both on the uncertainty in the basis factors from Marintek (Bremnes Nielsen & Stenersen 2009) and on the uncertainty in the allocations that are made of the factors between ship types and years. Marintek has estimated the uncertainty in their basis NO_x factors for different engine types to ± 5 per cent. Uncertainties in emission factors are shown in Table 3.19.

Table 3.19. Uncertainties in emission factors for ships and fishing vessels. Per cent

	Standard deviation (2σ)
SO_2	± 25
NO_x ¹	± 15
NMVOC	± 50

¹ It is assumed that the uncertainty might be lower now than in this estimate from Rypdal and Zhang (2001) since more measures have been performed in connection with the Business Sector's NO_x fund.

Source: Rypdal and Zhang (2001)

3.2.4.5.6 *Completeness*

Major missing emission sources are not likely.

3.2.4.5.7 *Source specific QA/QC*

As mentioned, emission estimates for ships have been made bottom-up for 1993 and 1998 (Tornsjø 2001) and for 2004 and 2007. These results have been compared with top-down data (from sales) on fuel consumption used in the annual estimates.

The outcome showed that data from sales were only 1 per cent higher than data from reported consumption in 2007. For 2004 the sales data were 27 per cent higher than the consumption data in the bottom-up analysis. This can be explained by the fact that the bottom-up method does not cover all ships, but it may also be that the domestic/international distinction is not specified precisely enough in the sales statistics. Another element, which not has been taken into account, is possible changes in stock. For the years 1993 and 1998 a deviation of -12 and -15 per cent, respectively, has been found. In the calculations, sales figures are used, as they are assumed to be more complete and are annually available.

3.2.4.6 **Pipeline**

NFR 1A3e i

Last update: 22.03.10

Figures on natural gas used in turbines for pipeline transport at two separate facilities are reported annually from the Norwegian Petroleum Directorate to Statistics Norway. However, energy generation for pipeline transport also takes place at the production facilities. Specific data on consumption for transport are not available. Thus, the consumption at the two pipeline facilities does not give a correct picture of the activity in this sector. As a consequence, all emissions from pipelines are reported under NFR 1A1.

3.2.4.7 **Motorized equipment**

NFR 1A2f g-vii etc.

Last update: 10.03.2015

3.2.4.7.1 *Description*

The category "motorized equipment" comprises all mobile combustion sources except road, sea, air, and railway transport. Farm and construction equipments are the most important categories. Other categories include mines and quarries, forestry, snow scooters, small boats and miscellaneous household equipment.

Emissions from motorized equipments are reported under several categories:

Table 3.20. Motorised equipment categories

	NFR
Manufacturing and construction	1A2g-vii
Commercial and institutional	1A4a-ii
Households	1A4b-ii
Agriculture/Forestry/Fishing	1A4c-ii
Military	1A5b

Source: Statistics Norway/Norwegian Environment Agency

Primarily consumption of gasoline and auto diesel is considered. A small amount of fuel oil used for equipment in construction is also accounted for.

3.2.4.7.2 Method

Emissions are estimated through the general methodology described earlier, involving consumption figures and appropriate emission factors.

3.2.4.7.3 Activity data

Gasoline and auto diesel are handled differently. Consumption of *gasoline* is estimated bottom-up for each type of machinery based on data on the number of each type of equipment, usage and specific consumption.

Snow scooters: Number of equipment is obtained annually from the Norwegian Public Roads Administration. We assume a mileage of 850 km/year and a specific consumption of 0.15 l/km (National institute of technology 1991). A portion of 16 per cent of petrol consumption in agriculture is assigned to snow scooters. The remaining snow scooter fuel consumption is assigned to households.

Chainsaws and other two-stroke equipments: Only consumption in forestry is considered, based on felling data. Felling statistics are gathered by Statistics Norway. 50 per cent is supposed to be felled with use of chain saws, with a consumption of 0.33 l/m³. Note: Consumption has been kept fixed since 1994 based on a calculation by the Institute of Technology (Bang 1996).

Lawn mowers and other four-stroke equipment: Only consumption in households is considered.

Consumption of *auto diesel* is based on data from the energy balance. Auto diesel used in off-road vehicles has no road tax from 1993. Total use of auto diesel in motorized equipment is given as the difference between total sales of tax free diesel and estimated use for railway transportation. It is important to bear in mind that the total consumption of auto diesel in motorized equipment from 1993 is considered being of good quality since there from 1993 is no road tax on this part of the auto diesel. There is CO₂ tax on the auto diesel used for motorized equipment as well as for road traffic.

Distribution of auto diesel between industries: From 2001, a certain fraction of the consumption in a number of industries is allocated to motorized equipment, based on the distribution of taxed and tax-free diesel – the latter is meant for use outside road transport. The distribution formulas are based on figures from the statistics on energy use in manufacturing industries or

from the sales statistics. The statistics on energy use in manufacturing industries did not have such a split before 2001, and therefore distribution formulas for 2001 are used for 2000 and earlier years.

3.2.4.7.4 Emission factors

Emission factors used are given in Appendix B.

For diesel machinery, emission factors for HC, CO, and PM₁₀ were estimated by Bang (1993), based on a literature survey and data on Norwegian usage profiles. Source for emission factor for NO_x from diesel machinery is from Bang (1993) for motor gasoline and light fuel oils. For autodiesel emission factors from a Danish report (Winther & Nielsen 2006) is used. NMVOC factors were calculated by subtracting an assumed CH₄ fraction of 0.3 g/kg diesel.

Emission factors for tractors are used for tax-free auto diesel consumption in agriculture and forestry, while emission factors for construction machinery are used for tax-free auto diesel consumption in all other industries and households.

The emission factors used in the emission model are calculated from the basic factors in Winther and Nielsen (2006), weighted by the age and engine rating distribution of the tractor and construction machinery populations, as well as assumptions on motor load and operating hours and the introduction scheme for emission regulations by the EU (Stage I, II, III and IV).

3.2.4.7.5 Uncertainties

The estimates of consumption are considered quite uncertain, particularly for gasoline. However, the total consumption of gasoline is well known. For auto diesel, some uncertainty in the delimitation of different middle distillates may cause variations in figures on total use in motorized equipment between years.

3.2.4.7.6 Completeness

Major missing emission sources are not likely.

3.2.4.7.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.8 Automobile tyre and brake wear

NFR 1A3b vi

Last update: 22.12.16

3.2.4.8.1 Tyre wear

Description

Tyre wear is a source for emission of particles, heavy metals and persistent organic pollutants. The tyres are worn down by 10 to 20 per cent of its total weight during its lifetime. Most of the rubber is lost during acceleration and braking. All rubber lost is assumed to be particles containing heavy metals and PAH.

Method

Particles

All rubber lost is assumed to be small particles. The emissions of particles are calculated based on emission factors and annual mileage.

Heavy metals

Rubber particles contain heavy metals. Emissions of the heavy metals As, Cd, Cu, Cr, Pb and Hg are calculated based on annual mileage and emission factors.

PAH

The particles emitted from tyre wear contain PAH. Emissions are calculated based on emission factors and annual mileage.

Activity data

Annual mileage is used for calculating the emissions from tyre wear. Annual mileage is given by the road traffic model, see section 3.2.4.2.

Emission factors

Particles

The emission factors used for calculating the emission of particles are given by TNO (Institute of environmental and energy technology 2002). The emission factors are based on several Dutch and British studies. Recommended emission factors for TSP and PM₁₀ are taken from (Institute of environmental and energy technology 2002). Emission factor for PM_{2.5} was set to be zero. A new report TNO (TNO 2008) presents emission factors for all three fractions of particulate matter. The emission factors for TSP and PM₁₀ are in the same range as the emissions factors given in Institute of environmental and energy technology (2002). In the Norwegian inventory, it has been chosen to include PM_{2.5} emissions using the same ratio between PM₁₀ and PM_{2.5} as the ratio between PM₁₀ and PM_{2.5} from TNO (2008). The emission factors used are given in Table 3.21.

Table 3.21. Emission factors for particles from tyre wear. kg/mill. km

	TSP	PM₁₀	PM_{2.5}
Private cars	69	3.45	0.69
Van	90	4.5	0.9
Heavy duty vehicles	371,25	18.563	3.71
MC	34,5	1.725	0.35

Source: TNO (Institute of environmental and energy technology 2002)

BC

BC is estimated as a fraction of PM_{2.5}. Emission factors depend on the type of vehicle. IIASA (Kupiainen & Klimont 2004) gives emission factors for Black Carbon and Organic Carbon as share of TSP. Since the sum of emissions of BC and OC has to be lower than PM_{2.5} emissions, the emission factors have been adjusted.

Table 3.22. Emission factors for BC from tyre wear in share of PM_{2.5}. Particles are shown in kg/mill. km

	TSP	PM_{2.5}	BC
Passenger cars	69	0.69	30%
Light duty vehicles	90	0.9	30%
Heavy duty vehicles	371,25	3.71	30%
MC	34,5	0.35	30%

Source: IIASA (Kupiainen & Klimont 2004)

Heavy metals

The emission factors used for the heavy metals As, Cd, Cu, Cr and Pb are derived from a particle-heavy metal distribution given by Dutch studies (van den Brink 1996). The content of heavy metals in the particles, given by this distribution, is multiplied by the PM₁₀ emission factor (Table 3.21). This gives the emission factors for the heavy metals As, Cd, Cu, Cr and Pb from tyre wear (Table 3.23).

Table 3.23. Emission factors for heavy metals from tyre wear. g/mill. Km

	As	Cd	Cu	Cr	Pb
Private cars	0.003	0.007	1.691	0.014	0.552
Van	0.005	0.009	2.205	0.018	0.720
Heavy duty vehicles	0.019	0.037	9.096	0.074	2.970
MC	0.002	0.003	0.845	0.007	0.276

Source: van den Brink (1996)

The emission factor used for the estimation of the emissions of Hg is 0.079 g/ mill. km. This emission factor is derived from a study of heavy metal content in tyres (Bækken 1993) and an estimate of the amount of tyre in Norway in 1993 of 6000 tonnes (Finstad et al. 2001).

PAH

Emission factors for PAH are given in Finstad et al. (2001), but there is no information about how much of the emissions that are emitted to air, and how much that goes to soil and to water. All emissions are therefore supposed to be emitted to air. There is also no PAH profile available, so in lack of other data the same PAH profile as for burning of tyres is used (EPA 1998). PAH emission factors for tyre wear are given in Table 3.24. There are no data available for Benzo(a)pyrene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene. All PAH-4 emissions are assumed to be benzo(b)fluoranthene. The PAH-4 profile is given in table 3.25.

Table 3.24. Emission factors for PAH from tyre wear. g/mill. Km

	PAH	PAH-4
Light duty vehicles	10.4	6
Heavy duty vehicles	0.1	0,035

Source: Finstad et al. (2001)

Table 3.25. PAH profile road dust, also used for tyre wear (only PAH-4 is shown)

	Per cent
Benzo(a)pyrene	..
benzo(b)fluoranthene	100
benzo(k)fluoranthene	..
indeno(1,2,3-cd)pyrene	..

Source: Finstad et al. (2001)

Uncertainties

The calculation of emissions from tyre wear is uncertain. First, the emission factors for particles used are based on international studies and not on Norwegian conditions. There is also uncertainty concerning how much of the particles that are emitted to air. According to a Dutch judgement, all particles emitted to air are PM₁₀. This is however only a judgement, and not based on scientific research.

The heavy metal emission factors are based on the particle emission factors for PM₁₀, and since this factor is uncertain, the heavy metal emission factors will also be uncertain. The content of heavy metals in the particles emitted from tyre wear is based on a Dutch study and can therefore differ from Norwegian conditions and type of tyres used.

Completeness

Tyre wear also leads to emissions of other heavy metal components, such as zinc, nickel etc., but these components are not included in the Norwegian emission inventory.

Until 2004, different methods for calculating the emissions of heavy metals from tyre wear were used. One method was used for calculating emissions of Pb, Cd and Hg (Finstad et al. 2001) and another for calculating emissions of Cu, Cr and As (Finstad & Rypdal 2003). From 2004, the same method has been used for all the heavy metal components.

Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.8.2 Brake wear

Description

Brake blocks will wear during braking and this generates dust containing various metals. In the inventory, emissions of particles and heavy metals are included from this source.

Method

Particles

Emissions of particles are calculated based on emission factors recommended by an annual mileage.

Heavy metals

Emissions of lead, copper and chromium are calculated after a method described in SLB (Stockholms luft- och bulleranalys 1998a). The calculations are based on annual brake wear, driven kilometers and the brake blocks' metal content.

Brake wear, private cars and vans

To calculate emissions, brake wear first has to be estimated. It is assumed that private cars change brake blocks every fourth year. The background for this assumption is that private cars,

by normal driving, change brake blocks at front after 30 – 40 thousand kilometers and at the back after 60 – 80 thousand kilometers. A private car drives in average 150 thousand kilometers each year. Assuming that the brake blocks are changed after 60 thousand kilometers, the car will be four years old when blocks first are changed.

The brake blocks at front weigh 0.13-0.15 kg and 0.09-0.11 kg at the back. It is assumed in the calculations that the brake blocks weigh 0.15 kg at the front and 0.11 kg at the back, that the brake blocks are worn 70 per cent before they are changed and that the front and back blocks are changed after 40 and 60 thousand kilometers, respectively. Brake wear per kilometer are given equations (3.4) and (3.5):

$$(3.4) \quad \text{Front brake blocks (private cars): } 0.7 \cdot 4 \cdot 0.15 / 40\,000$$

$$(3.5) \quad \text{Back brake blocks (private cars): } 0.7 \cdot 4 \cdot 0.11 / 60\,000$$

The same method is used for calculating emissions from brake wear for vans and minibuses.

Brake wear, heavy duty vehicles

The number of brake blocks at a heavy duty vehicle varies with both brand and model. It is assumed that each front brake block weighs 2.5 kg and 3.5 kg at the back (Stockholms luft- och bulleranalys 1998a). This means that a truck with four wheels have 12 kg of brake blocks. It is assumed that the blocks are changed after 100 thousand kilometers when the brake blocks are worn 70 per cent.

Metal content

The metal content in the brake blocks for cars have been tested (Stockholms luft- och bulleranalys 1998a). For calculating the emissions from brake blocks, annual brake wear has been multiplied by the metal content. The metal content in the brake blocks in front of the car differs from the content in the brake blocks at the back (Table 3.26). For heavy duty vehicles, the metal content is independent of age or type of brake block.

Table 3.26. Metal content in brake blocks. mg/kg

	Private cars		Heavy duty vehicles
	Front	Back	Front and back
Cr	137	73.4	165
Cu	117941	92198	9031
Pb	9052	18655	457

Source: Stockholms luft- och bulleranalys (1998a); Stockholms luft- och bulleranalys (1998b)

How much of the heavy metal emissions that are emitted to air were investigated by Sternbeck *et al.* (2001). Tunnel experiments showed that approximately 20 per cent of the brake wear emissions were emitted to air. This result is used in the calculations of brake wear emissions.

Activity data

For calculating the emissions of particles, are annual mileage given by the road traffic model, see section 3.2.4.2.

For calculating the emissions of heavy metals, annually driven kilometres are also given by the road traffic model.

Emission factors

Particles

Emission factors recommended by TNO (Institute of environmental and energy technology 2002), based on different European studies, are used (Table 3.27).

Table 3.27. Particle emission factors for brake wear. kg/mill. km

	PM_{2.5}	PM₁₀	TSP
Private cars (BM1+DM1)	6	6	6
Van (BN1+DN1)	7.5	7.5	7.5
Heavy duty vehicles	32.25	32.25	32.25
MC	3	3	3

Source: TNO (Institute of environmental and energy technology 2002)

BC

BC is estimated as a fraction of TSP from emission factors depending on the vehicle type, given by IIASA (Kupiainen & Klimont 2004).

Table 3.28. Emission factors for BC from tyre wear in share of TSP. Particles are shown in kg/mill. km

	TSP	BC
Passenger cars	6	1%
Light duty vehicles	7.5	1%
Heavy duty vehicles	32.25	1%
MC	3	1%

Source: IIASA (Kupiainen & Klimont 2004)

Heavy metals

Emission factors for Cr, Cu and Pb are derived based on the above information and are given in Table 3.29.

Table 3.29. Heavy metal emission factors for brake wear. g/mill. km

	Private cars and vans	Heavy duty vehicles
Cr	0.36	14.82
Cu	342.33	303.44
Pb	38.16	40.95

Source: Statistics Norway

Uncertainties

There is high uncertainty in different steps in the emission calculations of heavy metals from brake wear, since many assumptions have been done. For example, there is uncertainty connected to the weight and the metal content of the brake blocks, and to the number of driven

kilometres before blocks are changed.

Completeness

Brake wear also leads to emissions of other heavy metal components, such as zinc, nickel etc., but these components are not included in the Norwegian emission inventory.

No other major emission components are assumed missing.

Source specific QA/QC

There is no specific QA/QC procedure for this source. See section 1.6 for the description of the general QA/QC procedure.

3.2.4.9 Automobile road abrasion

NFR 1A3bvii

Last update: 22.12.16

3.2.4.9.1 Description

Asphalt dust is emitted to air while using studded tires. The abrasion layer on asphalt roads can contain approximately 90 per cent stones (rock/minerals) and 5 per cent filler. The rest is bitumen. During studded tyre abrasion, stone materials are worn down to minor particles and will together with detached filler and bitumen whirl up and become airborne. How much dust/particles studded tires generate depends on:

- Weight of the stud
- The road surface resistance against abrasion
- Vehicle velocity
- Share of heavy vehicle
- If the road surface is dry, wet or ice coated

A great share of the dust from studded tyres will bind up to the water film when the road surface is wet. Some of it will however whirl up again when the road surface dries up. This is not included in the calculation.

Bitumen is a mixture of a great number of organic components, including PAH components. The emissions of benzo(b)fluoranthene, from road abrasion are calculated and included in the emission inventory. Calculated emissions of Cd are also included.

PM emissions from road abrasion are declining due to implementation of measures. In the largest cities there is a tax to pay when you drive with studded tyres in the city. This, together with information from the authorities about problems caused by PM, has reduced the numbers of cars with studded tyres both in the cities and all over the country. In addition, the weight of the stud has been reduced and hence also the emissions of PM. Consequently, the emissions are decreasing even though the annual total driving length is increasing. In contrast, emissions from automobile tyre and brake wear are calculated by multiplying the driving length with an emission factor, not taking into account the type of tyres. Since the driving length is increasing, the emissions from this source category also increase.

3.2.4.9.2 Method

Particles

PM₁₀

The method is prepared by TI/SINTEF and documented in SFT (1999b). For calculating average emission Q (ton/year) of PM₁₀ formula (3.6) is used:

$$(3.6) \quad Q_{PM10} \text{ (ton/year)} = \sum_{\text{All vehicle categories}} SPS * n * l * m * p * w * \alpha / 10^6$$

SPS: The specific wear of studded tyres (SPS). Gives an estimate of how much of the road surface that is worn off on one road kilometer of a vehicle with studded tyres

n: Number of cars of a vehicle category in the area

l: Annual mileage for a vehicle category in the area

m: Part of the year with studded tyres in the area (between 0 and 1)

p: Share of the vehicle category using studded tyres

w: Correction factor for wet and frozen road surface. In the calculation of w, frozen surface is given 0, wet surface 0.5 and dry surface 1. If the mileage with studded tyres on a wet and frozen surface respectively is v and x, $w = (0.05*v) + (1*(1-v-x))$

α: Share of the road dust in air that is PM₁₀. There is no data for this factor. The share of PM₁₀ on ground is used as a reference. There is very varied data for the size of this factor (Hedalen 1994). Hedalen gives a PM₁₀ share of 3-4 per cent. In the calculations 3 per cent is used as a first estimate. Hedalen (1994) states further that the PM_{2.5} share of total road dust is 0.5-1 per cent.

The road surface has stronger wear resistance on roads with heavy traffic than on roads with little traffic. The SPS value can therefore vary with the amount of traffic. SPS values for different ÅDT⁵ intervals were estimated based on analysis of track depths over the years 1988-1995 (Norwegian public roads administration 1996).

SPS is also dependent on the weight of the studs. The studs have in the recent years become lighter. The requirement in 1988 was that the stud on light vehicles should not exceed 2.0 gram, in 1990 this was changed to 1.8 gram, and it changed again in 1992 to 1.1 gram (Norwegian public roads administration 1997). The so-called "light studs" has a weight on 0.7 gram. Studs used on tyres for heavy vehicles could until 1992 weigh 8.0 gram, but this demand was changed to 3.0 gram. There are also other factors influencing the SPS values, for example the road surface wear resistance and the quality of the stone materials used.

SPS values used in the calculations are given in Table 3.30. The SPS values are divided on classes of ÅDT (Evensen, *pers. comm.*⁶). Values are given for 1993-1997 and a prediction for 2002. For the years in between, a moving average is calculated. For the years after 2002, the 2002 SPS

⁵ ÅDT = Average annual daily traffic

⁶ Evensen, R. (2007). Note for Johnny Johansen. 14/12 1997. Bærum: ViaNova.

values are used. In the calculations average values for SPS, weighted after the size of traffic load on roads with different ÅDT, are used. The values are given in g/km and are valid for all vehicles. To estimate how much of the emissions that originate from heavy vehicles, it is provided that heavy vehicles wear 5 times more than light vehicles. The vehicle velocity is not given as an own factor, since it is included in the calculation of SPS.

Table 3.30. SPS values. g/km

ÅDT	1973-1980	1981-1987	1988-1992	1993-1997	2002-
0-1500	22	20	20	18	16
1500-3000	20	20	18	16	14
3000-5000	16	15	14	12	10
>5000	14	12	11	10	9
Average ¹	17.1	15.6	14.7	13.1	11.6

¹ Weight after traffic load on roads with different ÅDT.

Source: Evensen, pers.comm.⁶

Annual traffic load (trafikkarbeid) ($n \cdot l$ in the formula) used in the calculations are annual mileage given by the road traffic model, see section 3.2.4.2.

Use of studded tyres is forbidden in Norway from the first Monday after Easter and until 31st of October. There is an exception from this rule in the three northern counties, Nordland, Troms and Finnmark. In these counties, use of studded tyres is forbidden between 1st of May to 15th of October. It is assumed in the calculations that studded tyres are used the whole period when it is allowed. This means that m is 6.5/12 in the northern counties and 5.5/12 for rest of the country.

Shares of traffic load on studded tyres in the five largest towns in Norway are given in Table 3.31. There has been a decrease in use of studded tyres in Norway during the latest years. The factor p in the formula will therefore vary from one year to another. Information regarding the share of studded tyres originates from the Norwegian Public Roads Administration. There is also national data on share of the car fleet with studded tyres. The data material is based on interviews of car drivers (Norwegian public roads administration 1995a; Norwegian public roads administration 1995b; Norwegian public roads administration 1998). The questionnaires were given out at daytime and caused that most of the answers were from local car drivers. Accordingly, the survey included too many car drivers with annual mileage over 20 000 km. The survey from 1997 was however done differently. In the calculation program, the studded tyre share was decided to be 0.2. This value was adjusted by the different local road administrations, based on interviews or other available knowledge. In 2000, the Norwegian Public Roads Administration made a new investigation over local use of studded tyre (Johansen & Amundsen 2000). In 2006, Gjensidige made a survey over the use of studded tyres in different counties in Norway, winter 05/06 (Vaaje 2006). For 2001-2004 averages of the two investigations are calculated for the counties. For the five largest cities, data from the Norwegian Public Roads Administration was used also for 2001-2005, but for the rest of the country the results from Gjensidige (Vaaje 2006) was used. The data are given in Table 3.32. For the period 1973-1990 it is assumed that the studded tyre share was 90 per cent.

Table 3.31. Use of studded tyres in five prioritized communities. Share of traffic load with studded tyres.

Light duty vehicles

	1998/ 1999	1999/ 2000	2000/ 2001	2001/ 2002	2002/ 2003	2003/ 2004	2004/ 2005	
Oslo	51.9	32.4	21.2	31.3	29.2	28.4	24.0	
Drammen	49.6	48.7	52.1	41.8	42.3	40.6	31.5	
Stavanger	38.1	31.3	26.8	29.3	28.8	35.2	30.1	
Bergen	37.0	29.4	28.3	31.0	30.7	30.4	30.3	
Trondheim	67.0	64.4	62.1	44.4	40.2	38.8	38.1	
	2005/ 2006	2006/ 2007	2007/ 2008	2008/ 2009	2010/ 2011	2011/ 2012	2012/ 2013	2013/ 2014
Oslo	19.9	20.3	17.0	16.4	14.4	16.1	15.2	15.2
Drammen	27.0	28.0	27.3	22.9	25.2	25.0	20.6	20.6
Stavanger	32.2	28.4	33.2	19.6	27.9	28.9	26.8	26.8
Bergen	29.6	21.4	10.5	14.7	12.3	18.0	16.6	16.6
Trondheim	32.9	31.2	19.4	28.6	25.8	28.4	35.3	35.3
	2014/ 2015	2015/ 2016						
Oslo	15	14						
Drammen	20	20						
Stavanger	22	24						
Bergen	14	13						
Trondheim	36	36						

Source: The Norwegian Public Roads Administration

Table 3.32. Averaged studded tyre share in Norway weighted by traffic load in the different counties. Light duty vehicles

Year	
1990	0.90
1991	0.87
1992	0.88
1993	0.88
1994	0.87
1995	0.86
1996	0.83
1997	0.79
1998	0.70
1999	0.63
2000	0.58
2001	0.56
2002	0.55
2003	0.53
2004	0.51
2005	0.49
2006	0.48
2007	0.46
2008	0.45
2009	0.44
2010	0.43
2011	0.43
2012	0.42
2013	0.41
2014	0.40
2015	0.39

Source: Statistics Norway based on data from the Norwegian Public Roads Administration and Gjensidige

To calculate the correction factor for humid road surface, traffic load data is used. This is divided into different road conditions after Evensen (*pers. comm.*⁷) (Table 3.33). Share of wet and dry road surface will change some as a consequence of varied share of studded tyres. In the calculations for 1973-1997 a correction factor is used, based on the estimation that 80 per cent of light duty vehicles and 60 per cent of heavy duty vehicles use studded tyres.

⁷ Evensen, R. (1997) Personal information, telephone call 20/11 1997. Bærum: ViaNova.

Table 3.33. Grouping of wet, dry and icy road surface

	In the Norwegian emission inventory
Wet	Wet
Dry	Dry
Slush	Wet
Loose snow	Wet ¹
Hard snow	Hard snow/ice
Bare tracks	80 per cent dry and 20 per cent wet ²

¹ Assumption made of NILU and Statistics Norway.

² Assumption made by Evensen (pers. comm.,7).

Source: Statistics Norway

TSP

Hedalen and Myran (1994) analysed road dust depots from Trondheim and found that 30 weight percentage of the particles were below PM₁₀. This gives a distribution where PM₁₀ is 0.3*TSP.

This distribution is used in the inventory.

Cd

Emissions of Cd are calculated based on emission factors from Bækken (1993) and annually generated road dust of PM₁₀.

Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4)

Emissions of PAH are calculated based on emission factors from Larssen (1985) and annually generated road dust of PM₁₀.

3.2.4.9.3 Activity data

Cd and PAH

The activity data used for calculating the emissions of Cd and PAH are annually generated PM₁₀ of road dust, see section 3.2.4.9.2.

3.2.4.9.4 Emission factors

Particles

The emission factors can be derived from the factors given under section 3.2.4.9.2. The emission figures are calculated as a product of SPS values for the given year, the number of kilometers driven, part of the cars with studded tyres, part of the year with winter season, correction for icy surface and the PM₁₀ share of the emission (α). The emission factors do not reflect the whirl up of road dust. Heavy duty vehicles whirl up much more than light duty vehicles.

BC

BC is estimated as a fraction of PM_{2.5}. Emission factors depend on the type of vehicle. IIASA (Kupiainen & Klimont 2004) gives emission factors for Black Carbon and Organic Carbon as share of TSP. Since the sum of emissions of BC and OC has to be lower than PM_{2.5} emissions, the emission factors have been adjusted.

Table 3.34. Emission factors for BC from tyre wear in share of TSP

	BC
Light duty vehicles	0.83%
Heavy duty vehicles	0.83%

Source: IIASA (Kupiainen & Klimont 2004)

Cd

The Cd content in the bitumen is uncertain. According to Bækken (1993), the Cd content varies between 1.9 and 43 g Cd per tonne road dust. Statistics Norway has chosen an average emission factor of 22.5 g/ton, see Table 3.35.

Table 3.35. PAH and Cd emission factors from road dust¹. g/tonne. PM₁₀ of road dust

	Emission factor (g/tonne PM₁₀ from road dust)
Norwegian standard (PAH-total)	61.7
Benzo(a)pyrene,	..
benzo(b)fluoranthene	5,5
benzo(k)fluoranthene	..
indeno(1,2,3-cd)pyrene,	..
Cd	22.5

¹ Dry road surface.

Source: Finstad et al. (2001)

PAH

The PAH content in the bitumen is uncertain and can vary over time. According to Larssen (1985), the PAH content in airborne dust from wet roads is 330 ppm and 75 ppm from dry roads. Statistics Norway has chosen 85 ppm. In Table 3.35, the emission factor of 85 g/ton is converted to correspond to the PAH components included in NS9815. This gives an emission factor of 61.7 g/ton for PAH-total.

3.2.4.9.5 Uncertainties

Particle distribution of road dust has also been investigated by others than Hedalen and Myran, among them the Norwegian Institute for Air Research (NILU). The results from these measurements show another distribution than Hedalen and Myran, with a PM₁₀-fraction much lower than 30 weight percentage. In the calculation of PM₁₀, data from Hedalen and Myran (1994) are used, and for consistency reasons the same source is used for estimating TSP, despite the uncertainty and the discrepancy with NILU's estimations.

The value of α (PM₁₀ share in road dust) is very uncertain. An average velocity is assumed in the calculations. This is further complicated when road surface on roads with high velocities have another wear resistance than other road surfaces.

The emission factor used for calculating Cd emissions is uncertain since it is based on two measurements.

The estimation of the PAH content in road dust from Larssen (1985) is very uncertain, since it is based on only one measurement in Oslo, but it is the only estimate available, and is used in lack

of other data.

3.2.4.9.6 Completeness

Major missing emission sources are not likely.

3.2.4.9.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

3.2.5 Other sectors

NFR 1A4/1A5

Last update: 10.03.2015

3.2.5.1 Description

The source category "Other sectors" includes *all* military combustion, *stationary* combustion in agriculture, forestry, fishing, commercial and institutional sectors and households, motorized equipment and snow scooters in agriculture and forestry, and ships and boats in fishing.

3.2.5.2 Activity data

Motorized equipment is described in section 3.2.4.7.

Households

Use of wood in households for the years 2005-2011 are based on responses to questions relating to wood-burning in Statistics Norway's Travel and Holiday Survey. The figures in the survey refer to quantities of wood *used*. The survey quarterly gathers data that cover the preceding twelve months. The figure used in the emission calculations is the average of five quarterly surveys. For the years before 2005 and for 2012, figures are based on the amount of wood burned from the annual survey on consumer expenditure. The statistics cover purchase in physical units and estimates for self-harvest. The survey figures refer to quantities *acquired*, which not necessarily correspond to *use*. The survey gathers monthly data that cover the preceding twelve months; the figure used in the emission calculations (taken from the energy accounts), is the average of the survey figures from the year in question and the following year. Combustion takes place in small ovens in private households.

Figures on use of coal and coal coke are derived from information from the main importer. Formerly, Norway's only coal producing company had figures on coal sold for residential heating in Norway. From about 2000, this sale was replaced by imports from abroad. Figures for LPG are collected from the suppliers. Heavy fuel oil is taken from the sales statistics for petroleum products. As the consumption of each energy carrier shall balance against the total sales in the sales statistics, use of fuel oil, kerosene and heavy distillates in households is given as the residual after consumption in all other sectors has been assessed. Use of natural gas is based on sales figures reported to Statistics Norway from the distributors.

Agriculture

Data on energy use in hothouses are collected in surveys performed regularly. Sales figures are

used to project the figures for consumption of oil products in the years between. For biofuels and LPG, figures are interpolated for years not included in surveys. The Agricultural Budgeting Board has figures on the use of gasoline, auto diesel and fuel oil in agriculture excluding hothouses. A figure on the minor use of coal was previously collected annually from the only consumer. Since 2002, however, there has been no known use of coal in the Norwegian agricultural activities. Use of natural gas in agriculture, which has increased considerably since it first was registered in 2003, is based on sales figures reported to Statistics Norway from the distributors.

Fishing

Figures on the use of marine gas fuel, heavy distillate and heavy fuel oil are identical with the registered sales to fishing in the sales statistics for petroleum products. In addition to these figures on use in large fishing vessels, a minor figure on estimated use of gasoline in small fishing boats is also included.

Commercial and institutional sectors

Figures on energy use in wholesale and retail trade, and hotels and restaurants, are based on a survey for 2000, performed by Statistics Norway. For the following years, figures from this survey have been adjusted proportionally to the development in employment in the industries in question. For earlier years, the figures are based on a survey from the mid-1980s (Sagen 1987). LPG figures for the whole period from 1990 have, however, been estimated separately after consultation with an oil company.

For most other commercial and institutional sectors, the total use of fuel oil appears as a residual after the use in all other sectors has been estimated; the distribution of this residual between sub-sectors is done by using figures on energy use per man-labour year from the energy survey from the mid-1980s.

Use of heating kerosene in commercial industries is calculated by projecting a figure on use from the mid-1980s proportionally with the registered sales to buildings in industrial industries outside the manufacturing industries. The estimated total amount is distributed between sub-sectors by using figures on energy use per man-labour year from the mid-1980s survey.

Use of natural gas is based on sales figures reported to Statistics Norway from the distributors.

Calculated emissions from combustion of biogas at a sewage treatment plant are included for all years since 1993.

Military

Figures on fuel oil are annually collected directly from the military administration, while figures from the sales statistics for petroleum products are used for other energy carriers.

3.2.5.3 Emission factor

Emission factors used are given in Appendix B.

Emission factors for fuelwood are based on data for different oven technologies. Ovens made in 1998 and later have significantly improved combustion and reduced emissions. The factors are weighted based on information from the surveys of the amount of wood burned in ovens with the different technologies. The yearly weighted factors are given in appendix B.

The country specific emission factor for PAH-4 is split into benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene based on information from Guidebokk 2013, chapter 1A4.

Table 3.36. Emission factors for fuelwood, g/kg dry matter, BC in share of PM_{2.5} emissions

	Open fireplaces	Ovens -1997	Ovens 1998-
NO _x	1.3	0.97	0.97
CO	126.3	150	50.5
TSP	17.3	22.7	13.4
TSP large cities	17.3	17.4	12.2
PM ₁₀	17.0	22.2	13.1
PM ₁₀ , large cities	17.0	17.1	12.0
PM _{2.5}	16.4	21.6	12.7
PM _{2.5} large cities	16.4	16.5	11.6
BC	9%	0.96%	0.86%
BC large cities	9%	1.01%	0.9%
PCB	0.1184	0.1184	0.0156
Benzo(a)pyrene,	0,82	0,74	0,006
benzo(b)fluoranthene	1,29	1,16	0,01
benzo(k)fluoranthene	0,30	0,27	0,003
indeno(1,2,3-cd)pyrene,	0,59	0,53	0,005

Source: PAH : Finstad et al. (2001), TSP, PM₁₀, PM_{2.5} and BC : SINTEF (2013), PCB : (Nielsen et al. 2015), other pollutants : Haakonsen and Kvingedal (2001)

3.2.5.4 Uncertainties

Uncertainty in *fishing* is described together with navigation in section 3.2.4.5.5.

The method used for finding the use of fuel oil, kerosene and heavy distillates in households implies a great deal of uncertainty regarding the quality of these figures, particularly for fuel oil, which is the most important of these three energy carriers. Since the late 1990s, it has also been necessary to adjust figures for other sectors in order to get consumption figures for households that look reasonable. Hopefully, new surveys will improve the quality of these figures in the future.

As the total use of the different oil products is defined as equal to the registered sales, use in some sectors are given as a residual. This applies to use of heating kerosene and heavy distillates in households, and total use of fuel oil in commercial and institutional sectors. Accordingly, these quantities must be regarded as uncertain, as they are not based on direct calculations. This uncertainty, however, applies only to the distribution of use between sectors - the total use is defined as equal to registered sales, regardless of changes in stock.

There have been large variations in annual sales of military aviation kerosene; as stock changes are not taken into account, the actual annual use is uncertain.

3.2.5.5 Completeness

Major missing emission sources are not likely.

3.2.5.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

3.2.6 International bunkers

NFR - memo item

Last update: 31.03.06

3.2.6.1 Description

Emissions from international bunkers (marine and aviation) have been estimated and reported separately from national estimates, in accordance with the IPCC Guidelines. Differences between the IEA (International Energy Agency) data and the data reported to UNFCCC in sectoral data for marine shipping and aviation, are due to the fact that different definitions of domestic use are employed. In the Norwegian inventory, domestic consumption is based on a census in accordance with the IPCC good practice guidance. On the other hand, the IEA makes its own assessment with respect to the split between the domestic and the international market.

3.2.6.2 Shipping*3.2.6.2.1 Method*

Emissions are calculated by multiplying activity data with emission factors. The sales statistics for petroleum products, which is based on reports from the oil companies to Statistics Norway, has figures on sales for bunkers of marine gas oil, heavy distillates and heavy fuel oil. The same emission factors as in the Norwegian national calculations are used.

3.2.6.2.2 Activity data

Sales figures for international sea transport from Statistics Norway's sales statistics for petroleum products are used for marine gas oil, heavy distillates and heavy fuel oil.

3.2.6.2.3 Emission factor

Emission factors used for *Shipping* are described under *Navigation* in section 3.2.4.5.

3.2.6.3 Aviation*3.2.6.3.1 Method*

The consumption of aviation bunker fuel in Norway is estimated as the difference between total purchases of jet kerosene in Norway for civil aviation and reported domestic consumption. Figures on total aviation fuel consumption are derived from sales data reported to Statistics Norway from the oil companies. These data do not distinguish between national and international uses. Data on domestic fuel purchase and consumption are therefore collected by Statistics Norway from all airline companies operating domestic traffic in Norway. The figures on domestic consumption from airlines are subtracted from the total sales of jet kerosene to arrive

at the total fuel sales for international aviation. The bottom-up approach of Norway is the detailed Tier 2 CORINAIR methodology. The methodology is based on detailed information on types of aircraft and number of LTOs, as well as cruise distances.

3.2.6.3.2 Activity data

Statistics Norway annually collects data on use of fuel from the air traffic companies, including specifications on domestic use and purchases of fuel in Norway and abroad.

3.2.6.3.3 Emission factor

Emission factors used for *Aviation* are described under *Aviation* in section 3.2.4.1.

3.3 Energy production (fugitive emissions from fuels)

NFR 1B

3.3.1 Overview

Emission sources included in the inventory from the sector *Fugitive emissions from fuels* are fugitive emissions from coal mining and handling, and from oil and natural gas.

Fugitive emissions from oil and natural gas include emissions from loading and refining of oil, gasoline distribution, and fugitive emissions from the gas terminals on shore. There are also fugitive emissions in connection with venting and flaring offshore.

3.3.2 Fugitive emissions from coal mining and handling

NFR 1B1

Last update: 12.12.2013

3.3.2.1 Description

Coal has been shipped from Svalbard since 1907. There are today two coal mines at Spitsbergen (the largest island in the Svalbard archipelago) operated by a Norwegian company. They opened the second mine in 2001. As the Norwegian GHG inventory, according to official definitions, shall include emissions from all activities at Svalbard, also emissions from Russian coal production have been estimated. Until 1998, there was production in two Russian coal mines, Barentsburg and Pyramiden, but since then, production takes place only in the Barentsburg mine. The Norwegian mines and Pyramiden are defined as surface mines, whereas Barentsburg is an underground mine. Russian production has in recent years been considerably smaller than the Norwegian production, and due to a fire that started in 2008 the production in 2008 and 2009 was very small. In autumn 2010, ordinary production was restarted. Russian activity data are more uncertain than the Norwegian, which causes a correspondingly higher uncertainty in the emission figures.

At Svalbard there was a smouldering fire in Pyramiden, the Russian mine that was closed down in 1998. At an inspection in 2005, no emissions were registered, which indicates that the fire had burnt out. Due to lack of data, emissions for earlier years from this fire have not been estimated.

However, Norwegian authorities assume that these emissions are limited.

3.3.2.2 Method

NMVOC

NMVOC emissions from handling of coal are estimated by multiplying the amount of coal extracted (raw coal production) with Tier 2 emission factors from EMEP/EEA Guidebook 2013 (EEA 2013).

Particles

Emissions of particles from handling of coal are estimated by multiplying the amount of coal extracted (raw coal production) with Tier 1 emission factors from EMEP/EEA Guidebook 2013 (EEA 2013).

3.3.2.3 Activity data

Figures on Norwegian production (raw coal production) are reported by the plant to Statistics Norway. Russian figures are reported to the Norwegian authorities on Svalbard; these figures are, however, regarded as highly uncertain, consisting of a mixture of figures on production and shipments.

3.3.2.4 Emission factors

NMVOC

Emission factors for NMVOC are taken from EMEP/EEA Guidebook 2013 (EEA 2013). The Tier 2 factors used are 3 kg NMVOC per tonne coal for surface mines and 0.2 kg NMVOC per tonne coal for underground mines.

Particles

Emission factors for particles are taken from EMEP/EEA Guidebook 2013 (EEA 2013). The same Tier 1 factors are used for both surface and underground mines. The factors are 0.089 kg particles per tonne coal for TSP, 0.042 kg particles per tonne coal for PM₁₀ and 0.005 kg particles per tonne coal for PM_{2.5}.

3.3.2.5 Uncertainties

The uncertainty in the activity data concerning Norwegian coal production is regarded as being low. The uncertainty in Russian data is considerably higher.

3.3.3 Fugitive emissions from uncontrolled combustion and burning coal dumps

NFR 1B1b

Last update: 07.06.11

3.3.3.1 Description

In 2005, a fire broke out in one of the Norwegian coal mines at Spitsbergen, causing minor emissions.

3.3.3.2 Method

Emissions have been calculated by multiplication of the quantity of coal combusted by standard emission factors for combustion of coal.

3.3.3.3 Activity data

The company operating the mine has provided an estimate on the quantity of coal combusted in the fire.

3.3.3.4 Emission factors

Emission factors for direct-fired furnaces, as given in Appendix B, have been used in the calculations.

Emissions of BC have been estimated using the same share of PM_{2.5} as used for coal burning.

3.3.3.5 Uncertainties

The uncertainty in the activity data, that is the quantity of coal combusted, is unknown. However, as the emissions are small, the uncertainty is insignificant.

3.3.3.6 Completeness

The only fire in a Norwegian coal mine since 1990 is included. Emissions from a smouldering fire in a Russian mine, which is supposed to have lasted for several years, are not included in the emission inventory, due to lack of data. The same applies to another fire in 2008. These emissions are, however, probably insignificant.

3.3.3.7 Source specific QA/QC

There is no specific QA/QC procedure for this source.

3.3.4 Oil and natural gas

NFR 1B2

Last update: 10.03.2015

3.3.4.1 Description

1B2a covers emissions from loading and storage of crude oil, refining of oil and distribution of gasoline. Loading, unloading and storage of crude oil on the oil fields off shore and at oil terminals on shore cause emissions of NMVOC. Non-combustion emissions from Norway's two oil refineries (a third was closed down in 2000) include NO_x, NMVOC, SO₂ and particulates. Gasoline distribution causes emissions of NMVOC. Especially from 2007 there has been a shift in the consumption of fuels for road traffic from gasoline to auto diesel. This is mainly due to the introduction of a CO₂ tax on passenger cars (PC) from January 1st 2007. This resulted in diesel driven cars becoming less expensive than gasoline driven cars. From 2007, approximately 70 per cent of all new PCs were diesel driven. This has led to a reduction in the distributed amount of gasoline, and consequently of the NMVOC emissions from this source.

1B2b covers fugitive emissions of NMVOC from gas terminals on shore.

1B2c covers fugitive emissions from venting and flaring. Venting emissions include emissions of NMVOC from exploration and production drilling of gas and oil. The major source is cold vent and leakage of NMVOC from production drilling.

Most of the emissions in 1B2c come from flaring of natural gas offshore (during both well testing, extraction and pipeline transport) and at gas terminals and flaring of refinery gas at the refineries. This flaring causes emissions of NO_x, NMVOC, SO₂, CO, particulates, BC, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins. There is also some flaring of oil in connection with well testing - amounts flared and emissions are reported to NPD (the Norwegian Petroleum Directorate) and the Norwegian Environment Agency.

The major source in sector 1B2 is flaring of natural gas on the Norwegian continental shelf. Table 3.37 gives an overview over the calculations of the fugitive emissions of NMVOC.

Table 3.37 Fugitive emissions from oil and natural gas. Emission sources, compounds, methods, emission factors and activity data included in the Norwegian GHG Inventory

B Fugitive emissions from fuels	NMVOC	Method	Emission factor	Activity data
1.B.2.a Oil				
i. Exploration	IE	Tier II	CS	PS
ii. Production	IE	Tier II	CS	PS
iii. Transport	R/E	Tier II	CS	PS
iv. Refining/Storage	R	Tier I/II	CS	PS
v. Distribution of oil products	R/E	Tier I	C/CS	CS/PS
vi. Other	NO			
1.B.2.b Natural gas				
i. Exploration	IE	IE	IE	IE
ii. Production/Processing	IE	IE	IE	IE
iii. Transmission	IE	IE	IE	IE
iv. Distribution	IE	Tier II	OTH	CS/PS
v. Other leakage				
industrial plants, power stations	R	Tier II	CS	PS
residential/commercial sectors	NO			
1.B.2.c Venting				
i. Oil	IE	Tier II	CS/PS	PS
ii. Gas	IE	Tier II	CS/PS	PS
iii. Combined	R/E	Tier II	CS/PS	PS
Flaring				
i. Oil (well testing)	R/E	Tier II	CS	PS
ii. Gas				
Gas and oil fields	R/E	Tier II	CS	PS
Gas terminals	R/E	Tier I	CS	CS
Refineries	E	Tier I	CS	CS
iii. Combined	IE	Tier I	CS	CS

R = emission figures in the national emission inventory are based on figures reported by the plants. E = emission figures are estimated by Statistics Norway (Activity data * emission factor). IE = Included elsewhere, NO = Not occurring, CS = Country specific, PS = Plant specific, Tier = the qualitative level of the methodology used, C=Corinair, OTH=Other.

Source: Statistics Norway/Norwegian Environment Agency

3.3.4.2 Method

Loading and storage of crude oil off shore and on shore

NMVOC

From 2003, emission of NMVOC from loading and storage of crude oil on shuttle tankers included in the GHG inventory are based on reported emission figures from the oil companies. Emissions, activity, and to some extent emission factors, are reported from each field operator into the database *Environmental Web*/EPIM Environment Hub (EEH). The database is operated by Norwegian Oil and Gas Association. In addition the field operators each year deliver a report where they describe the activities during the last year.

Before 2003, the emissions of NMVOC were calculated by Statistics Norway. The calculation was based on the field specific amounts of crude oil loaded and stored multiplied with field specific emission factors. Field specific activity data and emission factors (the latter only to the Norwegian Environment Agency) used in the calculation were annually reported by the field operators to Statistics Norway and the Norwegian Environment Agency. Since 2000, an increasing share of the shuttle tankers have had installed vapour recovery units (VRU), and emissions from loading of crude oil on shuttle tankers with and without VRU were calculated separately for each field. In addition, emission figures were annually reported to the Norwegian Environment Agency and used in the QC of the calculated emission figures.

Only emissions from loading and storage of the Norwegian part of oil production are included in the inventory. For the Norwegian oil terminals on shore, the emissions from loading of crude oil are reported annually from the terminals to the Norwegian Environment Agency. At one of the terminals, VRU for recovering NMVOC was installed in 1996. The calculation of the emissions of NMVOC at the terminals is based upon the amount of crude oil loaded and oil specific emission factors dependent on the origin of the crude oil loaded.

Oil refineries

NO_x, NMVOC, SO₂ and particulates

Emission figures from the oil refineries are reported to the Norwegian Environment Agency, and are after QA/QC procedures used in the emission inventory.

NMVOC emissions are based on measurements carried out by Spectracyne in 2002 and 2005.

Gasoline distribution

NMVOC

Emissions from gasoline distribution are calculated from figures on amounts of gasoline sold and emission factors for, respectively, loading of tanker at gasoline depot, loading of tanks at gasoline stations, and loading of cars.

Gas terminals

NMVOC

Fugitive emissions of NMVOC from gas terminals are annually reported from the terminals to the Norwegian Environment Agency.

The emissions are calculated based on the number of sealed and leaky equipment units that is recorded through the measuring and maintenance program for reducing the leakage. The number of sealed and leaky equipment units is collected two times a year and the average number of the counting is used in the calculation. It is assumed in the calculation that a leakage has lasted the whole year if not the opposite is documented.

Measurements of the total emissions were carried out in 2002 and 2003.

Venting

NMVOC

Emissions of NMVOC from cold venting and diffuse emissions for each field are reported annually to the Norwegian Environment Agency from the field operator. The emissions are mostly calculated by multiplying the amount of gas produced with an emission factor for each emission source identified at the field.

Flaring

NO_x, NMVOC, CO, particulates, BC, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins

Emissions from flaring of natural gas off shore are calculated by Statistics Norway on the basis of field specific gas consumption data and emission factors. For NO_x, NMVOC and SO₂, calculated emissions are used in the inventory for the years until 2002. From 2003, emissions of these pollutants from flaring offshore have been reported by the oil companies to NPD and the Norwegian Environment Agency are used in the inventory. The same method is used in the calculation of emissions from flaring in connection with well testing.

Emissions of NO_x from flaring at gas terminals are reported for all years. For NMVOC, emissions are calculated for one gas terminal and reported figures used for the others. Other emissions from the gas terminals are based on activity data and emission factors.

3.3.4.3 Activity data

Loading and storage of crude oil off shore and on shore

The amount of oil buoy loaded and oil loaded from storage tankers is reported by the field operators in an annual report to the Norwegian Environment Agency and the Norwegian Petroleum Directorate (NPD). The amount of oil loaded on shuttle tankers with or without VRU is separated in the report.

Before 2003, Statistics Norway gathered data on amounts of crude oil loaded at shuttle tankers and stored at storage vessels from the NPD. The data from each field are reported monthly by the field operators to NPD on both a mass and a volume basis. The allocation of the amount of crude oil loaded at shuttle tankers and stored at storage vessels with or without VRU is from the annual report the field operators are committed to deliver to the Norwegian Environment Agency and NPD.

The amount of oil loaded at on shore oil terminals is also reported to the Norwegian Environment Agency and NPD.

Oil refineries

The crude oil throughput is annually reported by the plant to the Norwegian Environment Agency.

Gasoline distribution

Gasoline sold is annually collected in Statistics Norway's sales statistics for petroleum products.

Gas terminals

Activity data that the terminals use in their emission calculations are sampled through the terminals measuring and maintenance program, whose aim is to reduce leakage.

Venting

Amounts of gas produced or handled at the platforms are reported from NPD and used in the QC of the reported emissions.

Flaring

Amounts of gas flared at offshore oil and gas installations are reported monthly by the operators to the Norwegian Petroleum Directorate (NPD). Amounts flared at the gas terminals are reported to NPD and the Norwegian Environment Agency. Amounts of refinery gas flared are found by distributing the total amounts between different combustion technologies by using an old distribution key, based on data collected from the refineries in the early 1990s. This distribution was confirmed in 2003.

3.3.4.4 Emission factors

Loading and storage of crude oil offshore and on shore

For the years before 2003, emission factors used in the calculation of NMVOC emissions offshore are field specific and were reported to the Norwegian Environment Agency and NPD in an annual report. The Norwegian Environment Agency forwarded the emission factors to Statistics Norway. From 2003, the emission figures reported by the field operators are used in the inventory.

The evaporation rate varies from field to field and over time, and the emission factors are dependent on the composition of the crude oil as indicated by density and Reid vapour pressure (RVP). The VOC evaporation emission factors are obtained from measurements, which include emissions from loading and washing of shuttle tankers. For some fields the emission factors are not measured, only estimated. The CH₄ content of the VOC evaporated is also measured so that total emissions of VOC are split between CH₄ and NMVOC.

The emission factors that the field operators use in their calculations are reported to the Norwegian Environment Agency and NPD. They report emission factors with and without VRU and the split between CH₄ and NMVOC.

Loading on shore: The emission factors are considerably lower at one of Norway's two oil terminals than at the other, because the oil is transported by ship and therefore the lightest

fractions have already evaporated. At the other terminal the oil is delivered by pipeline. The latter terminal has installed VRU, which may reduce NMVOC emissions from loading of ships at the terminal by about 90 per cent. NMVOC emissions at this terminal are estimated to be more than 50 per cent lower than they would have been without VRU. However, the VRU technology is not designed to reduce methane and ethane emissions.

Oil refineries

The emission factor used in the calculation of methane emissions from the largest refinery is based upon measurements performed by Spectracyne in 2002 and 2005. The EF is deduced from the measured methane emissions and the crude oil throughput in 2005.

BC emissions have been estimated as a fraction of the PM_{2.5}. IIASA (Kupiainen & Klimont 2004) gives a fraction of 0.16 per cent.

Gasoline distribution

The emission factor for NMVOC from refuelling of gasoline in cars (1.48 kg NMVOC/tonne gasoline) is taken from inventory guidebook (EEA 2001).

Venting

The emission factors used are listed in Table 3.38.

Table 3.38. Emission factors for cold vents and leakage at gas fields offshore

	NMVOC Emission factor	Calculation method
Emission source	[g/Sm ³]	
Glycol regeneration	0.065	
Gas dissolved in liquid from K.O. Drum	0.004	
Gas from produced water system	0.03	
Seal oil systems	0.015	
Leaks through dry compressor gaskets	0.0014	
Start gas for turbines ¹	0.4	Tonne per start up
Depressurisation of equipment	0.005	
Instrument flushing and sampling	0.00021	
Purge and blanket gas ¹	0.032	
Extinguished flare	0.014	
Leaks in process	0.007	
Depressurisation of annulus	0.0000005	
Drilling	0.55	Tonne per well

¹ The gas source is standard fuel gas.

Source: Aker Engineering (1992)

Flaring

NO_x: A NO_x emission factor at 1.4 g NO_x/Sm³ flared gas at off shore installations is based upon studies conducted by Bakken et al. (2008). In the study two new experimental laws have been compared with DIAL-measurements of NO_x emissions made on onshore flares.

PM₁₀: The emission factor is based on McEwen and Johnson (2011). In fig. 7, this paper gives a regression formula for the emission factor as a function of the heating value (GCV) as EF =

0.0578(HV) – 2,09. For Norwegian offshore flaring a heating value of 48 MJ/Sm³ is suggested in Bakken et al. (2008). This gives an emission factor of 0.856 g PM₁₀/Sm³.

BC: Emissions are estimated using the same methodology as PM₁₀ emissions. The regression formula for the BC emission factor, given as a function of the heating value (GCV) is EF = 0.0578(HV) – 2,09. This gives an emission factor of 0.684 g BC/Sm³.

Other emission factors from flaring of gas are shown in appendix B. The same factors are used for flaring of gas in connection with *well testing*. For flaring of *oil*, the emission factors are shown in Table 3.39.

Table 3.39. Emission factors for flaring of oil in connection with well testing

Compounds (unit)	unit/tonnes flared oil	Source
NO _x (tonnes)	0.0037	(OLF 2009)
NM VOC (tonnes)	0.0033	
CO (tonnes)	0.018	
TSP (tonnes)	0.025	Measurements (OLF ¹)
PM ₁₀ (tonnes)	0.0215	Use the same distribution as for combustion of heavy fuel oil in industry (EPA 2002)
PM _{2.5} (tonnes)	0.014	
PAH (kg)	0.012	(OLF 1991)
Benzo(a)pyrene,	0.00024	Use the same distribution as for combustion of heavy fuel oil in industry (EPA 1998)
benzo(b)fluoranthene		
benzo(k)fluoranthene		
indeno(1,2,3-cd)pyrene,		
Dioxins (mg)	0.01	Measurements (OLF)
PCB (mg)	220	Langøren and Malvik (2010)

¹The Norwegian Oil Industry Association (OLF) now Norwegian Oil and Gas Association.

3.3.4.5 Uncertainties

The uncertainty in the emission factors for NM VOC (Rypdal & Zhang 2001) from *oil loading* is estimated to be ± 40 per cent and in the activity data ± 3 per cent.

The uncertainty in the amount of gas flared is regarded as being low, ±1.4 per cent, based on data reported in the emission trading scheme (Climate and Pollution Agency 2011) and assumptions in Rypdal and Zhang (2000). The uncertainty in NM VOC emissions from venting is much higher than for flaring.

The emission factors for both storage and transmission of natural gas are uncertain, since Austrian factors are used in lack of country specific Norwegian factors.

All uncertainty estimates for this source are given in Appendix C.

3.3.4.6 Source-specific QA/QC and verification

Statistics Norway gathers activity data on oil and gas activities from the Norwegian Petroleum Directorate (NPD). These data are reported monthly by the field operators to NPD. The activity data are quality controlled by comparing them with the figures reported in the field operator's annual report to the Norwegian Environment Agency and NPD. The emissions calculated by Statistics Norway for 1990-2002 were compared with the emission data that the field operators reported to the Norwegian Environment Agency and NPD. From 2003, Statistics Norway estimate emissions based on activity data that the field operators monthly report to NPD, and reported emission factors. When discrepancies are found between the two sets of data these are investigated and corrections are made if appropriate. If errors are found, the Norwegian Environment Agency contacts the plant to discuss the reported data and changes are made if necessary.

The reported emissions from the gas terminals are compared with previous years' emissions.

The Norwegian Environment Agency collects the activity data used for venting and flaring in the calculation from the NPD. The figures are quality controlled by comparing them with the figures reported in the field operators' annual report to the Norwegian Environment Agency and NPD, and time series are checked.

The Norwegian Environment Agency perform internal checks of the reported data for venting from the field operators. Some errors in the time-series are usually found and the field operators are contacted and changes are made. The same procedure is followed to check the amount of gas reported as flared. The quality of the activity data is considered to be high, due to the fact that there is a tax on gas flared offshore. NPD has a thorough control of the amount of gas reported as flared.

4 INDUSTRIAL PROCESSES AND PRODUCT USE (NFR sector 2)

NFR 2

4.1 Overview

This chapter provides descriptions of the methodologies employed to calculate emissions of greenhouse gases and long-range transboundary air pollutants from industrial processes and product use. Only non-combustion emissions are included in this chapter. Emissions from fuel combustion in the manufacturing industries are reported in the Energy chapter. Emission figures are either reported by plants to the Norwegian Environment Agency or calculated by Statistics Norway, based on emission factors and activity data. The emission factors are collected from different sources, while the activity data used in calculations carried out by Statistics Norway mainly come from official statistics collected by Statistics Norway.

4.2 Mineral products

NFR 2A

Last update: 22.10.16

The sector category Mineral products in the Norwegian inventory includes emissions from fourteen different products. SO₂, NO_x, NH₃, particles, BC, heavy metals dioxins, PAHs benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and HCB are components that are emitted during the production of mineral products and included in the inventory.

Table 4.1 shows the various components emitted from the different activities, and for which components the emission figures in the national inventory are based on figures reported by the plants (R) and for which the figures are estimated by Statistics Norway (E).

Table 4.1. Mineral products. Components emitted and included in the Norwegian inventory¹

Mineral products	SO ₂	NO _x	NH ₃	Particles	BC	Heavy metals	Dioxins	PAH	HCB
Cement production	R	NE	NE	R	E	R	R	NE	R
Lime production	NE	NE	NA	R	E	R	NA	NA	R
Glass and glassfibre production	NE	R	R	R	E	R	NA	NE	NE
Mining and extraction of stones and minerals	NA	NA	NA	R	NA	NA	NA	NA	NA
Construction and demolition	NA	NA	NA	E	NA	NA	NA	NA	NA
Ceramics	NA	NA	NA	R	E	NA	NA	NA	NE
Non-metallurgical magnesia production	R	NA	NA	R	E	NA	R	NA	NE
Sandpit and rock-crushing plants	NA	NA	NA	E	NA	NA	NA	NA	NA
Concrete pumice stone	R	NA	NA	R	E	NA	NA	NA	NA
Rock wool production	NA	R	R	R	E	R	NA	NA	NE
Production of mineral white	NA	NA	NA	R	E	R	NA	NA	NA
Construction /repairing of vessels - Sandblasting	NA	NA	NA	R	NA	NA	NA	NA	NA
Leather preparing	NA	NA	R	NA	NA	NA	NA	NA	NA
Production of asphalt	NA	NA	NA	E	E	NA	E	E	NE

¹ R means that emission figures in the national emission inventory are based on figures reported by the plants. E means that the figures are estimated by Statistics Norway (Activity data * emission factor). NA = Not Applicable.

Source: Statistics Norway/Norwegian Environment Agency

4.2.1 Cement production

NFR 2A1

Last update: 05.02.16

4.2.1.1 Description

Two plants in Norway produce cement. Production of cement gives rise to both non-combustion and combustion emissions of SO₂. The emission from combustion is reported in chapter 3 Energy. The non-combustion emissions originate from the raw material calcium carbonate (CaCO₃). The resulting calcium oxide (CaO) is heated to form clinker and then crushed to form cement. The emissions of SO₂ from non-combustion are reported to The Norwegian Environment Agency.

SO₂ from cement production is emitted from sulphur in the fuel (reported under Energy) and in the raw materials, especially pyrite in limestone. Only the SO₂ from the raw materials should be counted as non-combustion emissions. Particles as well as heavy metals are emitted during the production process. More than 90 per cent of the emission of mercury is due to mercury in the limestone, while the emissions of Pb, Cd, Cu, Cr and As originate both from processes and combustion of fuel. Emissions of dioxins are due to the thermal process in the clinker production.

4.2.1.2 Method

SO₂

The plants annually report emissions of SO₂ to the Norwegian Environment Agency. Figures are based on measurements at the plants.

SO₂ emissions from production of cement come from energy carriers like e.g. coal and oil and

from limestone. The sulphur from the energy carriers is to a large extent included in the clinker during the process. The emissions are distributed between combustion and non-combustion emissions based on studies conducted by Institute for Energy Technology in 1970 and 1999. Both studies indicate that 80-99 per cent of the sulphur from energy carriers is included in the clinker.

The total SO₂ emissions from the two plants are based on measurements. When the SO₂ emissions reported from the plant are not distributed between combustion and non-combustion emissions, the Norwegian Environment Agency distributes the total emissions, using the same percentage distribution as in the last year with reported distributed SO₂ emissions. The production technology is to some extent different for the two plants. In the last years, the distribution between combustion and non-combustion emissions is about 10/90 for one plant and 18/82 for the other plant. The difference is assumed to be due to the fact that one plant has a "by-pass" system where some of the flue gas is not in contact with the raw materials.

The amount of energy carriers used in cement production is subtracted from the energy balance to avoid double counting, see section 3.2.1.2.

Particles

Emissions have been reported to the Norwegian Environment Agency since 1991 for one plant and since 1992 for the other. It is believed that the reported figures also include emissions from combustion. Therefore, emissions from combustion of coal, coke and waste oil used in cement production are not calculated, to avoid double counting. The plants have installed particle filter.

Particle size distribution for emitted particles from cement production is found in TNO (Institute of environmental and energy technology 2002). In the Norwegian emission inventory, PM₁₀ and PM_{2.5} are assumed to be 85 and 30 per cent of TSP, respectively.

BC

Emissions have been estimated from a share of PM₁₀ emissions given by IIASA in Kupiainen and Klimont (2004). As a share of PM₁₀, BC emission factor is 0.25 per cent.

Heavy metals and POPs

Emission figures for heavy metals are reported to the Norwegian Environment Agency. It is believed that these figures also include emissions from combustion. Therefore emissions from combustion of coal, coke and waste oil used in cement production are not calculated, to avoid double counting.

Dioxin figures are reported to the Norwegian Environment Agency. It is also here assumed that the reported figures include emissions from fuel combustion, therefore emissions from combustion are not calculated.

HCB emissions were reported by the plant to the Norwegian Environment Agency in 2010. This reporting has been used to estimate emissions in 2010. For other years in the time series,

emissions have been estimated using a Tier 2 emission factor from the EEA (2016).

4.2.1.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Reported emission figures for particles have varied a great deal as a result of changes the plants have undergone to reduce emissions. There are also uncertain measurements due to annual variations.

Regarding the heavy metals, it has varied when the two plants started reporting the various components, and therefore estimations have been necessary for the years when reporting have been insufficient. The reported figures also vary from a year to another due to process technical conditions, variations in the metal content in the limestone used, and uncertain measurements.

4.2.1.4 Completeness

Major missing emission components are not likely.

4.2.1.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure

4.2.2 Lime production

NFR 2A2

Last update: 05.02.16

4.2.2.1 Description

Emissions of particles, black carbon and HCB from lime production are included in the Norwegian inventory.

4.2.2.2 Method

Particles

One plant has reported emission figures for particulate matter to the Norwegian Environment Agency since 1990. Emission figures from 1990 to 1995 are based on calculations, using emission factors and production volume. Since 1996, the figures are a result of measurements at the plant. The plant has installed particle filter.

In the inventory, a particle size distribution suggested by TNO (Institute of environmental and energy technology 2002) is used. PM_{10} is $0.4 \cdot TSP$ while $PM_{2.5}$ is $0.08 \cdot TSP$.

BC

For the same plant that reports particles, BC emissions have been estimated from a share of TSP emissions given by IIASA in Kupiainen and Klimont (2004). As a share of TSP, BC emission factor is 0.2 per cent.

HCB

HCB may unintentionally be formed in the production and extracation of lime in the thermic process. One plant has reported emisisions in 2010. Emissions for the rest of the timeseries are estimated based on lime production data. Emissions for two other plants that do not report emisisions are also estimated based on lime production. The emission factor used is 0.008 mg HCB per tonne lime from Japan (Toda 2006). It is also assumed that, for this category, the reported figures include emissions from fuel combustion, therefore emissions from combustion are not calculated.

4.2.2.3 Activity data

The activity data is the input of limestone and dolomite, these amounts are annually reported to the Norwegian Environment Agency.

4.2.2.4 Uncertainties

The particle distribution used is not specified for the plants, and the particles emitted might therefore have another distribution than the one suggested from TNO (Institute of environmental and energy technology 2002).

4.2.2.5 Completeness

Major missing emission components are not likely.

4.2.2.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.3 Glass and glassfibre production

NFR 2A3

Last update: 22.12.16

4.2.3.1 Description

Five plants producing glass, glass wool or glass fibre are included in the emission inventory, with figures based on emission reports to the Norwegian Environment Agency. PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxin emissions are neither calculated nor measured although glass production might be a dioxin emissions source.

4.2.3.2 Method

NO_x

The two glass wool producing plants and the one producing glass fibre annually report emission figures for NO_x to the Norwegian Environment Agency. The emission figures are based on calculations.

NH₃

The two glass wool producing plants annually report emission figures for NH₃ to the Norwegian Environment Agency. The emission figures are based on measurements.

Particles

The two plants producing glass wool have reported emission figures to the Norwegian Environment Agency since 1990. The glass fibre producing plant has reported emissions from 1996; for the period 1990-1995, the 1996 figure is used in the inventory. One glass-producer with particle emissions reported figures from 1995. Emission figures from 1990 to 1994 were assumed to be the same as the reported 1995 figure. This plant was closed down in 1999.

TNO (Institute of environmental and energy technology 2002) suggests using a particle size distribution of the emissions where $PM_{2.5}$ is 80 per cent of TSP and PM_{10} is 90 per cent of TSP, and this size distribution is used in the Norwegian inventory.

BC

Emissions have been estimated from a share of $PM_{2.5}$ emissions given by IIASA in Kupiainen and Klimont (2004). BC has been considered to be 0.06 per cent of $PM_{2.5}$ emissions.

Heavy metals and POPs

Emission of lead has been reported from two glass-producers to the Norwegian Environment Agency. One of them was closed down in 1999. The emission of lead is due to the lead content in the raw material used. Emissions of other heavy metals are reported under Energy.

4.2.3.3 Uncertainties

For the years where reported emission figures for particles do not exist, Statistics Norway has assumed that emissions are in the same order of magnitude as for the first year of reporting. This is uncertain and only an estimate, since it does not consider annual changes in raw materials, production rates, nor possible cleaning devices.

4.2.3.4 Completeness

Production of glass can be a source for dioxin emissions, but no reported figures are available. Emission factors are found in literature, but since activity data (production rate) is not available and it is assumed that the emission factor is dependent on type of glass produced, emissions are not calculated.

Emissions of particles are also reported from three other glass-producers in Norway, but since annual emissions are low (less than 1 tonne), they are not included in the inventory.

4.2.3.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.4 Mining and extraction of stones and minerals

NFR 2A5A

Last update: 14.09.15

4.2.4.1 Description

Mining and extraction of stones and minerals are done by several plants. Particles are emitted

during these processes.

4.2.4.2 Method

Particles

Emission figures are reported to the Norwegian Environment Agency. Reported figures exist from 1992. Emission figures for 1990 and 1991 are assumed by Statistics Norway, in accordance with the Norwegian Environment Agency, to be the same as reported figures in 1992. An exception is one plant, which only reported emissions for 1992. For this plant, Statistics Norway has calculated emissions based on production rates for previous and later years.

It is given for most plants that they use fabric filter or textile fibre to clean their particle emissions. It is assumed by the Norwegian Environment Agency that the particles emitted are larger than PM₁₀. Thus, BC has not been considered for this category.

The Norwegian inventory uses the size distribution recommended by TNO (Institute of environmental and energy technology 2002) for sandpits and rock-crushing plants (Table 4.3).

4.2.4.3 Uncertainties

For years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order of size as for the first year of reporting. This is uncertain and a result of lack of better data. The size of the particles emitted from mining and extraction will also depend on the type of stone/mineral and production process. The particle size distribution used in the inventory does not consider these differences.

4.2.4.4 Completeness

Emission of particles is often a source of heavy metal emissions since particles often contain heavy metals. Type of metals will however depend on the origin of the particles. Metals might therefore be emitted during mining and extraction of stones and minerals. There are, however, no data available for calculating emissions of heavy metals.

4.2.4.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.5 Construction and demolition

NFR 2A5B

Last update: 10.01.17

4.2.5.1 Description

Construction and building includes a lot of different activities that will generate particle emissions.

4.2.5.2 Method

Particles

Emission factors and activity data are used to estimate the Norwegian emissions.

4.2.5.3 Emission factors

The emission factors used are shown in Table 4.2.

Table 4.2. Particle emission factors for building and construction. Tonne/hectare/year

Component	Tonne/hectare/year
TSP	1.62
PM ₁₀	0.8121.52
PM _{2.5}	0.08120.52

Source: EEA (2013)

Statistics Norway assumes that none of the processes used in building and construction will lead to BC emissions. Hence, BC has not been considered for this activity.

4.2.5.4 Activity data

The activity data used is the annual area of completed buildings from the building statistics at Statistics Norway.

4.2.5.5 Uncertainties

The particle emissions depend on climate conditions as well as building traditions and building materials. Since the emission factors used are based on surveys in other countries than Norway, these factors might not be ideal for Norwegian conditions.

4.2.5.6 Completeness

Building of roads, railways, tunnels and demolition of buildings is also a source of particle emissions, but no emission factors are found in the literature, and therefore such emissions are not included in the inventory.

4.2.5.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.6 Ceramics

NFR 2A6

Last update: 27.10.16

4.2.6.1 Description

One plant producing bricks reports emission of particulate matter from limestone and dolomite use to the Norwegian Environment Agency.

4.2.6.2 Method

Particles

Emissions have been reported to the Norwegian Environment Agency since 2000. Reported figure for 2000 have been used for all years since 1990. The same particle size distribution is used as for production of cement as given in TNO (Institute of environmental and energy technology 2002). PM₁₀ and PM_{2.5} are assumed to be 85 and 30 per cent of TSP, respectively.

BC

Emissions has been estimated from a share of PM_{2.5} emissions. Values for bricks production are given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 37.5 per cent.

4.2.6.3 Uncertainties

Uncertainty estimates are given in Appendix C.

4.2.6.4 Completeness

Major missing emission components are not likely.

4.2.6.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.7 Non-metallurgical Magnesia Production

NFR 2A6

Last update: 27.10.16

4.2.7.1 Description

One plant whose main activity is producing magnesium oxide and calcium oxide from limestone and dolomite is included in the emission inventory. The plant was established in 2005.

4.2.7.2 Method

Particles

Emissions have been reported to the Norwegian Environment Agency for the years 2005-2008 and 2013 and onwards. Linear interpolation has been used for the intervening years. No information is found regarding the particle size distribution for particles emitted during production. In lack of other data we used the same distribution as for aluminium production. PM₁₀ and PM_{2.5} are assumed to be 100 and 43 per cent of TSP, respectively.

BC

Emissions has been estimated from a share of PM_{2.5} emissions. Values for bricks production are given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 37.5 per cent.

SO₂

Emissions have been reported to the Norwegian Environment Agency since 2006.

Dioxins

Emissions have been reported to the Norwegian Environment Agency for the years 2011, 2013 and onwards.

PCB

Emissions have been measured and reported to the Norwegian Environment Agency for 2010. An emission factor has been built from these emissions measurements to estimate emissions for

the whole timeseries.

4.2.7.3 Activity data

The amount of limestone and dolomite used by the plant in their calculation is annually reported to the Norwegian Environment Agency.

4.2.7.4 Uncertainties

Uncertainty estimates are given in Appendix C.

The particle distribution used is not specified for the plants, and the particles emitted might therefore have another distribution than the one suggested.

4.2.7.5 Completeness

Major missing emission components are not likely.

4.2.7.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.8 Sandpit and rock-crushing plant

NFR 2A6

Last update: 10.01.17

4.2.8.1 Method

Particles will be emitted during crushing of rocks and at sandpits. In the inventory, emissions are estimated based on the production of sand and crushed stone from the production statistics at Statistics Norway, and emission factors recommended by EEA (2016).

4.2.8.2 Activity data

The production of sand and crushed stone is annually given by Statistics Norway's production statistics.

4.2.8.3 Emission factors

The emission factors used are shown in Table 4.3.

Table 4.3. Particle emission factors for sandpits and rock-crushing plants. Ratio X^1 /TSP

Component	g/tonne produced
TSP	102
PM ₁₀	50
PM _{2.5}	5

¹ X is either PM_{2.5}, PM₁₀ or TSP.

Source: EEA (2016)

All particles are assumed to be larger than PM_{2.5}. Thus, no emission of BC has been estimated.

4.2.8.4 Uncertainties

This emission source is highly uncertain since the emissions will vary from one place to another depending on the different processes in use, type of raw materials and of course the activity

level. Little information is available in the literature. The emission factors used are only based on one source and are uncertain. In addition, there is uncertainty regarding the activity data. The PRODCOM codes used in the production statistics include total production of sand and crushed stone in Norway, but some of it might not be relevant for these calculations.

4.2.8.5 Completeness

Emission of particles is often a source of heavy metal emissions since particles often contain heavy metals. Type of metals will however depend on the origin of the particles. Metals might therefore be emitted during crushing at sandpits and rock-crushing plants. There are however no data available for calculating emission of heavy metals.

4.2.8.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.9 Concrete pumice stone

NFR 2A6

Last update: 27.10.16

4.2.9.1 Description

Two factories have reported emissions of SO₂ and particles from concrete pumice stone production to the Norwegian Environment Agency until 2004 when one of them was closed down.. Non-combustion emissions of SO₂ originate from the clay used in the production process.

4.2.9.2 Method

SO₂

Emission figures for SO₂ are reported to the Norwegian Environment Agency, based on measurements at the two manufacturing plants in Norway. The plants have installed flue gas desulphurisation equipment.

Particles

The plants have reported emissions of particles to the Norwegian Environment Agency since 1990. It is assumed that the reported figures include both process and combustion emissions, so emission calculations from fuel combustion are not done for these two plants. The plants have installed particle filters.

No information concerning particle size is found in national or international literature, but the Norwegian Environment Agency assumes that most of the particles emitted from these plants are smaller than PM₁₀. Statistics Norway has decided to use the same particle size distribution for production of cement as given in TNO (Institute of environmental and energy technology 2002). PM₁₀ is therefore assumed to be 0.85*TSP and PM_{2.5} is 0.3*TSP.

BC

Emissions has been estimated from a share of PM_{2.5} emissions. Values for bricks production are

given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 37.5 per cent.

4.2.9.3 Uncertainties

The particle size distribution used is not specific for production of concrete pumice stone, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate.

4.2.9.4 Completeness

Particles often contain heavy metals, but type of metals and volumes will depend on the origin of the particles. Metals might therefore be emitted during production of concrete pumice stone. Statistics Norway and the Norwegian Environment Agency have, however, no data available for calculating emissions of heavy metals from this source.

4.2.9.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.10 Rock wool production

NFR 2A6

Last update: 12.01.17

4.2.10.1 Description

Three plants in Norway produced rock wool until 2003 when one of them was closed down. In the inventory, emission figures for NH₃, particles and heavy metals are included. Particles originate from the cutting of the mineral wool and from fuel used in the production. The emissions of heavy metals are partly due to use of coal/coke, but mainly due to the stone used in the production. Emissions of dioxins and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) are neither reported nor calculated since emissions of these components are minor or not occurring.

4.2.10.2 Method

NO_x

Emission figures are reported to the Norwegian Environment Agency.

NH₃

Emission figures are reported to the Norwegian Environment Agency. Figures exist from 1992. It is assumed in the inventory that emission figures for 1990 and 1991 are the same as the reported figure in 1992.

Particles

Emission figures are reported to the Norwegian Environment Agency. Most of the emissions come from the spin chamber, and the particle size is assumed to be less than 1 µm. Particles emitted from the fabric filter are also assumed to be smaller than 1 µm. All emissions are therefore set to be smaller than PM_{2.5}. All assumptions are made by the Norwegian Environment

Agency in accordance with the industry. It is assumed that the reported figures include both non-combustion and combustion emissions. Combustion emissions of particles are therefore not calculated.

BC

Emissions have been estimated from a share of PM_{2.5} emissions. Values for glass fiber production are given by IIASA in Kupiainen and Klimont (2004). As a share of PM_{2.5}, emission factor is 0.06 per cent.

Heavy metals and POPs

Emission figures for Pb, Cd, As and Cr have been reported annually from one of the plants to the Norwegian Environment Agency since 1999. The figures are based on measurements. It is assumed that the reported figures include combustion emissions, and emission calculations from fuel combustion are not done for these heavy metals. Statistics Norway has calculated the emission figures for missing years (1990-1998) based on reported figures in 1999 and production rates for previous years. For the two plants not reporting, Statistics Norway calculates emissions based on derived emission factors from the one plant that reports and production volumes at each plant.

4.2.10.3 Activity data

Production volumes of rock wool are annually reported from the plants to the Climate and Pollution Agency.

4.2.10.4 Emission factors

BC

Emissions have been estimated from a share of PM_{2.5} emissions. Value given by IIASA in Kupiainen and Klimont (2004) for glass production have been used. As a share of PM₁₀, emission factor is 0.25 per cent.

Heavy metals

A default emission factor is derived for each component (Pb, Cd, As and Cr) based on the annually reported emission figures and production rates from the one plant reporting. The derived emission factors are used to calculate emissions from the two other plants (one of these were closed down in 2003).

*Table 4.4. Emission factors for Pb, Cd, As and Cr from production of rock wool.
g/tonne produced rock wool*

Component	Emission factors (g/tonne produced rock wool)
Lead (Pb)	0.164
Cadmium (Cd)	0.001
Arsenic (As)	0.031
Chromium (Cr)	0.703

Source: Statistics Norway/Norwegian Environment Agency

4.2.10.5 Uncertainties

Activity data

The activity data is assumed to be of good quality since this is production rates reported from each plant to the Climate and Pollution Agency.

Emission factors

Several conditions influence the emission of heavy metals, such as production rates and raw materials, and these factors can vary from one plant to another. To derive emission factors based on one plant's reported emission figures and production volume and use these factors to estimate emissions at other plants is therefore quite uncertain.

4.2.10.6 Completeness

Major missing emission components are not likely.

4.2.10.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.11 Production of mineral white (plaster)

NFR 2A6

Last update: 11.09.15

4.2.11.1 Description

Two plants producing mineral white in Norway are included in the inventory with their emissions of mercury and particles. The mercury content in the raw materials leads to emission of mercury, and during the production process, particles are emitted.

4.2.11.2 Method

Particles

Emission figures are reported to the Norwegian Environment Agency. Reported emission figures exist since 1992 and figures for 1990 and 1991 are assumed by Statistics Norway, in accordance with the Norwegian Environment Agency, to be the same as the figures reported in 1992. The particles are purified through a fabric filter, and it is assumed by the Norwegian Environment Agency that the particles emitted after the filter are smaller than PM₁₀.

According to TNO (Institute of environmental and energy technology 2002), PM_{2.5} is 30 per cent of TSP, while PM₁₀ is assumed to be the same as TSP. The Norwegian inventory uses this distribution.

BC

Emissions are estimated from a share of PM_{2.5} emissions. As no share for BC was found in the literature, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

Heavy metals

The plants have reported emission figures to the Norwegian Environment Agency since 2000. For one of the plants, historical emissions are based on reported figures for 2000 and

production volumes. For the other plant, emission figures for 1990-1999 are assumed to be the same as the reported figure for 2000, due to lack of production data for previous years. Annual emissions are assumed to be low.

4.2.11.3 Activity data

Production volumes for calculation of historical emissions of mercury for one of the plants are reported to the Norwegian Environment Agency.

4.2.11.4 Emission factors

Emission factors for mercury are derived from historical calculations for one plant, based on reported figures for the first year of reporting and production volumes.

4.2.11.5 Uncertainties

Historical emissions of mercury for both plants are uncertain. For one plant, the emission figures are based on a derived emission factor and production volumes, and do not take into account changes in raw materials and possible cleaning devices. For the other plant, it is assumed, due to lack of historical production data, that the historical emissions are the same as the reported figures for 2000. This is just an estimate and does not consider annual changes in raw materials, production rates, or possible cleaning devices.

The particle size distribution used in the inventory is not specific for the plants. The particles emitted might therefore have another distribution than the one suggested by TNO, which is used in the inventory.

4.2.11.6 Completeness

Major missing emission components are not likely.

4.2.11.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.12 Construction and repairing of vessels - Sandblasting

NFR 2A6

Last update: 14.09.15

4.2.12.1 Description

Five plants constructing and repairing vessels are included in the inventory with their particle emissions. One of the plants was closed down in 2000. Emission of particles is due to the different processes during construction and repairing of vessels, but most of the particles are emitted from sandblasting.

4.2.12.2 Method

Particles

Emission figures are reported to the Norwegian Environment Agency.

For four of the five plants, there are no information regarding cleaning device, but it is assumed

by the Norwegian Environment Agency that they have fabric filter and/or wet washer. For the last one, particle emissions are purified in cyclones, and the size of the particles emitted is larger than PM₁₀.

It is difficult to decide particle size of the particles emitted based on the above information. It is however assumed by the Norwegian Environment Agency that most of the particles are larger than PM₁₀ and therefore no PM_{2.5} and PM₁₀ is considered for this category. Thus, no BC emission has been estimated.

4.2.12.3 Uncertainties

The size of the particles emitted is uncertain and will depend on the cleaning device used at each plant. The different activities during construction and repairing can also result in emission of particles of different sizes.

4.2.12.4 Completeness

Emission of particles is often a source of heavy metal emissions since particles often contain heavy metals. Type of metals will however depend on the origin of the particles. Metals might therefore be emitted during sandblasting and repairing/construction of vessels. There are however no data available for calculating emissions of heavy metals.

4.2.12.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.2.13 Leather preparing

NFR 2A6

Last update: 05.02.15

4.2.13.1 Method

NH₃

NH₃ is used to adjust the pH level in the fattening and colouring process in leather preparing. This means that NH₃ is dissolved in an aqueous solution to feed fatty substances to leather. One plant reports emission figures for NH₃ to the Norwegian Environment Agency. Emission figures are available from 1994. Emissions for the years 1990-1993 are assumed by Statistics Norway and the Norwegian Environment Agency to be the same as the reported figure for 1994. The emission of NH₃ reported by the plant is equal to the consumption of NH₃.

4.2.13.2 Uncertainties

It is not clear if it is correct to assume that all NH₃ consumed is emitted to air. This assumption may have to be revised.

4.2.13.3 Completeness

Major missing emission components are not likely.

4.2.13.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description

of the general QA/QC procedure.

4.2.14 Production of asphalt

NFR 2A6

Last update: 22.12.16

4.2.14.1 Method

Dioxins

Asphalt preparations and asphalt recycling are supposed to be a possible dioxin source, especially in countries using extensive recycling, and that use salt on the roads during winter. A lot of salt is used on Norwegian roads during winter, and when this asphalt is heated during recycling, it is assumed to give emissions of dioxins (Hansen 2000).

4.2.14.2 Activity data

The activity data used is the annual weight of asphalt used for road paving in Norway, EBA (2014).

4.2.14.3 Emission factors

BC

Emissions are estimated from a share of PM_{2.5} emissions. As no share for BC was found in the literature, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

Dioxins

Two emission factors are found in the literature. According to SFT (2001), the Oslo and Paris Convention (OSPAR) suggests an emission factor of 0.047 µg/tonne asphalt. This emission factor is however assumed to be very high since it is based on data from a plant only re-circulating old asphalt. Fyns Amt (2000) operates with a much lower emission factor, which probably reflects dioxin emissions from preparation of new asphalt. Since Norway both makes new asphalt and recycles old asphalt, it is assumed that an emission factor in between those suggested from OSPAR and Fyns Amt would be most correct for Norwegian conditions (Table 4.5).

Table 4.5. Dioxin emission factor for asphalt production. µg I-TEQ/tonne produced asphalt

Source	Emission factor
SFT (2001)	0.047
Fyns Amt (2000)	0.0022
Emission factor chosen	0.025

4.2.14.4 Uncertainties

The emission factor used, for estimating dioxins, is uncertain.

4.2.14.5 Completeness

No major missing emission components are likely.

4.2.14.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3 Chemical Industry

NFR 2B

Last update: 10.01.17

In the Norwegian emission inventory, there are 13 different activities included under chemical industry. Nearly all emission figures from this industry included in the inventory are reported from the plants to the Norwegian Environment Agency. Production of carbides causes emissions of many components, but most of the other activities within the sector chemical industry cause only emissions of one or two components (Table 4.6).

Table 4.6. Chemical industry. Components emitted and included in the Norwegian inventory

Chemical industry	CO	NO _x	NMVOC	SO ₂	NH ₃	PM	BC	HM	POP
Production of:									
Ammonia	NE	IE ¹	NA	NA	NE	NE	NE	NA	NA
Nitric acid	NA	R	NA	NA	R	R	E	NA	NA
Silicon carbide	E	NA	R	R	NA	R	NE	R	R
Calcium carbide	NA	R	R	NA	NA	R	E	R	NA
Titanium dioxide	NE	NA/R	NE	R	NE	R	E	R	NA/R
Methanol	NA	R	R	NA	NA	NA	NA	NA	NA
Sulphuric acid	NA	NA	NA	R	NA	NA	NA	NA	NA
Plastic	NA	NA	R	NA	R	R	E	NA	R
Explosives	NA	R	NA	NA	NA	NA	NA	NA	NA
Chloralkali	NA	NA	NA	NA	NA	NA	NA	R	NA
Pigments	NA	NA	NA	NA	NA	NA	NA	R	NA
Soap	NA	NA	NA	NA	NA	R	E	NA	NA
Paint/varnish	NA	NA	NA	NA	NA	R	E	NA	NA

E = Figures estimated by Statistics Norway.

R = Figures reported by the plant to the Norwegian Environment Agency.

NA = Not Applicable.

IE = Included Elsewhere.

¹ Included in reported figures for nitric acid and other fertilisers.

Source: Statistics Norway/Norwegian Environment Agency

4.3.1 Ammonia Production

NFR 2B1

Last update: 16.02.15

4.3.1.1 Description

In Norway, ammonia is produced by catalytic steam reforming of wet fuel gas (containing ethane, propane and some butane). This is one of the steps during fertiliser production. Hydrogen is needed to produce ammonia, and wet fuel gas is the basis for the production of hydrogen.

4.3.1.2 Method

NO_x

During the production of ammonia there are some non-combustion emissions of NO_x. These emission figures are included in the reported NO_x emission from nitric acid production and production of other fertilisers.

4.3.1.3 Uncertainties

The uncertainties in the figures reported by the plant are believed to be limited. Uncertainty estimates are given in Appendix C.

4.3.1.4 Completeness

Major missing emission components are not likely.

4.3.1.5 Source specific QA/QC

The plant reports the emissions of NO_x to the Norwegian Environment Agency. They examine the trends as a quality check.

4.3.2 Production of nitric acid

NFR 2B2

Last update: 10.01.17

4.3.2.1 Description

There are two plants in Norway where nitric acid is produced. Nitric acid is used as a raw material in the manufacture of nitrogenous-based fertiliser. The production of nitric acid (HNO₃) generates NO_x as by-products of high temperature catalytic oxidation of ammonia (NH₃). The production of nitrogenous-based fertiliser also leads to emissions of particles.

4.3.2.2 Method

NO_x

The two plants report the emissions of NO_x to the Norwegian Environment Agency.

NH₃

Emission figures for NH₃ are annually reported to the Norwegian Environment Agency.

Particles

Both plants report emission figures to the Norwegian Environment Agency and have done so since 1990 and 1992. One of the plants has also reported emissions from combustion, but since it is only 1 per cent of the non-combustion emissions, these figures are included in the figures for non-combustion emissions.

In lack of plant specific information regarding particle size distribution of the emitted particles, Statistics Norway uses the distribution given by TNO (Institute of environmental and energy technology 2002) for production of nitrogenous-based fertilisers where PM₁₀ is 0.8*TSP and PM_{2.5} is 0.6*TSP.

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). BC=1,8 per cent of PM_{2.5}.

4.3.2.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

There is uncertainty regarding the size of the particles emitted since there is no plant specific information available. The distribution recommended by TNO is used in lack of other data.

4.3.2.4 Completeness

Major missing emission components are not likely.

4.3.2.5 Source specific QA/QC

The plants report the production of HNO₃ to the Norwegian Environment Agency. They compare the trends in the production data with the trend in N₂O emission and use this as a quality check.

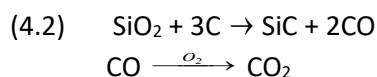
4.3.3 Silicon carbide

NFR 2B5

Last update: 12.01.17

4.3.3.1 Description

Silicon carbide (SiC) is produced by reduction of quartz (SiO₂) with petrol coke as a reducing agent.



In the production of silicon carbide CO are released as by-products from the reaction between quartz and carbon. Sulphur, NMVOCs, particles, heavy metals and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) may also be emitted during the production process. Sulphur originates from the petrol coke.

4.3.3.2 Method

NMVOC

Emission figures are reported to the Norwegian Environment Agency by the plants. The emissions are calculated by multiplying annual production of silicon carbide by an emission factor. From 2007 and onwards, the emission factor is based on measures made once a year. For previous years, an average of the measured emissions in 2007 and 2008 is applied.

CO

The emissions of CO are calculated by Statistics Norway from the consumption of petrol coke and an emission factor in accordance with the IPCC Guidelines (IPCC 1997).

SO₂

Emission figures are reported to the Norwegian Environment Agency by the plants. The emissions are calculated from the consumption of petrol coke in dry weight and the sulphur content in the coke. It is assumed that 3 per cent of the sulphur is left in the product or as wastage.

Particles

Emission figures for particles are reported to the Norwegian Environment Agency. Two of the plants have reported since 1990 while the third has reported since 1991. Emission figures for 1990 for this plant are assumed by Statistics Norway and the Norwegian Environment Agency to be the same as the reported figure for 1991. For one of the plants, reported figures have not been used in the inventory for 1990-1993, since the plant means these emission figures are not representative, but a result of different measurement and calculation methods. For this plant, reported emission figures for 1994 have been used for 1990-1993.

There is no detailed information about the particle size distribution for the emissions from silicon carbide production. The Norwegian Environment Agency assumes the emissions have the same particle size distribution as emissions of particles from production of ferroalloys, where all particles are expected to be smaller than $PM_{2.5}$. This is however an uncertain estimate. This leads to a distribution where $TSP=PM_{10}=PM_{2.5}$.

Heavy metals

Emission figures have been reported to the Norwegian Environment Agency since 1999/2000. For Pb, Hg and Cd, historical emissions are based on emission factors derived from reported emission figures and production rates for the first year of reporting. Using these emission factors for each plant together with production rates for previous years, historical emissions have been calculated. Cd is reported from one plant for the years after 1992. The calculations for Pb and Cd have been corrected for dust regulations, while emissions of mercury are not affected by these regulations.

Historical emissions of Cu, Cr and As are based on dust emissions for each plant. This has been recommended by the Norwegian Environment Agency, since historical production rate data lack for some years and because changes in emissions will be easier to find when installation of dust control systems reduces the emissions of these metals. Emissions of As are reported to the Norwegian Environment Agency from one plant. Reported figures exist since 1992, and emissions in 1990 and 1991 are assumed to be the same as reported figures in 1992.

Emission figures for Cu, Cr and Pb are annually reported for all the three plants. In 1999, the plants also reported Hg and Cd due to a heavy metal investigation under the leadership of the Norwegian Environment Agency. After 1999, the plants have not been required to report these metals due to low emissions. Still, one of the plants have reported Cd and Hg figures for all following years, whereas another has reported only Cd; for this plant the 1999 figure for Hg has been used for all later years. For the plant which now has been closed down, the 1999 figures for both Cd and Hg have been used for all later years when the plant still was operating.

POPs

Emission figures for PAH are reported from the plants to the Norwegian Environment Agency. Two of the plants have reported emissions since 1991, while the third one has only reported since 1997. Historical emissions back to 1990 have been calculated based on production rates and an emission factor derived from the first year of reporting and production rate for that year. No PAH profile is available for this source, so lacking of other information, the same profile as for aluminium production is used (Table 4.7). No emissions of dioxins are reported or calculated.

Table 4.7. Distribution of PAH emissions from silicon carbide production. Ratio X^1 /PAH

Component	Distribution of PAH emissions (ratio)
PAH (Norwegian standard)	1
PAH-4 (CLRTAP)	0.15

¹ X is either PAH, PAH-6 or PAH-4.

Source: Finstad et al. (2001)

Table 4.8. Distribution of PAH-4 emissions from silicon carbide production. Share of PAH-4

Component	Distribution of PAH emissions (ratio)
benzo(a)pyrene	0.2
benzo(b)fluoranthene	0.45
benzo(k)fluoranthene	0.25
indeno(1,2,3-cd)pyrene	0.1

Source: Norwegian Environment Agency (2016): Expert judgement, Oslo, Norway

4.3.3.3 Activity data

The activity data used by the plants for the calculation of SO₂ emissions is the consumption of petrol coke in dry weight. The activity data used by Statistics Norway for the calculation of CO emissions is the consumption of petrol coke reported to Statistics Norway. Historical calculations of particle emissions are based on annual production rates and dust emission figures reported to the Norwegian Environment Agency.

4.3.3.4 Emission factors

CO

CO emissions are calculated from the consumption of petrol coke, using a factor of 0.4 tonnes CO/tonnes petrol coke, as recommended by Rosland (1987).

NMVOC

From 2007 and onwards, the emission factor is based on measurements made once a year. The emission factors for 2007 are 10.906 tonne NMVOC/kilotonne Sic for one of the plants in operation and 10.84 tonne NMVOC/kilotonne Sic for the other. For previous years, the emission factor for the latter plant has been more or less constant whereas the emission factor for the first plant varies.

4.3.3.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Heavy metals

The historical calculations for heavy metals are based on derived emission factors for each plant

and either production or dust data for previous years, and can only be seen as estimates. The emission figures reported also vary from one year to another, and this is assumed to be, in addition to differences in raw materials, a result of few and uncertain measurements. For the one plant that has not reported emission figures for Hg and Cd since 1999, the same emission figures as those reported in 1999 are used for later years. For the other plant, emissions of Cd have been reported for all years since 1992. Emission figures for Hg have not been reported since 1999. The emission figure for 1999 is used for later years. This is also highly uncertain, but the emission figures are very small and have only marginal impact on the total emissions of these metals.

Particles

The particle size distribution used is not specific for production of silicon carbide, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as the first year of reporting. This is uncertain and a result of lack of better data.

4.3.3.6 Completeness

Major missing emission components are not likely.

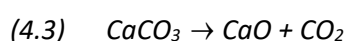
4.3.4 Production of calcium carbide

NFR 2B5

Last update: 10.01.17

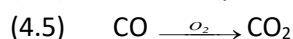
4.3.4.1 Description

One plant in Norway was producing calcium carbide until 2003. The production of calcium carbide generates CO₂ emissions when limestone is heated and when petrol coke is used as a reducing agent. The process can be described through the following equations:



which takes place when limestone (calcium carbonate) is heated.

and



where petrol coke is used as a reducing agent to reduce the CaO to calcium carbide.

Some of the carbon from petrol coke will be sequestered in the product, but not permanently. Thus, this carbon is included in the emission estimate. NMVOC originates from the use of petrol coke in the production process, and NO_x is mainly produced during the high temperature oxidation of nitrogen in the air. Particles are also emitted during the production process. Emission of heavy metals is a result of the heavy metal content in the raw materials.

4.3.4.2 Method

NO_x

Emission figures for NO_x were annually reported to the Norwegian Environment Agency. The reported values are based on calculations.

NMVOC

Reported figures were annually reported to the Norwegian Environment Agency, based on calculations.

Particles

Emission figures for particles were reported from 1992. Figures for 1990 and 1991 are assumed to be the same as for 1992. It does not exist any detailed information about the particle size distribution of the emissions from calcium carbide production. The Norwegian Environment Agency assumes that the emissions are in the same order as emission of particles from production of ferroalloys, where all particles are expected to be smaller than PM_{2.5}. This is however an uncertain estimate. A particle size distribution where PM₁₀ and PM_{2.5} is expected to be the same as TSP, is used in the Norwegian Inventory.

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). BC=1,8 per cent of PM_{2.5}.

Heavy metals and POPs

Emission figures for heavy metals were reported to the Norwegian Environment Agency from 1999. Historical emissions are calculated based on production rates for Pb, Cd and Hg, and based on particle emissions for As, Cu and Cr.

No emission figures for PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) or dioxins are available.

4.3.4.3 Activity data

Particle emissions used in the calculations of As, Cu and Cr have been reported to the Norwegian Environment Agency.

4.3.4.4 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Heavy metals

Historical emissions are based on a derived emission factor for the first year of reporting (1999) and calculated with production/particle emission figures for previous years. This is uncertain and only an estimate in lack of other data.

Particles

The particle size distribution used is not specific for production of calcium carbide, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order of magnitude as for

the first year of reporting. This is uncertain and a result of lack of better data.

4.3.4.5 Completeness

Major missing emission components are not likely.

4.3.5 Production of titanium dioxide

NFR 2B6

Last update: 10.01.17

4.3.5.1 Description

One plant in Norway produces titanium dioxide. The ore is crushed and pulverized in mills. The crushed raw material is separated in various steps. Ilmenite and the by-product magnetite are cleaned during acid treatment and flotation. The ilmenite concentrate is drained and the water content is reduced to approximately 3.5 per cent. Emissions of SO₂, heavy metals and particles from the plant are included in the inventory. The particle emissions are a result of the crushing of the ore in the mills and from the annealing furnace, while the heavy metal emissions are due to the metal content in the raw material used.

Another plant produces titanium dioxide slag and also pig iron as a by-product. The raw material is the mineral ilmenite, and coal is used as a reducing agent. SO₂ originates from the sulphur in the reducing agent used, while NO_x is produced primarily by the high temperature oxidation of nitrogen in the air. Heavy metal emissions are due to the metallurgical melting process and the content of heavy metals in the raw materials used.

4.3.5.2 Method

SO₂

The emission figures for SO₂ are based on calculations and are reported annually to the Norwegian Environment Agency.

NO_x

The emission figures for NO_x for the plant producing titanium dioxide slag are estimated and reported to the Norwegian Environment Agency.

Particles

Since 1990, emissions of particles have been reported annually to the Norwegian Environment Agency. The particles are assumed to be of a size less than PM_{2.5}.

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). BC=1,8 per cent of PM_{2.5}.

Heavy metals and POPs

Both plants report emission figures to the Norwegian Environment Agency.

One plant reported emission figures for Pb, Cd and Hg for the period 1990 to 1999. After 1999, there has not been any reporting, as a result of very small emission figures. No emissions of

persistent organic pollutants are reported or calculated.

The other plant reports emission figures for Pb, Cd, Cr, Cu, As and Hg. Emissions exist from 1990, 1992 or later, depending on type of heavy metal. For dioxins and PAH, reported figures have only been available from 1999. In lack of production rate data for previous years, it has been assumed that yearly emissions are the same as in the first year of reporting. PCB emissions have been measured and reported since 2006. Emissions from 1990 to 2006 are based on reported emissions from 2006. Emission figures for PAH are reported from the plant to the Norwegian Environment Agency. No PAH profile is available for this source. The Norwegian Environment Agency suggests a distribution of the emissions where PAH-4 is 15 per cent of reported PAH emissions. Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are estimated using the same PAH-profile as for aluminium production.

4.3.5.3 Uncertainties

Heavy metals and POPs

Reported emission figures vary from one year to another, partly due to differences in raw materials, but mainly as a result of uncertain measurements. The reported figures are based on a limited number of measurements, and the emissions will vary from minute to minute, since the production of pig-iron is a non-continuous process. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as the first year of reporting. This is uncertain and a result of lack of better data.

Particles

The particle size distribution used is only an assumption, and we can not preclude that the distribution is different from the one used in the inventory.

4.3.5.4 Completeness

Major missing emission components are not likely.

4.3.5.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.6 Production of methanol

NFR 2B10A

Last update: 22.12.16

4.3.6.1 Description

One plant in Norway produces methanol. Natural gas and oxygen are used in the production of methanol. The conversion from the raw materials to methanol is done in various steps and on different locations at the plant. NMVOC are emitted during the production process. Emissions from flaring of natural gas in connection with production of methanol are now reported under 2B10A, as recommended by UNFCCC's expert review team.

4.3.6.2 Method

The plant reports emission figures for NMVOC and NO_x to the Norwegian Environment Agency. The reported emissions are based on measurements. Emissions from flaring of natural gas are estimated by multiplying the amount of gas flared with the emission factors shown in Table 4.9.

Table 4.9. Emission factors for flare

Component	Flare natural gas kg/1000 Sm ³
SO ₂	0
CO	1.5
NO _x	¹
Particles	0.0018
NMVOC	0.06
	mg/1000 Sm ³
Pb	0.25
Cd	1.7
Hg	1
Cu	16
Cr	21
As	3.8
Dioxins	0.00005
Benzo(a)pyrene	0.02
benzo(b)fluoranthene	0.04
benzo(k)fluoranthene	0.02
indeno(1,2,3-cd)pyrene	0.02

¹ Reported to the Norwegian Environment Agency since 2000.

Source: Statistics Norway/Norwegian Environment Agency. PAH: EEA (2016)

BC emissions have been estimated using the same emission factor as for flaring of natural gas 1B2c.

4.3.6.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

4.3.6.4 Completeness

Major missing emission components are not likely.

4.3.6.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.7 Production of sulphuric acid

NFR 2B10A

Last update: 05.02.15

4.3.7.1 Description

Three plants in Norway produced sulphuric acid until March 2006 when one of them was closed down. The production of sulphuric acid leads to emissions of SO₂. All the three plants report the emissions from the production to the Norwegian Environment Agency, but only one plant have

specified that the emissions come from the production of sulphuric acid. For the two other plants, the emissions have been included in the reported emissions from the plants' main production (production of nickel and zinc, respectively).

4.3.7.2 Method

The plant reports annually emission figures for SO₂ to the Norwegian Environment Agency. The reported figures are based on measurements.

4.3.7.3 Uncertainties

No source specific uncertainty is known.

4.3.7.4 Completeness

Major missing emission components are not likely.

4.3.7.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.8 Production of plastic

NFR 2B10A Last update: 10.01.17

4.3.8.1 Description

Three plants report emissions to the Norwegian Environment Agency under this source category. One of the plants produces ethylene, one propylene and polyethylene, and the third plant has vinyl chloride production. Two of the reporting plants were merged up to 2001. Various components are emitted during the production of plastic. NMVOC emissions are from leakages in the process.

During the production process of ethylene and vinyl chloride there is an oxide chloride step for production of ethylene chloride, followed by cracking to vinyl chloride monomer and hydrochloric acid. Various chloride components are produced during these processes, including dioxins. However, most of the dioxins end up in the EDC-tar, which is combusted in an own chloride recycling installation. Particles (PVC-dust) are also emitted during the production of vinyl chloride.

Emissions from flaring of fuel gas in connection with production of plastic are now reported under 2B10a.

4.3.8.2 Method

NH₃ and NMVOC

Emission figures are annually reported to the Norwegian Environment Agency. Reported NMVOC emissions are based on measurements. The emissions of NH₃ are regarded as equal to use. As some of the ammonia is stored in the product, the emissions are probably somewhat overestimated.

Particles

Emission figures have been reported to the Norwegian Environment Agency since 1992. Emission figures for 1991 and 1990 are assumed to be the same as reported figures in 1992. The particle emissions have decreased since 1996 as a result of installation of cleaning devices. The emissions are purified in cyclones, but there is no available information regarding particle size. In lack of plant specific information, the distribution $TSP=PM_{10}=PM_{2.5}$, as in TNO (Institute of environmental and energy technology 2002), is used in the calculation.

BC

Emissions are estimated from a share of $PM_{2.5}$ emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2013). $BC=1,8$ per cent of $PM_{2.5}$.

Dioxins

The plant producing vinyl chloride reports dioxin emission figures. Figures are reported since 1990 except for 1992 and 1994. Emission figures for 1992 and 1994 are based on the reported data for 1991 and 1993.

HCB

The plant producing vinyl chloride reports HCB emission figures since 1996. Emissions from 1990 to 1995 are based on the 1996 reported emissions.

PCB

PCB emissions have been reported since 2010. Emissions from 1990 to 2010 are based on the 2010 reported emissions.

4.3.8.3 Uncertainties

It is difficult to measure leakages of NMVOC and therefore the uncertainty is regarded as being high.

The particle size distribution used is not specific for the plants, and the particles emitted might therefore have another distribution than the one suggested by TNO.

4.3.8.4 Completeness

Major missing emission components are not likely.

4.3.8.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.9 Production of explosives

NFR 2B10A

Last update: 05.02.15

4.3.9.1 Description

There has been one plant in Norway producing explosives, but the plant was closed down in

2001. Nitric acid was used as a raw material in the manufacture of explosives, and during the production of nitric acid, NO_x was emitted.

4.3.9.1.1 Method

NO_x

Emission figures were annually reported to the Norwegian Environment Agency, and the figures were based on calculations.

4.3.9.2 Uncertainties

No source specific uncertainty is known.

4.3.9.3 Completeness

Particles

Reported emission figures to the Norwegian Environment Agency exist only for 1997-1999.

Annual emissions were so low that they have not been included in the Norwegian inventory.

4.3.9.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.10 Chloralkali production

NFR 2B10A

Last update: 05.02.15

4.3.10.1 Description

One plant in Norway produced chloralkali until 2005. Before 1997, mercury was used in the chloralkali production and emitted during the process. In 1997, the plant changed its production process and stopped using mercury, but in the following years there were still some mercury emissions.

4.3.10.2 Method

Hg

Emission figures were reported to the Norwegian Environment Agency.

4.3.10.3 Uncertainties

No source specific uncertainty is known.

4.3.10.4 Completeness

Major missing emission components are not likely.

4.3.10.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.11 Production of pigments

NFR 2B10A

Last update: 05.02.15

4.3.11.1 Description

Two plants are included in the inventory. One plant produces copper oxide for bottom paint and emits copper to air during the production process. Emissions of Cd and Pb have been reported since 2002. Emissions for 1990-2001 are set to be the same as the reported figure in 2002. Also minor amounts of arsenic and chromium are emitted. The other plant produces zinc chromate, and chromium is emitted.

4.3.11.2 Method

Emission figures are reported to the Norwegian Environment Agency.

4.3.11.3 Uncertainties

Reported emission figures for 1990 and 1991 for the plant producing zinc chromate are not occurring. In the inventory, the same figure as reported for 1992 is used for 1990 and 1991.

4.3.11.4 Completeness

Major missing emission components are not likely.

4.3.11.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.12 Production of soap

NFR 2B10A

Last update: 10.01.17

4.3.12.1 Method

Two plants producing soap have reported emission figures for particles to the Norwegian Environment Agency. One of the plants has only reported for 1990 and 1991. The plant has after 1991 had a temporary permission without reporting requirements and is therefore not included after 1991 due to lack of data. The other plant reported figures for 1992-1994. Emissions for 1990 and 1991 are assumed to be the same as reported figure in 1992, while emissions for 1995-1997 are assumed to be the same as reported figure in 1994. Annual emission figures are low.

The particles have been purified through filters and scrubbers and the Norwegian Environment Agency assumes the sizes of the particles are smaller than PM_{2.5}. BC emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2016). BC=1,8 per cent of PM_{2.5}.

4.3.12.2 Uncertainties

For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as reported in one of the other years. This is uncertain and a result of lack of better data.

4.3.12.3 Completeness

Major missing emission components are not likely.

4.3.12.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.3.13 Paint and varnish production

NFR 2B10A

Last update: 10.01.17

4.3.13.1 Method

One plant producing paint has reported emission figures for particles to the Norwegian Environment Agency since 1995, after first getting an emission permit in 1994. Annual emissions are small. It is assumed by the Norwegian Environment Agency that the particles emitted are smaller than PM_{2.5}. BC emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for 'general' chemical industry is used, EEA (2016). BC=1,8 per cent of PM_{2.5}.

4.3.13.2 Uncertainties

No source specific uncertainty is known.

4.3.13.3 Completeness

Major missing emission components are not likely.

4.3.13.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure QA/QC procedure

4.4 Metal production

NFR 2C

Last update: 10.01.17

Metal production in Norway includes plants producing iron and steel, ferroalloys, aluminium, nickel and zinc and also magnesium until spring 2006. Production of anodes is also included in this chapter. As shown in Table 4.10, most of the figures in the national inventory are from the plants' annual reports to the Norwegian Environment Agency.

Table 4.10. Metal production. Components emitted and included in the Norwegian inventory

	SO ₂	NO _x	NH ₃	NMVOC	CO	PM	BC	HM	POP
Production of:									
2C1 Iron and steel	NE	NE	NE	NE	NE	R	E	R	R
2C2 Ferroalloys	R	R	NE	E	NE	R	E	R	R
2C3 Primary aluminium	R	E	NE	NE	E	R	E	R	R/E
2C3 Secondary aluminium	NA	NA	R	NA	NA	R	E	R	R
2C4 Magnesium	R	NA	NA	NA	R	R	NA	R	R
2C6 Zinc	R	NE	NE	NE	NE	R	NA	R	NE
2C7B Nickel	R	R	R	NE	NE	R	NA	R	NE
2C7C Anodes	R	R	NE	NA	NA	R	E	R	R

E = Figures estimated by Statistics Norway (Activity data * emission factor). R = Figures reported by the plant to the Norwegian Environment Agency. NA = Not Applicable.

Source: Statistics Norway/Norwegian Environment Agency

4.4.1 Production of iron and steel

NFR 2C1

Last update: 22.12.16

4.4.1.1 Description

Several plants are included in the time series for the production of iron and steel, but not all plants are currently in production. The components included in the inventory are particles, black carbon, heavy metals and POPs. One plant producing titanium dioxide slag also produces pig iron as a by-product, but the emissions from this plant are registered under 2B6.

4.4.1.2 Method

Particles

One plant has reported figures since 1990 while the other only has reported since 1998. For this plant, historical emissions in the period 1990-1997 have been assumed to be the same as the reported figure in 1998, since production rate data for previous years are not available.

The Norwegian Environment Agency assumes that the particles emitted in the production of iron and steel are smaller than PM_{2.5}. We can, however, not disregard that some of the particles emitted are larger than PM_{2.5}.

BC

Emissions have been estimated as a share of PM_{2.5} emissions. Measurements from one plant showed that 0.1 per cent of particles are carbon dust. This value is used to estimate BC. As no information on the share of BC and OC was found in the literature for iron and steel production, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM). Hence, BC emissions represents 0.05 per cent of PM_{2.5} emissions.

Heavy metals and POPs

Heavy metal emissions are due to the metallurgical melting process and the content of heavy metals in the raw materials used. One plant reports emission figures to the Norwegian Environment Agency. Reported figures for heavy metals (Pb, Cd, Cr, Cu, As and Hg) exist from

1990, 1992 or later, depending on type of heavy metal. For dioxins and PAH, reported figures have only been available from 1997 and 1999. Diffuse emissions have been included from one plant. In lack of production rate data for previous years, it has been assumed that yearly emissions are the same as in the first year of reporting.

Plants reports HCB emissions to the Norwegian Environment Agency since 2011. Emissions for the previous years have been estimated using the emission factor from the EMEP inventory guidebook (EEA 2013) and the crude steel production.

Plants reported PCB emissions in 2010. Emissions for the other years have been estimated using the data reported in 2010.

4.4.1.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Heavy metals and POPs

Reported emission figures vary from one year to another, partly due to differences in raw materials, but mainly as a result of uncertain measurements. The reported figures are based on a limited number of measurements, and the emissions will vary from minute to minute, since the production of iron and steel is a non-continuous process. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as the first year of reporting. This is uncertain and a result of lack of better data.

Particles

The particle size distribution used is only an assumption, and we can not preclude that the distribution is different from the one used in the inventory. For the years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order of magnitude as for the first year of reporting. This is an uncertain estimate due to lack of better data.

4.4.1.4 Completeness

Major missing emission components are not likely.

4.4.1.5 Source specific QA/QC

Annually reported emission figures are first controlled by the Norwegian Environment Agency and then by Statistics Norway.

Adjustments and recalculations have been done for years where reported emission figures seem to be unreasonably high or low compared with previous years. This is applicable when the variations in the reported emission figures do not have a natural explanation.

4.4.2 Production of ferroalloys

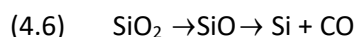
NFR 2C2

Last update: 22.12.16

4.4.2.1 Description

There were 12 plants producing ferroalloys in Norway in 2014. One plant closed down in 2001, two plants were closed down during 2003 and two in 2006. One plant was out of production in 2006, but started up again in 2007.

Ferrosilicon, silicon metal, ferromanganese and silicon manganese are now produced in Norway. Ferrochromium was produced until summer in 2001. Ferrosilicon with 65 to 96 per cent Si and silicon metal with 98-99 per cent Si is produced. The raw material for silicon is quartz (SiO_2). SiO_2 is reduced to Si and CO using reducing agents like coal, coke and charcoal.



The waste gas CO and some SiO burns to form CO_2 and SiO_2 (silica dust).

Some of the CO generated from coal is sold for energy use to other industries. The amount of CO gas sold is hence subtracted from the emissions reported under this category and included in energy use in manufacturing industries and construction (NFR 1A2).

In ferroalloy production, raw ore, carbon materials and slag forming materials are mixed and heated to high temperatures for reduction and smelting. The carbon materials used are coal, coke and some biocarbon (charcoal and wood). Electric submerged arc furnaces with graphite electrodes or consumable Soederberg electrodes are used. The heat is produced by the electric arcs and by the resistance in the charge materials. The furnaces used in Norway are open, semi-covered or covered.

The CO stems from the production process. In open or semi-closed furnaces the CO reacts with air and forms CO_2 before it is emitted. This is due to high temperature and access to air in the process. In a closed furnace the CO does not develop to CO_2 as there is no access to air (oxygen) in the process. The waste gas is then led from the furnace and used as an energy source or flared, and is reported under the relevant energy sectors. The technical specification of the furnaces is irrelevant since emissions are calculated using a mass balance or calculated by multiplying the amount of reducing agents in dry weight with country specific emission factors

Several components are emitted from production of ferroalloys. SO_2 originates from the sulphur in the reducing agent used, while NO_x is produced primarily by the high temperature oxidation of nitrogen in the air. NMVOC emissions originate from the use of coal and coke in the production processes by producing ferrosilicon and silicon metal. Heavy metals are emitted from the raw materials (ore) during the metallurgical process, and the particles emitted are mainly silica dust generated during the production process.

4.4.2.2 Method

SO_2

Each plant annually reports emission figures to the Norwegian Environment Agency. Some of the sulphur is trapped in the product. For production of ferromanganese and silicon manganese, 98-99 per cent of the sulphur is trapped, while for other ferroalloys it is assumed that about 5

per cent is trapped. The emissions are calculated from the consumption of reducing agents and electrodes and the content of sulphur in the materials.

NO_x

Emissions of NO_x originate from production of ferrosilicon and silicon metal. Ferromanganese, ferrochrome and silicomanganese do not have significant emissions of NO_x. Emission figures are annually reported by each plant to the Norwegian Environment Agency. The reported emissions are calculated either from the production of metal and metal specific emission factors, see Table 4.13, or on the basis of continuous measurements.

NMVOC

The emissions are estimated by Statistics Norway from the consumption of reducing agents and an emission factor.

Particles

All plants producing ferroalloys report emission figures to the Norwegian Environment Agency. Some have reported since 1990, others since 1992. For plants reported since 1992, emission figures from 1990 and 1991 have been assumed to be the same as reported figures in 1992. According to the ferroalloy industry, particles emitted are smaller than PM_{2.5} (Eikeland, *pers.comm.*⁸). This is, however, an assumption, and we can not preclude that some of the particles might be larger than PM_{2.5}. In the inventory, we have decided to use this distribution for all particles emitted from the production of ferroalloys. This means that TSP=PM₁₀= PM_{2.5}.

BC

Emissions have been estimated as a share of PM_{2.5} emissions. Measurements of particles composition from several plants were used to estimate the dust carbon content. This value was used to estimate BC. As no information on the share of BC and OC was found in the literature for ferroalloys production, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

Heavy metals

Emission figures for heavy metals are reported from all plants producing ferroalloys after the Norwegian Environment Agency in 1999 imposed larger metallurgical plants to map their emissions of heavy metals. Most plants have therefore reported figures to the Norwegian Environment Agency since 1999, but some reported for the first time in 2000 and 2001. An emission factor has been derived for each plant, based on the emission figure and production rate for the first year of reporting. These emission factors have been used together with production rates for each year to calculate the emissions back to 1990 for each plant.

Dioxins

All plants producing ferrosilicon report emission figures for dioxins to the Norwegian

⁸ Eikeland (2002): Personal information, e-mail dated 29/05 2002. Elkem@elkem.no

Environment Agency. It varies, however, when the plants started reporting, so calculations of historical figures back to 1990 have been necessary. An emission factor was derived for each plant based on reported emission data and production rates, and this factor was used to calculate historical emissions based on production rates for each year.

None of the four plants producing ferromanganese and ferrochromium⁹ report emission figures for dioxins to the Norwegian Environment Agency. The reason is probably that the emissions are so small that they are not measured and therefore not reported (the Norwegian Pollution Control Authority, *pers. comm.*¹⁰). Instead, the emissions are calculated by Statistics Norway based on the general emission factor for combustion of coke and coal in the industry (Table 4.15).

PCB

As for dioxins emissions, PCB emissions are only considered in the ferrosilicon production. Plants reported emissions in 2010 and reported data has been used to estimate emissions for the whole period.

PAHs (Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)

Emissions of PAH from the production of ferroalloys are reported to the Norwegian Environment Agency for plants producing ferrosilicon and silicon metal. All these plants have reported emission figures since 2000. Historical emissions back to 1990 have been calculated based on production rates for each year and an emission factor derived for each plant based on reported figures for 2000, 2001 and 2002. Reported figures and historical calculations are only done for plants producing ferrosilicon and silicon metal. This is based on the assumption that these alloys are produced in open ovens and therefore cause larger emissions of PAH compared to other alloys that are produced in closed ovens, and are assumed to cause no or minor emissions of PAH. No PAH profile is available for this source. The Norwegian Environment Agency suggests a distribution of the emissions where PAH-4 is 15 per cent of reported PAH emissions.

The PAH emission figures are reported according to Norwegian Standard, but no PAH profile for benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene is available. In lack of other data, the same profile as for aluminium production is used.

Table 4.11. Distribution of PAH emissions from production of ferroalloys

Component	Distribution of PAH emissions (ratio)
PAH (Norwegian standard)	1
PAH-4 (CLRTAP)	0.15

Source: Finstad et al. (2001)

⁹ The ferrochromium plant was closed down in 2003.

¹⁰ Norwegian Pollution Control Authority (2001): Units for dioxins (dioxins.doc). Personal information C. Benestad, 13/03 2001, Oslo: Norwegian Pollution Control Authority.

Table 4.12 Distribution of PAH-4 emissions from production of ferroalloys. Share of PAH-4

Component	Distribution of PAH emissions (ratio)
benzo(a)pyrene	0.2
benzo(b)fluoranthene	0.45
benzo(k)fluoranthene	0.25
indeno(1,2,3-cd)pyrene	0.1

Source: Norwegian Environment Agency (2016): Expert judgement, Oslo, Norway

4.4.2.3 Activity data

NMVOC

The amounts of reducing agents that are used for the calculation of NMVOC emissions are annually reported to Statistics Norway from each plant.

4.4.2.4 Emission factors

NO_x

The emission factors used by the ferrosilicon plants in the calculations are based on measurements carried out at three plants.

The emission factors in Table 4.13 are based on several measuring campaigns at four different ferroalloy plants that were carried out from 1995 to 2007. Each measurement period lasted 4 to 8 hours with different operation conditions. Based on this, emission factors for different ferroalloys and operational conditions have been established. The measurements have been carried out by Det norske Veritas, Norsk Energi, SINTEF and TÜV.

The silicon plants have applied a new method. They have used online measurement instruments to measure the emissions of NO_x. The measurements were undertaken in 2010. The instrument applied is NEO laser gas and Testo 350 as a control of the results from the NEO laser gas device. So far there are only two plants where the online measurement devices are installed on a permanent basis. For the other plants the online measurement instruments are used periodically to derive emission factors. The following emission factors are used by four plants in one major ferroalloy producing company (kg NO_x/tonne metal produced): 27 (based on measuring campaigns), 34 and 39 (based on online measurements) and 45 (based on a combination of online measurements and campaigns).

The uncertainties associated with the measurements mainly come from measurement of off-gas flow and measurement of concentration of the NO_x in the off-gas. In addition, the periodical measurement campaigns will not include all variations in the emissions gained over time.

Table 4.13. NO_x emission factors for production of ferrosilicon. Kg NO_x/tonne metal produced

	Normal operations	Dryss - charging	Dryss- charging > 750 °C	Source
Ferrosilicon 75 per cent	15.3	7.0	8.3	Measured in 1995 at Rana Metal and the Thamshavn plant 2005
Ferrosilicon 65 per cent	6.0	4.0	5.0	Estimations 1

¹ Estimations means that this emission factor is not measured but estimated by the plants based on general process experiences.

NM VOC

Statistics Norway uses an emission factor of 1.7 kg NMVOC/tonne coal or coke (EPA 1986) in the calculations.

BC

BC is assumed to be 50 per cent of the particles carbon content. Hence, BC emissions represents 3.5 per cent of PM_{2.5} emissions from ferro-manganese production and 0.23 per cent of PM_{2.5} from ferro-silicon production.

Dioxins

The emission factors used by the plants in the calculations are given in Table 4.14.

Table 4.14. Emission factors for production of ferroalloys. µg dioxin/tonne metal produced

	Normal operations	Dryss - charging	Dryss- charging > 750 °C	Source
Silicon metal	3	1.2	0.2	Measured in 1995 at the Fiskaa plant
Ferrosilicon 90 per cent	4	1.2	0.2	Estimations ¹
Ferrosilicon 75 per cent	5	1.2	0.2	Measured in 1995 at Rana Metall
Ferrosilicon 65 per cent	5	1.2	0.2	Estimations
Si96	3	1.2	0.2	Estimations

¹ Estimations means that this emission factor is not measured but estimated by the plants based on general process experiences.

Emission calculations of dioxins for plants not reporting figures to the Norwegian Environment Agency use an emission factor for combustion of coke and coal in the industry (Table 4.15).

Table 4.15. Emission factor used by Statistics Norway to calculate dioxin emissions from production of ferro manganese/chromium

	Emission factor
Coal and coke	1.6 µg/tonne

Source: Bremmer et al. (1994) and Finstad et al. (2002b)

PAH

The emission factors used by the plants in the calculations are given in Table 4.16.

Table 4.16. Emission factors for production of ferroalloys. g PAH /tonne metal produced

	Normal operations	Dryss - charging	Dryss- charging > 750 °C	Source
Silicon metal	3	2.6	1.6	Measured in 1995 at the Fiskaa plant
Ferrosilicon 90 per cent	2	2	1	Estimations ¹
Ferrosilicon 75 per cent	1.5	1.3	0.8	Measured in 1995 at Rana Metal and the Thamshavn plant
Ferrosilicon 65 per cent	1	1.3	0.8	Estimations
Si96	3	2.6	1.6	Estimations

¹ Estimations means that this emission factor is not measured but estimated by the plants based on general process experiences.

4.4.2.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

The inventory uses a particle size distribution which is an assumption from the ferroalloy industry and not based on measurements. We can therefore not preclude that some of the particles might be larger than PM_{2.5}.

Heavy metals and POPs

Historical emissions are based on derived emission factors for the first year of reporting, and calculated using production figures for previous years. This is uncertain since the calculation method does not consider quality changes of the raw materials or changes in the production profile at each plant that can have big impact on yearly emissions.

4.4.2.6 Source specific QA/QC

NO_x, NMVOC and CO

The reported emission figures for NO_x, NMVOC and CO are compared with calculations at Statistics Norway.

Emission figures for NMVOC are controlled by multiplying the amount of reducing agents with an emission factor recommended by EPA (1986).

PAH

PAH was first included in the Norwegian Inventory in 2000, and only two plants producing ferrosilicon and silicon metal reported emission figures to the Norwegian Environment Agency for the year 1999. In 2004, a specific emission factor for each plant was derived based on the plants' reported emission figures for 2000, 2001 and 2002 and production volumes,. These factors were then used to recalculate the plants' historical emissions of PAH. A specific emission factor for each plant was considered better to use for historical emissions, instead of using a default emission factor for all plants. The specific emission factors derived for each plant with the new method were lower than those suggested by Benestad (*pers. Comm.*), and this caused approximately 2-12 per cent lower yearly PAH emissions from 1990 to 1999 from this source.

4.4.3 Production of primary aluminium

NFR 2C3

Last update: 10.01.17

4.4.3.1 Description

There are seven plants in Norway producing aluminium. Both prebaked anode and the Soederberg production methods are used.

In the Soederberg technology, the anodes are baked in the electrolysis oven, while in the prebaked technology the anodes are baked in a separate plant. In general, the emissions are larger from the Soederberg technology than from the prebaked technology. There has been a shift from Soederberg to prebaked technology. In 1990, 57 per cent of the aluminium production in Norway was produced with prebaked technology and the share of aluminium production from prebaked increased to 92 per cent in 2009. Two new plants with prebaked technology were established in 2002 and two plants using Soederberg technology were closed down in 2001 and 2003. Since 2007, three plants using Soederberg technology have been closed down, one in 2007, one in 2008 and one in 2009. There is now only one plant left where Soederberg technology is used.

Production of aluminium leads to emissions of various components, such as SO₂, NO_x, heavy metals and persistent organic pollutants. The emission of SO₂ are from the sulphur in the reducing agents used. NO_x is primarily produced by the high temperature oxidation of nitrogen in the air. All plants also report emissions of particles, heavy metals and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene). Emissions of heavy metals are due to the metal content in the raw materials used and the reducing agents.

4.4.3.2 Method

SO₂

The plants report emission figures of SO₂ to the Norwegian Environment Agency. The figures are estimated by each plant based on the amounts of reducing agents used and their sulphur content. All plants have installed flue gas treatment, like, for example, sea water scrubber.

NO_x

NO_x emissions are estimated by Statistics Norway from the level of production and an emission factor derived from measurements at two Norwegian plants. The figure is rather uncertain.

CO

CO emissions are estimated by Statistics Norway from the level of production and an emission factor from EEA 2016. 120 kg CO/tonn aluminium.

Particles

Emission figures have been reported to the Norwegian Environment Agency since 1990. The Norwegian Environment Agency assumes that the particles emitted are smaller than PM₁₀. According to TNO (Institute of environmental and energy technology 2002), PM₁₀ is 97 per cent

of TSP, and PM_{2.5} is 43 per cent of TSP. The Norwegian inventory uses the particle size distribution suggested by TNO (Institute of environmental and energy technology 2002).

BC

Emissions are estimated from a share of PM_{2.5} emissions. Tier 1 emission factor for BC applicable for aluminium production is used, EEA (2016). BC=2.3 per cent of PM_{2.5}.

Heavy metals

The plants report emission figures to the Norwegian Environment Agency. The first requirement for reporting came in 1999, so emission figures before that are insufficient. The concentrations of heavy metals in the air emissions are very low and therefore impossible to measure.

Emissions are therefore calculated at each plant, based on the mass flow.

Dioxins

Since the process uses coal and coke as reducing agents, it is assumed that production of primary aluminium gives dioxin emissions. Reported figures for dioxins are not available. The emissions are believed to be so small that reporting is not necessary. Emissions are therefore calculated based on the combustion factor for coal in the industry.

HCB

Emissions are so small that primary aluminium plants do not report them. Hence, emissions have been considered negligible and have not been estimated.

PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)

The reported emission data are assumed to be according to Norwegian standard (NS9815). It is further assumed by the Norwegian Environment Agency that the emissions of PAH-4 accounts for 5 -10 per cent of total PAH emissions reported from production of aluminium. Historical emission figures have been calculated based on changes in production of aluminium after the Soederberg method.

The PAH profile has been measured at three plants, in addition benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene have been measured at some plants for the last year. Based on these profiles it is believed that PAH-4 accounts for 10 per cent of total PAH emissions from production of aluminium from one plant, 7.5 per cent is used for the other. Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene have been measured in 2016 and in 2001. Based on these measurements a PAH-profile has been made by Hetland, the Norwegian Environment Agency (*pers. comm*)¹¹. The PAH-4 profile used for aluminium production is shown in Table 4.18.

¹¹ Pers. comm, email from Øyvind Hetland, 22. Nov. 2015, Norwegian Environment Agency.

Table 4.17. Distribution of PAH emissions from production of primary aluminium. Ratio

Component	Distribution of PAH emissions (ratio)
PAH (Norwegian standard)	1
PAH-4 (CLRTAP)	0.05-0.1

Source: Statistics Norway/Norwegian Environment Agency

Table 4.18. Distribution of PAH-4 emissions from production of primary aluminium. Share of PAH-4

Component	Distribution of PAH emissions (ratio)
benzo(a)pyrene	0.2
benzo(b)fluoranthene	0.45
benzo(k)fluoranthene	0.25
indeno(1,2,3-cd)pyrene	0.1

Source: Hetland (2016)

4.4.3.3 Activity data

NO_x

The activity data for the NO_x calculation are production figures, which are reported annually from the plants to the Norwegian Environment Agency.

Dioxins

The calculation of emissions of dioxins is based on consumption of raw materials. The figures are reported annually from the plants to Statistics Norway.

4.4.3.4 Emission factors

NO_x

Statistics Norway uses the emission factor 0.00071 tonnes NO_x/ tonne produced aluminium in the calculations. This emission factor is assumed by the Norwegian Environment Agency and is based on measurements.

Dioxins

Emissions of dioxins are calculated based on the consumption of coal and an emission factor from Bremmer et al. (1994).

Table 4.19. Emission factor used to calculate dioxin emissions from aluminium production.

	Emission factor	Source
Coal and coke	1.6 µg/tonne	Bremmer et al. (1994)

4.4.3.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

The particle size distribution are not reported by the plants. Actual emissions are probably somewhat different from those estimated with the size distribution from TNO (Institute of environmental and energy technology 2002).

4.4.3.6 Completeness

Major missing emission components are not likely.

4.4.3.7 Source specific QA/QC

PAH

In 2014, the Norwegian Environment Agency had audits at all aluminium plants. For the four plants that have emissions of PAH, their systems for monitoring emissions of PAH were checked.

Heavy metals

First requirement for reporting of heavy metals was given in 1999, and the reported figures were that year based on concentration measurements. The concentration of heavy metals in the air emissions are very low and therefore subject to high degree of uncertainty. The reported emission figures showed large differences from plant to plant, also in the cases where the raw materials came from the same supplier. The Norwegian Environment Agency has had a long discussion with the aluminium industry to find a better method to estimate heavy metals from aluminium production. In 2001 it was decided that reported figures should be based on calculations. New calculations have shown that earlier calculations gave too high emissions of heavy metals. It was therefore recommended by the Norwegian Environment Agency to recalculate historical reported data based on the new calculation method. Recalculation of historical data are normally based on production rate data, but due to very low emissions and relative stable production rates, historical data are set to be the same as the first year of reporting.

4.4.4 Production of secondary aluminium

NFR 2C3

Last update: 12.01.17

4.4.4.1 Description

One open mill in Norway is handling secondary aluminium production. Heavy metals and persistent organic pollutants (dioxins and PAHs, (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)) are emitted in the production of secondary aluminium due to the remelting process. Particles are also emitted during the production process. For earlier years there have also been some emissions of NH₃ and SF₆ from another plant which closed down in 2001.

4.4.4.2 Method

NH₃

For the years 1993-2001, emissions of NH₃ were reported from one plant. This plant closed down in 2001.

Particles

The plant has reported emission figures to the Norwegian Environment Agency from 1993. Emission figures for 1990 to 1992 are in the inventory assumed to be the same as the reported figure in 1993. The following particle size distribution is assumed and used in the Norwegian inventory; PM₁₀ is 0.8*TSP and PM_{2.5} is 0.32*TSP (Institute of environmental and energy technology 2002).

BC

BC has been estimated as fraction of PM_{2.5} emissions. Shares of BC in PM_{2.5} given by IIASA (Kupiainen & Klimont 2004) have been used.

Heavy metals and POPs

The figures are reported annually to the Norwegian Environment Agency. Emission figures exist since 1993, and emissions before 1993 have been supposed to be the same as reported figures in 1993.

The emission figures for heavy metals are based on metal analyses of dust samples. Figures of Pb, Cd and Cr have been reported since 1997. Annual figures can vary a lot from one year to another, and therefore we have used mean values for years when the changes can not be explained by the industry. We have assumed that the emission figures for 1990-1996 are the same as reported figures in 1997, since there are no reported figures of heavy metals before 1997.

PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)

No PAH profile is available for this source. The Norwegian Environment Agency suggests a distribution of the emissions where PAH-4 is 15 per cent of reported PAH emissions. Since no PAH profile is available, emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are assumed to be 25 per cent each of PAH-4.

HCB emissions have been reported since 2010. For the period 1990-2009, aluminium production has been used with a emission factor from Japan (Toda 2006) to estimate HCB emissions. Emission factor is 1.7 mg/tonn secondary aluminium.

PCB emissions have been reported from 2006 to 2008 and in 2010. Data reported have been used to build an emission factor and estimate emissions from 1990.

4.4.4.3 Uncertainties

Heavy metals and POPs

The reported figures for heavy metals are estimated based on heavy metal content in the dust samples. The metal content were only analysed for a few dust samples yearly and the reported figures are therefore only a presumption of yearly emission figures. Calculation of emission figures before 1997 are assumed to be the same as reported figures in 1997, and this gives highly uncertain figures since raw materials and production variations may have changed during the period.

The reported emission figures for dioxins and particles vary from one year to another, and it is assumed that this is due to uncertain measurements and process readjustments.

4.4.4.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.4.5 Production of magnesium

NFR 2C4

Last update: 16.09.15

4.4.5.1 Description

There has been one magnesium producing plant in Norway. The plant closed down the production of primary magnesium in 2002, but the production of cast magnesium continued. During 2006, the production of remelting Mg also stopped. From the mid-1970s, both the magnesium chloride brine process and the chlorination process were used for magnesium production. Since 1991, only the chlorination process was in use.

Production of magnesium leads to non-combustion CO emissions. During the calcination of dolomite ($\text{MgCa}(\text{CO}_3)_2$) to magnesium oxide, CO_2 is emitted. During the next step, magnesium oxide is chlorinated to magnesium chloride, and coke is added to bind the oxygen as CO and CO_2 . SO_2 is emitted due to the sulphur in the reducing agent used.

4.4.5.2 Method

CO

Emission figures of CO were reported annually to the Norwegian Environment Agency. These emissions disappeared when the plant closed down the production of primary magnesium in 2002.

SO_2

The SO_2 emissions were estimated from the amounts of reducing agent used (coke) and their sulphur content and reported from the plants to the Norwegian Environment Agency.

Particles

The plant reported emission figures for particles for the first time for the year 1992. Emissions of particles for 1990 and 1991 are assumed to be larger than the reported figure in 1992, since a cleaning device was installed in 1992. Statistics Norway has no information that can be used to estimate emissions in 1990 and 1991, so the inventory uses the reported emission figure for 1992 also for 1990 and 1991. The Norwegian Environment Agency assumes that reported figures also include emissions from combustion.

No information is found regarding the particle size distribution for particles emitted during magnesium production. In lack of other data, we use the same distribution as for aluminium production (PM_{10} is 97 per cent of TSP, and $\text{PM}_{2.5}$ is 43 per cent of TSP).

Heavy metals and POPs

Emission of heavy metals is due to the metal content in the reducing agent used. Emission data of Hg, As, Cr and dioxins were reported to the Norwegian Environment Agency. When the plant closed down the production of primary magnesium in 2002, the emissions of As disappeared. Reported figures of heavy metals have only been available since 2000. Emission figures are calculated back to 1990 based on the production rate for each year.

During the chlorination process and the use of coke as a reducing agent, dioxins and HCB are emitted. Emission figures for dioxins were reported to the Norwegian Environment Agency from 1990 while emissions from HCB have been reported from 1992. For 1990 and 1991, 1992 figures have been considered. As no reports were available in 2004 and 2006, emissions have been estimated using 2003 and 2005 figures.

4.4.5.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

For years where reported emission figures do not exist, Statistics Norway has assumed that emissions are in the same order as that of the first year of reporting. This is uncertain and a result of lack of better data. The particle size distribution used is not specific for production of magnesium, but used due to lack of specific size distribution data for this source. The particle size distribution can therefore only be seen as an estimate.

Heavy metals

Historical emissions are based on a derived emission factor for the first year of reporting and calculated with production figures for previous years. This is uncertain and only an estimate since it does not consider annually changes in raw materials nor possible cleaning devices.

4.4.5.4 Completeness

Major missing emission components are not likely.

4.4.5.5 Source specific QA/QC

The latest reported emission data from the plant were compared with previous reported data and the emissions were compared with the production.

4.4.6 Production of zinc

NFR 2C6

Last update: 05.02.15

4.4.6.1 Description

One plant in Norway produces zinc. SO₂, particles and heavy metals are emitted during the process. Emission of SO₂ originates from the sulphur in the reducing agent used.

4.4.6.2 Method

SO₂

The plant reports emission figures to the Norwegian Environment Agency. The SO₂ emissions are estimated from infrequent measurements combined with calculations.

Particles

Emission figures for particles have been reported since 1991. Emissions for 1990 are assumed to be the same as the reported figure for 1991. It is assumed that of the particles emitted, 90 per

cent is PM₁₀ and 80 per cent is PM_{2.5} (Institute of environmental and energy technology 2002) and this particle size distribution is used in the Norwegian inventory.

Heavy metals and POPs

The plant reports emission figures for Cd, Pb, Hg, Cu, Cr and As. Reported figures exist since 1992, and emissions in 1990 and 1991 are assumed to be the same as reported figures in 1992.

Figures are not reported for PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins.

PCB emissions have been estimated using the emission factor given by EEA (2013). This emissions factor, which amounts to 0.9 mg/tonn of zinc, is used for the whole period.

4.4.6.3 Completeness

Major missing emission components are not likely.

4.4.6.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.4.7 Production of nickel

NFR 2C7B

Last update: 22.12.16

4.4.7.1 Description

One plant in Norway produces nickel. During the production of nickel SO₂, NO_x, NH₃, particles and heavy metals are emitted. CO₂ is emitted in the production of nickel, due to the soda from the production of nickel carbonate and use of coke as a reducing agent, while SO₂ is a result of the sulphur content in the coke used. NO_x is produced primarily by the high temperature oxidation of nitrogen in the air. Emission of heavy metals is due to the metal content in reducing agent used. Particles are also emitted during the production process. PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins are not reported or calculated.

4.4.7.2 Method

SO₂

Emission figures of SO₂ are reported from the plant to the Norwegian Environment Agency based on continuous measurements. Flue gas treatment is installed at the plant.

NO_x

Emission figures of NO_x are annually reported from the plant to the Norwegian Environment Agency. The emission figures are based on calculations.

NH₃

Emission figures based on calculations are annually reported from the plant to the Norwegian Environment Agency.

Particles

Emission figures for particles have been reported to the Norwegian Environment Agency since 1992. Emissions in 1990 and 1991 are assumed to be the same as the reported figure in 1992. The emission permit sets requirements to emissions from the melting furnace, transport, crushing and packing of the raw materials and products. The Norwegian Environment Agency assumes that the particles emitted are smaller than PM_{2.5}. This means that TSP=PM₁₀=PM_{2.5} is used in the inventory.

Heavy metals and POPs

Emission figures for Cu have been reported to the Norwegian Environment Agency since 1990. Reported figures for Cd, Hg and Pb were available from 1990-1994, but because of low emissions the plant stopped reporting these metals.

4.4.7.3 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

Particles

The particle size distribution used is only an assumption and we can not preclude that the distribution might be different than the one suggested. The particle size distribution can therefore only be seen as an estimate.

4.4.7.4 Completeness

Major missing emission components are not likely.

4.4.7.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.4.8 Manufacture of anodes

NFR 2C7C

Last update: 12.01.17

4.4.8.1 Description

Four plants in Norway produce anodes. Three plants produce prebaked anodes and one plant produced coal electrodes. These are alternatives to the use of coal and coke as reducing agents in the production process for aluminium and ferroalloys. The anodes and coal electrodes are produced from coal and coke. The production of anodes and coal electrodes leads to emissions of NO_x, SO₂, particles, BC, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and heavy metals.

4.4.8.2 Method

SO₂ and NO_x

Emission figures of SO₂ are based on measurements while NO_x emissions are calculated by the plants and reported to the Norwegian Environment Agency.

Particles

Production of anodes leads to emission of particles. One of the plants has reported emissions since 1990, while the other one has reported since 1992. Emission figures for 1990 and 1991 are assumed to be the same as the reported figure in 1992 for this plant. The Norwegian Environment Agency assumes that the particles emitted are smaller than PM_{10} , but also expects some to be smaller than $PM_{2.5}$. No information has been found regarding the particle size distribution, so in lack of other data we use the same distribution profile as used for production of aluminium where PM_{10} is 97 per cent of TSP and $PM_{2.5}$ is 43 per cent of TSP.

BC

Emissions have been estimated as a share of $PM_{2.5}$ emissions. Measurements of particles composition from one plant were used to estimate the dust carbon content. This value was used to estimate BC. As no information on the share of BC and OC was found in the literature for ferroalloys production, BC share has been set to be 50 per cent of $PM_{2.5}$. Indeed, the amount of $PM_{2.5}$ is assumed to be equally shared between BC and organic mass (OM).

As the dust carbon content was measured to be 8 per cent, BC emission have been set to 4 per cent of $PM_{2.5}$ emissions.

PAHs (Benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene)

Emission figures for PAH are based on measurements and reported from both plants to the Norwegian Environment Agency. One plant has developed a new and better method for measuring PAH. This method is used for the period 1992 to 2003. The reported figures of PAH are assumed to be according to the Norwegian standard (NS9815). Measurements from production of Soederberg paste (at three Norwegians plants) and a PAH-profile of baked anodes from EPA are used to derive a PAH-profile to find the emission of PAH-OSPAR and PAH-4. Based on these profiles it is assumed that PAH-4 account for 5 per cent of the total PAH emissions (Table 4.20). Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are estimated using the same PAH-profile as for aluminium production, see Table 4.18

Table 4.20. Distribution of PAH emissions from production of anodes. Ratio

Component	Distribution of PAH emissions (ratio)
PAH (Norwegian standard)	1
PAH-4 (CLRTAP)	0.05

Source: Norwegian pollution control authority (SFT 1999a)

Heavy metals

Production of anodes leads to emission of heavy metals due to the metal content in the reducing agents (coke and coal). Emission figures are based on measurements and are reported for arsenic and mercury from one plant since 2001, and for lead since 2004. Emission figures have not been measured or reported before 2001 for As and Hg and before 2004 for Pb and are therefore not available for previous years. Historical emission figures back to 1990 are assumed to be the same as reported figures for 2001 for As and Hg and 2004 for Pb.

4.4.8.3 Uncertainties

Historical calculations of heavy metals from 1990 to 2001 are very uncertain since they are assumed to be the same as reported figures for the first year of reporting (2001). Annual changes in production volumes, coke quality and the amount of heavy metals in the reducing agents are not taken into account, and the historical emissions can only be seen as an estimate in lack of better data.

4.4.8.4 Completeness

Major missing emission components are not likely.

4.4.8.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5 Solvents and product use

NFR 2D

Last update: 22.12.16

This chapter describes emissions from solvents and other products. Use of solvents and products containing solvents result in emissions of non-methane volatile organic compounds (NMVOC).

In addition to solvents emitting NMVOC, there are other products that emit other volatile components. Creosote treated materials and tarry jointing paste cause emissions of PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene). Dioxins are also emitted during road paving with asphalt (2D3B). Emissions of mercury from mercury-containing products as well as emissions from combustion of tobacco are also included in the Norwegian inventory.

4.5.1 Solvent losses (NMVOC)

NFR 2D3A, 2D3D, 2D3E, 2D3F, 2D3G, 2D3H, 2D3I.

Last update: 05.02.15

4.5.1.1 Method

Our general model is a simplified version of the detailed methodology described in chapter 6 of the EMEP/CORINAIR Guidebook 2007 (EEA 2007). It represents a mass balance *per substance*, where emissions are calculated by multiplying relevant activity data with an emission factor. For better coverage, point sources reported from industries to the Norwegian Environment Agency and calculated emissions from a side model for cosmetics, are added to the estimates. For a detailed description of method and activity data, see Holmengen and Kittilsen (2009).

It is assumed that all products are used the same year as they are registered, and substances are not assumed to accumulate in long-lived products. In other words, it is assumed that all emissions generated by the use of a given product during its lifetime take place in the same year

as the product is declared to our data source, the Norwegian Product Register. In sum, this leads to emission estimates that do not fully reflect the actual emissions taking place in a given year. Emissions that in real life are spread out over several years all appear in the emission estimate for the year of registration. However, this systematic overestimation for a given year probably more or less compensates for emissions due to previously accumulated amounts not being included in the estimate figures.

No official definition of solvents exists, and a list of substances to be included in the inventory on NMVOC emissions was thus created. The substance list used in the Swedish NMVOC inventory (Skårman et al. 2006) was used as a basis. This substance list is based on the definition stated in the UNECE Guidelines¹². The list is supplemented by NMVOC reported in the UK's National Atmospheric Emissions Inventory (NAEI) (AEA 2007). The resulting list comprises 678 substances. Of these, 355 were found in the Norwegian Product Register for one or more years in the period 2005-2007.

Cosmetics

Cosmetics are not subject to the duty of declaration. The side model is based on a study in 2004, when the Norwegian Environment Agency calculated the consumption of pharmaceuticals and cosmetics (SFT 2005). The consumption was calculated for product groups such as shaving products, hair dye, body lotions and antiperspirants. The consumption in tonnes each year is calculated by using the relationship between consumption in Norwegian kroner and in tonnes in 2004. Figures on VOC content and emission factors for each product group were taken for the most part from a study in the Netherlands (IVAM 2005), with some supplements from the previous Norwegian solvent balance (the previous NMVOC emission model).

4.5.1.2 Activity data

The data source is the Norwegian Product Register. Any person placing dangerous chemicals on the Norwegian market for professional or private use has a duty of declaration to the Product Register, and import, export and manufacturing is reported annually. The only exception is when the amount of a given product placed on the market by a given importer/producer is less than 100 kg per year.

The information in the data from the Product Register makes it possible to analyse the activity data on a substance level, distributed over product types (given in UCN codes; (Norwegian Product Register 2007)), industrial sectors (following standard industrial classification (NACE), including private households (no NACE), or a combination of both. As a consequence, the identification of specific substances, products or industrial sectors that have a major influence on the emissions is greatly facilitated.

¹² "Volatile compound (VOC) shall mean any organic compound having at 293.15 degrees K a vapor pressure of 0.01 kPa or more, or having a corresponding volatility under the particular conditions of use."

Cosmetics

The side model for cosmetics is updated each year with data on from the Norwegian Association of Cosmetics, Toiletries and Fragrance Suppliers (KLF).

Point sources

Data from nine point sources provided by the Norwegian Environment Agency are added to the emissions estimates. The point sources are reported from the industrial sector “Manufacture of chemicals and chemical products” (NACE 20). In order to avoid double counting, NMVOC used as raw materials in this sector are excluded from the emission estimates from the Product Register data.

4.5.1.3 Emission factors

Emission factors are specific for combinations of product type and industrial sector. Emission factors from the Swedish model for estimating NMVOC emissions from solvent and other product use (Skårman et al. 2006) are used. The emission factors take into account different application techniques, abating measures and alternative pathways of release (e.g. waste or water). These country-specific emission factors apply to 12 different industries or activities that correspond to sub-divisions of the four major emission source categories for solvents used in international reporting of air pollution (EEA 2007).

It is assumed that the factors developed for Sweden are representative for Norwegian conditions, as we at present have no reason to believe that product types, patterns of use or abatement measures differ significantly between the two countries. Some adjustments in the Swedish emission factors were made when the model was first developed (see Holmengen and Kittilsen (2009)) and several improvements of single emission factors have been made in the following years.

In accordance with the Swedish model, emission factors were set to zero for a few products that are assumed to be completely converted through combustion processes, such as EP-additives, soldering agents and welding auxiliaries. Quantities that have not been registered to industrial sector or product type are given emission factor 0.95 (maximum). Emission factors may change over time, and such changes may be included in this model. However, all emission factors are at the moment constant for all years.

4.5.1.4 UncertaintiesUncertainty in emission factors

The emission factors are more detailed in the new NMVOC model than in the previous model, as this model can take into account that emissions are different in different sectors and products, even when the substance is the same. However, for this to be correct, a thorough evaluation of each area of use is desirable, but not possible within a limited time frame. Thus, the emission factor is set with general evaluations, which leads to uncertainty.

The emission factors are taken from several different sources, with different level of accuracy. The uncertainties in emission factors depend on how detailed assessment has been undertaken when the emission factor was established. Some emission factors are assumed to be unbiased,

while others are set close to the expected maximum of the range of probable emission factors. This, together with the fact that the parameter range is limited, gives us a non-symmetrical confidence interval around some of the emission factors. For each emission factor we thus have two uncertainties; one negative (n) and one positive (p). These are aggregated separately, and the aggregated uncertainty is thus not necessarily symmetrical.

Uncertainty in activity data

For the activity data, the simplified declarations and the negative figures due to exports lead to known overestimations, for which the uncertainty to a large extent is known. A more elaborate problem in calculations of uncertainty is estimating the level of omissions in declaration for products where the duty of declaration does apply. In addition, while declarations with large, incorrect consumption figures are routinely identified during the QA/QC procedure, faulty declarations with small consumption figures will only occasionally be discovered. There is however no reason to believe that the Product Register data are more uncertain than the data source used in the previous model (statistics on production and external trade), as similar QA/QC routines are used for these statistics.

The errors in activity data are not directly quantifiable. Any under-coverage in the Product Register is not taken into account. Skårman et al. (2006) found that the activity data from the Swedish Product register had an uncertainty of about 15 per cent. The Norwegian Product Register is assumed to be comparable to the Swedish, and thus the uncertainty in the activity data is assumed to be 15 per cent. For some products, simplified declarations give an indication of maximum and minimum possible amounts. In these cases, the maximum amount is used, and the positive uncertainty is set to 15 per cent as for other activity data, while the negative uncertainty is assumed to be the interval between maximum and minimum amount. All activity data are set to zero if negative.

For a detailed description of the uncertainty analysis, see Holmengen and Kittilsen (2009). The variance of total emission was estimated from the variance estimates obtained for emission factors and activity data, using standard formulas for the variance of a sum and the variance of a product of independent random variables. The aggregated uncertainties in level and trend are given in Table 4.21 and Table 4.22.

Table 4.21 Uncertainty estimates for level of NMVOC emissions, 2005-2007. Tonnes and per cent

Uncertainty in level	Negative (n)	Negative (n) (per cent of total emissions)	Positive (p)	Positive (p) (per cent of total emissions)
2005	2 288	4.58	1 437	2.88
2006	1 651	3.70	1 103	2.47
2007	1 299	2.79	1 168	2.51

Source: Holmengen and Kittilsen (2009)

Table 4.22. Uncertainty estimates for trend in NMVOC emissions, 2005-2007. Tonnes

Uncertainty in trend	Negative (n)	Positive (p)	95% confidence interval for change
2005-2006	2 135	1 067	(-7 366, -4 164)
2006-2007	1 420	947	(407, 2 774)
2005-2007	1 882	1 076	(-5 286, -2 328)

Source: Holmengen and Kittilsen (2009)

4.5.1.5 Completeness

No major missing emission sources are likely.

4.5.1.6 Source specific QA/QC

Large between-year discrepancies in the time series of substance quantities are routinely identified and investigated, in order to correct errors in consumption figures. Large within-year discrepancies between minimum and maximum quantities in simplified declarations are routinely identified and investigated, in order to prevent overestimation for substances where consumption figures are given in intervals. Large within-year discrepancies between totals for industrial sectors (NACE) and totals for products (UCN) are routinely identified and investigated, in order to detect erroneous or incomplete industrial sectoral and product type distribution.

4.5.2 Creosote-treated materials

NFR 2D3G

Last update: 02.01.17

4.5.2.1 Description

Creosote is mainly used in quay materials and conduction poles, but also in fence poles and roof boards. In Norway there is a requirement that all creosote in use should contain less than 50 mg/kg benzo(a)pyrene (Ministry of the Environment 2004). PAH-components will evaporate from the creosote-treated materials in hot weather. In addition, PAH-components will evaporate during impregnation. The smallest PAH-components, like naphthalene, are most volatile, but several components used in wood treatment will not evaporate.

4.5.2.2 Method

Emissions of PAHs benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene are calculated based on the import of creosote oil and emission factors. For simplicity, it is assumed that all PAH is emitted the same year as the materials are produced.

Activity data

Data on imported amounts of creosote oil are taken from Statistics Norway's statistics on external trade.

4.5.2.3 Emission factors

The emission factor used is taken from Finstad et al. (2001). It is assumed that imported creosote oil contains on average 55 per cent PAH and that one per cent will evaporate during the lifetime of the creosote-treated materials. It is assumed that PAH-4 account for 0,018 per cent of the total PAH emissions. Emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene

and indeno(1,2,3-cd)pyrene are estimated using the a PAH-profile for creosot oil, Finstad et al (2001), see Table 4.23.

Table 4.23. Distribution of PAH-4 emissions from creosote oil. Share of PAH-4

Component	Distribution of PAH emissions (ratio)
benzo(a)pyrene	..
benzo(b)fluoranthene	0.50
benzo(k)fluoranthene	0.50
indeno(1,2,3-cd)pyrene	..

Source: Finstad et al (2001)

4.5.2.4 Uncertainties

In the inventory it is assumed that all PAH is emitted the same year as the materials are used. This is however not the case, since PAH will be emitted as long as the creosote-treated materials are in use. However, most of it is likely to be emitted during the first years.

4.5.2.5 Completeness

No major missing emission components or sources are likely.

4.5.2.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.3 Road paving with asphalt

NFR 2D3B

Last update: 02.01.17

4.5.3.1 Method

The emissions from road paving are being calculated in accordance with a Tier 1 approach (EEA 2013) for NMVOC, TSP, PM₁₀ and PM_{2.5}. Emissions of dioxins from production of asphalt are also included.

$$E_{\text{pollutant}} = AR_{\text{production}} * EF_{\text{pollutant}}$$

Where:

$E_{\text{pollutant}}$ = the emission of the specified pollutant

$AR_{\text{production}}$ = the activity rate for the road paving with asphalt

$EF_{\text{pollutant}}$ = the emission factor for this pollutant

Dioxins

Asphalt preparations and asphalt recycling are supposed to be a possible dioxin source, especially in countries using extensive recycling, and that use salt on the roads during winter. A lot of salt is used on Norwegian roads during winter, and when this asphalt is heated during recycling, it is assumed to give emissions of dioxins (Hansen 2000).

4.5.3.2 Activity data

The activity data used is the annual weight of asphalt used for road paving in Norway. EBA, *pers. comm*¹³).

4.5.3.3 Emission factors

Emissions of NMVOC, TSP, PM₁₀ and PM_{2.5} from road paving with asphalt are estimated using Tier 1 emission factors from the 2016 EEA Guidebook.

Table 4.24. Emission factor for road paving with asphalt. g/tonn

NMVOC	16
TSP	14 000
PM ₁₀	3 000
PM _{2.5}	400

Source: EEA (2016)

Dioxins

Two emission factors are found in the literature. According to SFT (2001), the Oslo and Paris Convention (OSPAR) suggests an emission factor of 0.047 µg/tonne asphalt. This emission factor is however assumed to be very high since it is based on data from a plant only re-circulating old asphalt. Fyns Amt (2000) operates with a much lower emission factor, which probably reflects dioxin emissions from preparation of new asphalt. Since Norway both makes new asphalt and recycles old asphalt, it is assumed that an emission factor in between those suggested from OSPAR and Fyns Amt would be most correct for Norwegian conditions (Table 4.25).

Table 4.25. Dioxin emission factor for asphalt production. µg I-TEQ/tonne produced asphalt

Source	Emission factor
SFT (2001)	0.047
Fyns Amt (2000)	0.0022
Emission factor chosen	0.025

4.5.3.4 Uncertainties

The activity data used are uncertain. The emission factors used are also uncertain. The annual emissions are low however, and will not have any impact on the total level of these emissions.

4.5.3.5 Completeness

No major missing emission components are likely.

4.5.3.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.4 Other product use

NFR 2G

¹³ EBA (2014): Expert judgement by Contractors Association - Building and Construction (EBA), Oslo, Norway

Last update: 05.02.15

4.5.5 Mercury-containing products

NFR 2G

Last update: 05.02.15

4.5.5.1 Method

Breakage of mercury-containing thermometers, fluorescent tubes, economy bulbs, and various measuring and analytical instruments lead to emissions of mercury. The emission estimates are based on an annual report from the Norwegian Environment Agency ("Miljøgifter i produkter"). The sale of mercury-containing thermometers and fluorescent tubes has decreased strongly since the mid-1990s, and the mercury content in these products has been reduced. A prohibition against the production, import and export of mercury-containing products entered into force in 1998, except for some thermometers for professional use, which were prohibited in 2001. Since these products have long operating life times, there will be emissions from these products for many years. In the calculations, however, it is assumed that the emissions occur the same year as the product is sold.

For thermometers, it is assumed that all mercury is emitted in hospitals, despite some breakage of mercury-containing thermometers that occur in households. For fluorescent tubes and economy bulbs, all emissions are placed in households, although emissions occur in all sectors. For measuring and analytical instruments, all emissions are placed under research and development work.

4.5.5.2 Uncertainties

The emissions are assumed to be emitted the same year as the products are sold. This is not accurate, since most of these products have long operating life times. It is however impossible to predict the annual breakage and the mercury content in each of them.

4.5.5.3 Completeness

No major missing emission components are likely.

4.5.5.4 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.5.6 Tobacco

NFR 2G

Last update: 10.01.17

4.5.6.1 Method

NO_x, NMVOC, CO, NH₃, particles, BC, heavy metals and POPs

The emission components included from the combustion of tobacco are NO_x, NMVOC, CO, NH₃, particles, heavy metals and POPs (Persistent organic pollutants). Emission figures have been calculated by multiplying the annual consumption of tobacco with emission factors for each

pollutant. As recommended by the UNFCCC ERT (expert review team), emissions from use of tobacco are reported under NFR 2G.

4.5.6.2 Activity data

The total consumption of tobacco in Norway is given by the net import of tobacco from Statistics Norway's external trade statistics.

Emission factors

Table 4.26 gives emission factors used for tobacco combustion. For NO_x, NMVOC and CO the emission factors are calculated by Statistics Norway, based on values given in Norwegian Directorate of Health (1990).

Table 4.26. Emission factors used for tobacco combustion

	Tobacco (unit/kg tobacco)	Source
NO _x (kg)	0.0034652	Statistics Norway, Norwegian Directorate of Health (1990)
NMVOC (kg)	0.0048374	Statistics Norway, Norwegian Directorate of Health (1990)
CO (kg)	0.1215475	Statistics Norway, Norwegian Directorate of Health (1990)
NH ₃ (kg)	0.00415	EEA (2013)
TSP (kg)	0.04	Institute of environmental and energy technology (2002)
PM ₁₀ (kg)	0.04	Institute of environmental and energy technology (2002)
PM _{2.5} (kg)	0.04	Institute of environmental and energy technology (2002)
BC	0.5% of PM _{2.5}	IIASA (Kupiainen & Klimont 2004)
Pb (g)	0.00005	Finstad et al. (2001)
Cd (g)	0.0001	Finstad et al. (2001)
Hg (g)	0.00001	Finstad et al. (2001)
As (g)	0.000159	Finstad and Rypdal (2003)
Cr (g)	0.000125	Finstad and Rypdal (2003)
Cu (g)	0.000354	Finstad and Rypdal (2003)
Benzo(a)pyrene (g)	0.000111	Finstad et al. (2001)
Benzo(b)fluoranthene (g)	0.000045	EEA (2016)
Indeno(1,2,3-cd)pyrene (g)	0.000045	EEA (2016)
Indeno(1,2,3-cd)pyrene (g)	0.000045	EEA (2016)
Dioxins (µg)	0.0013	Finstad et al. (2002b)

4.5.6.3 Uncertainties

The emissions are assumed to be emitted the same year as the products are imported.

4.5.6.4 Completeness

Tobacco bought tax free abroad and tobacco smuggled are not included in the inventory.

4.5.6.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.6 Other production

NFR 2H

Within other production, Norway includes emissions from pulp and paper, food and beverages industry and ore mines.

Table 4.27. Other production. Components emitted and included in the Norwegian inventory.

Other production	SO ₂	NO _x	NH ₃	NMVOC	CO	Particles	BC	Heavy metals	Dioxins
Pulp and paper	R	E	NE	E	E	R	E	NA	NA
Food and beverages industry	NA	NA	NA	E	NA	NE	NE	NA	NA
Ore mines	R	NA	NA	NA	NA	R	NA	NA	R

E = Figures estimated by Statistics Norway (Activity data * emission factor). R = Figures reported by the plant to the Norwegian Environment Agency. NA = Not Applicable.

Source: Statistics Norway/ Norwegian Environment Agency

4.6.1 Pulp and paper

NFR 2H1

Last update: 13.01.17

4.6.1.1 Description

Pulp and paper production has three major processing steps; pulping, bleaching and paper production. Kraft (sulphate) pulping is the most widely used pulping process and is generally used to produce strong paper products. The Kraft pulping process includes bleaching, chemical recovery and by-products recovery. The sulphite pulping is another chemical pulping process. It produces a weaker paper than some other types of pulping, but the pulp is less coloured, making it more suitable for printing, often with little bleaching. In Norway, SO₂ and particles are reported emitted from production of pulp and paper. In the Kraft pulping process, sodium sulphide and sodium hydroxide are used to chemically dissolve the lignin that binds the cellulose fibres, and in the acid sulphite pulping process, sulphurous acid solution is used. SO₂ is emitted in these processes. Emissions of NO_x, NMVOC and CO are estimated.

4.6.1.2 Method

SO₂

Emission figures are reported from producers of chemical pulp to the Norwegian Environment Agency. SO₂ is measured continuously and emission estimates are made from these measurements.

Particles

Four plants producing pulp and paper, report non-combustion emissions of particles to the Norwegian Environment Agency. Two of these plants have not reported emission figures from combustion and it is assumed that the reported non-combustion emission figures include emissions from combustion. It varies when the plants started reporting emission figures for particles, and due to lack of data, emission for those years is assumed to be the same as in the first year of reporting.

Two of the plants state that they clean the emissions by electric filter and wet scrubbers, and it is assumed by the Norwegian Environment Agency that the particles emitted are smaller than PM_{2.5}. The other two clean their emissions using only wet scrubbers, and it is assumed the

particles are smaller than PM₁₀. According to TNO (Institute of environmental and energy technology 2002), PM_{2.5} is 20 per cent of PM₁₀ and PM₁₀ is the same as TSP.

NO_x, NMVOC and CO

Emissions of NO_x, NMVOC and CO are estimated based on annual production levels and emission factors from the 2013 Guidebook.

BC

BC emissions have been estimated using shares of PM_{2.5} as emission factors. Shares given by IIASA (Kupiainen & Klimont 2004) have been used.

4.6.1.3 Activity data

For the estimates of NO_x, NMVOC and CO, the plants reported production levels of pulp by different processing steps are used.

4.6.1.4 Emission factors

For the estimates of NO_x, NMVOC and CO, emission factors as shown in Table 4.28 from the 2016 Guidebook are used.

Table 4.28. Emission factors for pulp and paper. kg/Mg air dried pulp

NO _x	1 (Kraft), 2 (Acid sulphite)
NMVOC	2 (Kraft), 0.2 (Acid sulphite), 0.05 (neutral sulphite semi)
CO	5.5 (Kraft)

Source: EEA (2016)

4.6.1.5 Uncertainties

The particle size distribution used is not plant specific and might therefore be different from the one suggested by TNO.

4.6.1.6 Completeness

Major missing emission components are not likely.

4.6.1.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.6.2 Food and beverages industry

NFR 2H2

Last update: 10.01.17

4.6.2.1 Description

This source category includes NMVOC emissions from production of bread and beer. Emissions of NMVOC from spirit manufacture are considered insignificant and are not included in the inventory.

4.6.2.2 Method

NMVOC

Production of bread and beer (and other similar yeast products) involves fermentation processes that lead to emission of NMVOC (ethanol). Emissions are calculated based on production volumes and emission factors.

4.6.2.3 Activity data

Production volumes of bread and beverages are annually reported to Statistics Norway.

4.6.2.4 Emission factors

The emission factors are taken from EEA (1996).

Table 4.29. NMVOC emission factors from production of bread and beverage

	Emission factor	Unit
Production of bread	0.0045	tonnes/tonnes produced
Production of beverage	0.35	kg/1000 litres

Source: EEA (1996)

4.6.2.5 Uncertainties

The emission factors used are recommended by EEA (1996) and are not specific for Norwegian conditions.

4.6.2.6 Completeness

Major missing emission components are not likely.

4.6.2.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.6.3 Ore mines

NFR 2H3

Last update: 05.02.15

4.6.3.1 Description

Three ore mines are included in the Norwegian Inventory, but one of the mines was closed down in 1996. Emission figures of SO₂, particles and dioxins are included. The treatment of ore generates emissions of SO₂, and particles are also emitted. Dioxin emissions are due to the thermal process during the pellet production. The ore mine which closed down in 1996, had large dioxin emissions due to the thermal process during the pellet production.

4.6.3.2 Method

SO₂

The ore mine which was closed down in 1996, reported emission figures for SO₂ to the Norwegian Environment Agency. None of the two other ore mines report any non-combustion SO₂ emissions.

Particles

All the three ore mines report emission figures for particles to the Norwegian Environment Agency. Emissions for the two existing ore mines are reported from respectively 1994 and 1996 and it is assumed by Statistics Norway, in accordance with the Norwegian Environment Agency, that emissions for previous years have been in the same order of size.

The size distribution used in the Norwegian inventory is according to TNO (Institute of environmental and energy technology 2002) (Table 4.30).

Table 4.30. Particle size distribution for particles emitted from ore mining. Ratio X^1 /TSP

Component	Particle size distribution (ratio)
TSP	1
PM ₁₀	0.49
PM _{2.5}	0.07

¹ X is either PM_{2.5}, PM₁₀ or TSP.

Source: TNO (Institute of environmental and energy technology 2002).

Dioxins

Emissions of dioxins are registered only for the ore mine which was closed down in 1996. Emission figures were first reported to the Norwegian Environment Agency in 1994 and emissions for previous years have been assumed by Statistics Norway, in accordance with the Norwegian Environment Agency, to be in the same order of size as the reported figure in 1994.

4.6.3.3 Uncertainties

For years where reported emission figures do not exist for particles and dioxins, Statistics Norway has assumed, in accordance with the Norwegian Environment Agency, that the emissions are in the same order of size as for the first year of reporting. This is uncertain and a result of lack of better data. The size of the particles emitted from ore mining will also depend on the type of ore and production process. The particle size distribution used in the inventory does not consider these differences.

4.6.3.4 Completeness

SO₂ emissions are only included in the inventory for the ore mine that was closed down in 1996. The SO₂ emissions from the two other ore mines are not included in the inventory.

4.6.3.5 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

4.7 Wood processing

NFR 2I

Table 4.31. Wood production. Components emitted and included in the Norwegian inventory

Other production	SO ₂	NO _x	NH ₃	NMVOC	CO	Particles	TSP
Wood production	NE	NE	NE	NE	NE	NE	E

E = Figures estimated by Norwegian Environment Agency (Activity data * emission factor). NE = Not Estimated

Source: Statistics Norway/ Norwegian Environment Agency

4.7.1 Wood processing

NFR 2I

Last update: 13.01.17

4.7.1.1 Description

This source category includes TSP emissions from four plants from wood processing.

4.7.1.2 Method

The emissions are calculated based on production volumes and emission factors.

4.7.1.3 Activity data

The production volumes of wood processing products are annually reported to the Norwegian Environment Agency.

4.7.1.4 Emission factors

The emission factor is taken from 2016 EEA Guidebook.

Table 4.32. TSP emission factor for wood processing

	Emission factor	Unit
Wood processing	1	Kg/Mg wood product

Source: EEA (2016)

4.7.1.5 Uncertainties

The emission factor is not specific for Norwegian conditions.

4.7.1.6 Completeness

Major missing emission components are not likely.

4.7.1.7 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

5 AGRICULTURE (NFR sector 3)

NFR 3

180

5.1 Overview

Agriculture is an important contributor to NH₃ emissions. Animal manure management, grazing animals and the use of fertiliser (manure, synthetic fertiliser, sewage sludge and other organic fertilisers applied to soils) generate emissions of NH₃. Another source of NH₃ is treatment of straw using NH₃ as a chemical.

Animal manure management and the use of fertiliser (manure, synthetic fertiliser, sewage sludge and other organic fertilisers applied to soils) also generates emissions of NO_x.

Emissions of NMVOC from manure management and from cultivated agricultural soils are also included in the inventory.

Non-combustion emissions of particles from manure management and agricultural soils are also calculated. Additionally, there are long-range transboundary air emissions arising from the burning of agricultural residues.

Table 5.1. Agriculture emissions as per cent of total emissions of the component, 2015

Pollutant	Percent of emissions
NO _x	4.8 %
NMVOC	6.9 %
SO _x	0.1 %
NH ₃	92.1 %
PM _{2.5}	2.8 %
PM ₁₀	7.9 %
TSP	6.9 %
BC	1.0 %
CO	0.8 %
Pb	0.1 %
Cd	8.8 %
Hg	2.7 %
As	0.0 %
Cr	0.1 %
Cu	0.0 %
Diox	0.1 %
Benzo(a) pyrene	1.4 %
Benzo(b) fluoranthene	1.7 %
Benzo(k) fluoranthene	1.5 %
Indeno (1,2,3-cd) pyrene	1.7 %
PAH-4	1.6 %
HCB	0.0 %
PCB	0.0 %

Source: Statistics Norway/ Norwegian Environment Agency

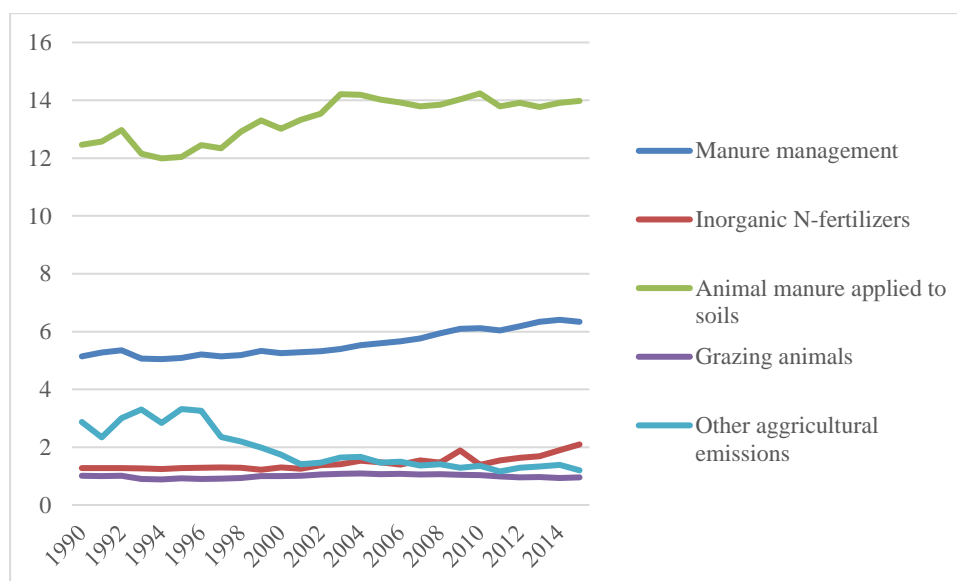


Figure 5.1. Trends for the NH_3 emissions for agricultural sources. 1000 tonnes NH_3 . 1990-2015

Source: Statistics Norway/ Norwegian Environment Agency

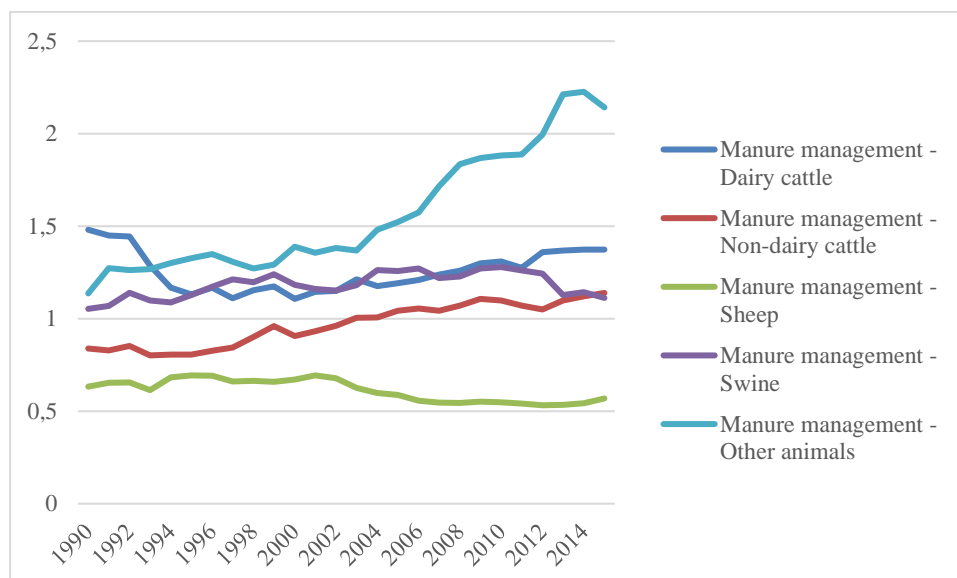


Figure 5.2: Trends for the NH_3 emissions from manure management. 1000 tonnes NH_3 . 1990-2015

Source: Statistics Norway/ Norwegian Environment Agency

The total emissions of NH_3 from agriculture have been relatively stable but with a slight increase since 1990. Figure 5.1 and Figure 5.2 shows the NH_3 trends for the different agriculture sources.

Combustion of straw is a source to many different emission compounds, which all have the same decreasing trend, primarily due to reduced amount of straw burned since 1990.

More information is given in the trend chapter, section 2.

5.2 Livestock population characterisation

Last update: 01.03.2017

5.2.1 Data sources

The main sources of the livestock statistics are the register of production subsidies (sheep for breeding, goats, breeding pigs, poultry for egg production and beef cows), statistics of approved carcasses (animals for slaughter) and the Cow Recording System at TINE BA¹⁴ (heifers for breeding and dairy cows). These sources cover 90-100 per cent of the animal populations. The coverage in the register of production subsidies is shown in Table 5.2.

Table 5.2. Estimated coverage of animal populations in the and the Cow Recording System at TINE BA . 2015 register of production subsidies

	Percentage covered in the statistics
Dairy cows ¹	98,6
Heifers for breeding ¹	98,6
Beef cows	100
Sheep	99.7
Goats	100
Laying hens	100
Chicks for breeding	96.1
Other poultry for breeding	99.9
Sows	98.4
Young pigs for breeding	100
Deer	100

¹ Share of livestock herds.

Source: Estimations by Statistics Norway and the Cow Recording System (dairy cows and heifers).

The statistics of approved carcasses covers close to 100 per cent of all slaughtered animals. Home slaughter is not included, but the extent of home slaughter is very low due to legal restrictions. Even animals consumed by producers are in most cases registered at the slaughterhouses. The number of dairy cows and heifers for breeding derive from the Cow Recording Systems. Between 98 and 99 per cent of all dairy cows are assumed to be registered here.

The registers are updated annually. In addition to the animals included in these registers, an estimate of the number of horses that are not used in farming is obtained from the Norwegian Institute of Bioeconomy Research (NIBIO)¹⁵. The number of reindeer is obtained from the Norwegian Reindeer Husbandry Administration.

For the categories of animals living shorter than a year or two, generally animals for slaughter, lifetime is taken into account to get a yearly average for the number of animals.

¹⁴ TINE BA is the sales and marketing organisation for Norway's dairy cooperative and covers most of the milk production and the meat production induced by milk production.

¹⁵ Former named the Agricultural Economics Research Institute (NILF).

5.2.2 Method for estimating number of cattle

For dairy cows, additional information from the Cow Recording System concerning annual milk production and proportion of concentrate in the diet is used (TINE BA Annually). The Cow Recording System also supplies annual information about slaughter age for heifers and bulls and data for estimating live weight of dairy cows and heifers for breeding, and also the age of young cows at their first calving. (Moen, *pers. comm.*¹⁶).

For heifers and bulls for slaughter, animal numbers are based on data from statistics of approved carcasses which provide data on numbers slaughtered and slaughter weights. Combined with slaughter age from the Cow Recording System (TINE BA Annually), this gives a precise estimation of animal life time for each animal slaughtered. One principal draw-back of this method for estimating animal population is that emissions in all stages of these animals' lives will be accounted for in the year of slaughter, even though the emissions in the early stages of the lives of these animals to a large extent took place in the previous year. In a stable population of animals, this error is automatically adjusted for. Since animal populations are relatively stable, this error is considered much smaller compared to errors related to estimating animal year based on animal populations in the register of production subsidies which was previously used. The data sources used also ensure a better coherence between animal numbers, life time and weight. Estimated animal years for cattle are given in Table 5.3.

The number of milk cows calving their first time (=heifers for replacement) and their average age at time of calving is reported by the Cow Recording System (TINE BA Annually) on request from Statistics Norway. These data date back to 2004. For the years 1990-2003, average fraction (number of heifers)/(number of milk cows) for the years 2004-2011 is used to estimate number of heifers based on number of milk cows. Number of heifers for replacement in beef production is collected from annual reports from Animalia (Norwegian Meat and Poultry Research Center (www.animalia.no)). Figures exist from 2007. For previous years, the number is estimated with the same method as for heifers for milk production.

¹⁶ Moen, O. (*annually*): Personal information, email from Oddvar Moen Tine Rådgivning annually.

Table 5.3 Estimated animal years for cattle

	Heifers for replacement	Heifers for slaughter	Bulls for slaughter	Beef cows ¹	Dairy cows
1990	311 279	47 020	289 945	8 193	325 896
1995	299 284	47 103	284 237	20 334	310 346
2000	280 121	63 512	285 349	42 324	284 880
2005	255 862	57 619	263 170	54 841	255 663
2006	246 711	58 446	255 963	55 706	250 903
2007	235 282	56 607	247 578	57 609	246 624
2008	240 399	54 831	238 111	60 401	238 550
2009	247 902	53 397	235 689	63 803	235 480
2010	239 839	53 410	230 872	67 110	232 294
2011	239 007	48 778	223 536	68 539	224 721
2012	235 891	42 863	217 050	71 834	229 767
2013	239 386	47 294	220 401	70 969	225 163
2014	244 601	67 624	208 979	73 894	221 032
2015	238 485	63 815	205 073	77 408	217 576

¹ Counted animals

Source: Cow Recording System at TINE BA (dairy cows), slaughter statistics and estimations by Statistics Norway

5.2.3 Method for estimating number of sheep

In the estimations, the sheep population is divided between sheep > one year and sheep < one year. Data from both the register of production subsidies and slaughter statistics is used in estimating the number of animals.

Sheep over one year is estimated as the number of sheep registered 1. of January deducted for the number of sheep slaughtered Jan.-May. The sheep slaughtered later in the year are counted as living the whole year.

Sheep under one year is estimated as number of sheep under one year registered 1. of January + number of lambs slaughtered June-December *143/365. Lambs slaughtered before June are assumed to be registered as sheep under one year the 1. of January. Practically all lambs slaughtered after June are born in the spring. An expert judgment suggests an average lifetime of 143 days for slaughtered lambs born in the spring (UMB, *pers. comm*¹⁷).

5.2.4 Deviations from FAO statistics

There are some differences between the number of animals used in these calculations and the FAO statistics. The general reason that animal statistics used in the emission inventory differ from the statistics delivered to FAO is that the statistics are used for different purposes. Animal

¹⁷ UMB (2001): Expert judgement by Department of Animal Science, Ås: Norwegian University of Life Sciences.

statistics used in the inventory has to be categorized so that the categories fit the recommended methodology and the various emission factors used in the emission estimations. The figures reported to the FAO are provided by the Norwegian Institute of Bioeconomy Research (NIBIO)¹⁸. NIBIO makes an overall estimation for the agricultural sector, which is the basis for the annual negotiations for the economic support to the sector. This estimate includes a grouping of all agricultural activities, comprising area, number of animals and production data. Differences include:

- Different emphasis on the dates for counting, 31.07 and 31.12
- NIBIO does not register pigs under 8 weeks, whilst Statistics Norway does. For the number of animals for slaughter, SN uses the statistics of approved carcasses and estimates animal years (average population through the year) on basis of this, while NIBIO uses figures for registered animals at specific dates.
- For the number of dairy cows and heifers for replacement, Statistics Norway uses statistics from the Cow Recording System (TINE BA Annually)

5.2.5 Uncertainties

Activity data

The uncertainty in the data is considered to be within ± 5 per cent. There is also an uncertainty related to the fact that some animals are only alive part of the year and how long this part is.

5.2.6 Source specific QA/QC

In 2001, a project was initiated to improve the estimate of the number of animals. This was completed in 2002. In 2012, a further revision of the numbers of bulls and heifers was implemented. In 2016, the method for estimating the number of sheep was revised. The revised data on animal populations form the basis for the emission calculations for all years.

5.3 Nitrogen in animal manure as basis for emission estimates

Access to nitrogen is vital for all plant growth; hence nitrogen is added to the soil from i.a. animal manure. This causes emissions to air of compounds containing nitrogen at various points. Of the nitrogen compounds emitted to air from animal manure, N_2O , NO_x and NH_3 are estimated.

According to the IPPC and LRTAP guidelines, process emissions of nitrogen compounds from use of animal manure are calculated from the following sources:

1. Manure management systems (N_2O , NO_x and NH_3)
2. Application of manure on soil (N_2O , NO_x and NH_3)
3. Droppings from animals on pastures (N_2O and NH_3)
4. Leakage of nitrogen through manure management systems and soils (N_2O)

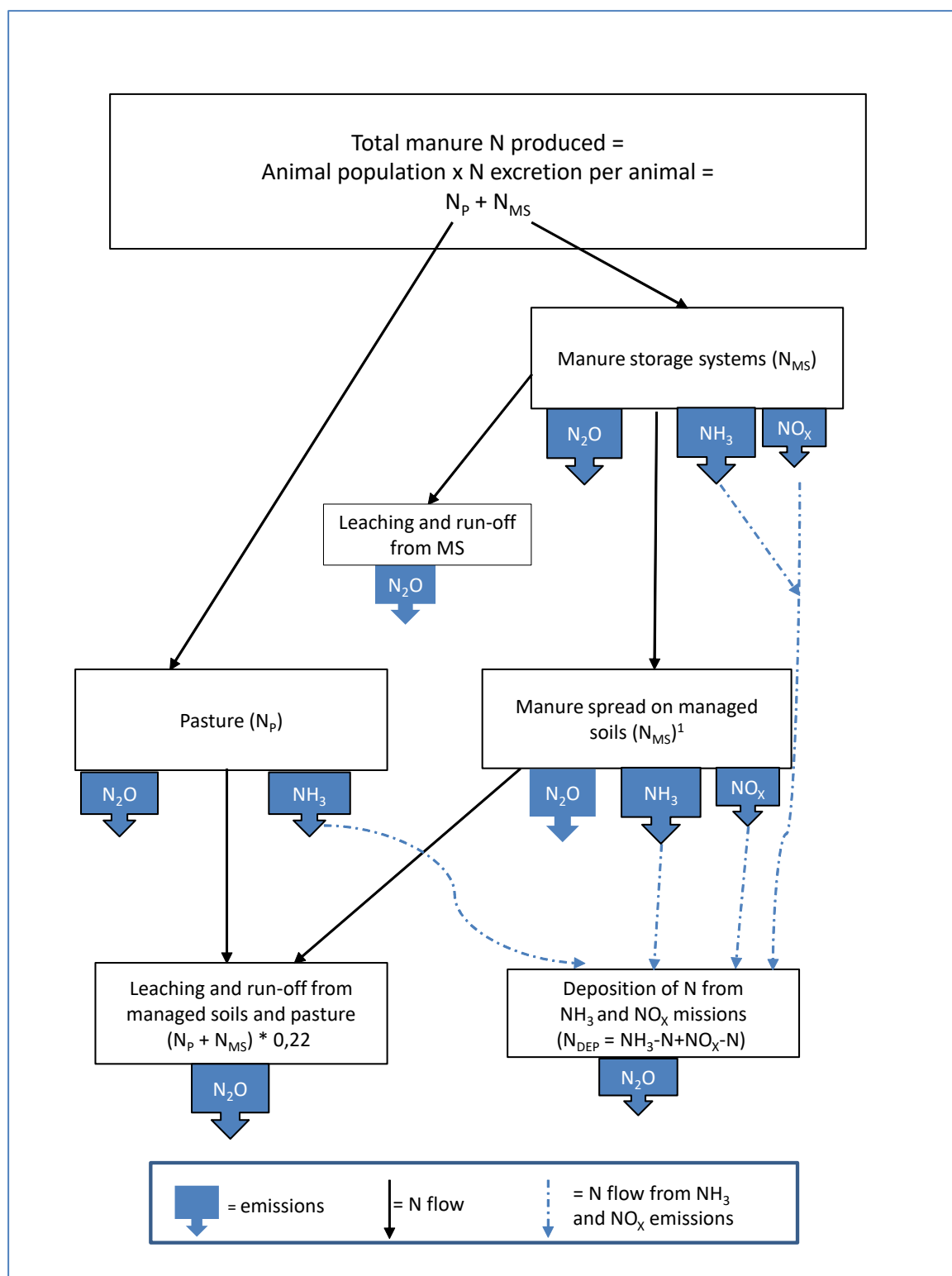
¹⁸ Former named the Agricultural Economics Research Institute (NILF).

5. Deposition of nitrogen from emissions of NH_3 and NO_x (N_2O)

Though the nitrogen flow is continuously depending on its surroundings (soil characteristics, temperature, moisture etc.) and the preceding supplies and losses of N, the emission estimates of each of the sources above are generally done independently of emissions from the other sources. Figure 5.3 gives an overview of the manure nitrogen flows in the Norwegian greenhouse gas inventory.

The following decides the amounts of N that are used as the basis for the respective emission calculations:

- The amount of N in manure systems is calculated as total N in manure adjusted for the N that is dropped on pastures.
- N_2O emitted during spreading is calculated from the amounts of N in manure storage. This means that N lost through leaching in manure storage and as N_2O , NO_x and NH_3 in manure storage and during spreading is not deducted.
- NH_3 emitted during and after spreading of manure is based on the amounts of N in manure storage minus N lost as NH_3 volatilization in manure storage. NO_x emitted during and after spreading of manure is based on the amounts of N in manure storage, and *not* deducting N lost as NH_3 or NO_x volatilization in manure storage. Losses of N through leaching and N_2O emissions in manure storage are not deducted.
- Emissions of N_2O and NH_3 from pasture are calculated independently of each other, and are based on the amounts of N estimated in manure dropped during grazing. NO_x emissions from pastures is not estimated.
- N_2O lost through leaching is based on total N in manure storage and N dropped on pastures. This means that N emitted as N_2O , NO_x and NH_3 or lost in other ways is not deducted.
- The nitrogen in NH_3 and NO_x volatilised both during storage, pasture and spreading of manure is the basis for the calculation of N_2O emissions from atmospheric deposition. How the amounts of N are estimated in the various emission estimates, is described in more details in the respective chapters below.



¹ N_{MS} is the N basis for the N_2O and NO_x emission estimations, while ($N_{MS} - N$ lost as NH_3 in manure storage systems) is the N basis for the NH_3 emission estimations.

Figure 5.3 Overview of the manure nitrogen flows in the Norwegian greenhouse gas inventory

5.4 Emissions from manure management

NFR 3B

Last update:

5.4.1 Description

Manure management in Norway is a source of emissions to air of NH₃, NO_x, NMVOC and PM.

5.4.2 NH₃ emissions from manure management

5.4.2.1 Description

The dominating pollutant emitted from this source category are NH₃ (NFR 3B). Emissions from cattle are most important in Norway. Emissions of NH₃ from manure depend on several factors, e.g. type of animal, nitrogen content in fodder, manure management, climate, time of spreading of manure, cultivation practices and characteristics of the soil. In the IPCC default method a NH₃ volatilisation fraction of 20 per cent is used for the total N excretion by animals in the country. However, in the Norwegian emission inventory, yearly updated NH₃ volatilisation values are used, because this is expected to give more correct values for Norway. The estimated national volatilisation fractions from manure have differed between 19-21 per cent since 1990.

5.4.2.2 Method

A model is used for calculating the emissions of NH₃ from manure management. In Norway, all animal excreta that are not deposited during grazing are managed as manure. The estimations are made in accordance with the IPCC tier 2 method (IPCC 2006), using Norwegian values for N in excreta from different animals according to Table 5.4. The rationale for the Norwegian values for N in excreta is given in Karlengen et al. (2012). The N-excretion factors for cattle, poultry and pigs have been scientifically investigated, while the remaining categories have been given by expert judgements (Karlengen et al. 2012). Based on typical Norwegian feedstock ratios, the excretion of nitrogen (N) were calculated by subtracting N in growth and products from assimilated N and P. Comparisons have also been made with emission factors used in other Nordic countries and IPCC default factors.

The factors for cattle are based on equations using animal weight, production (milking cows), life time (young cattle) and protein content in the fodder as activity data.

The Nordic feed evaluation system (NorFor) was used to develop the nitrogen factors for cattle. Excretions of N in the manure were calculated as the difference between their intake, and the sum of what is excreted in milk, fetus and deposited in the animal itself. The procedure used for calculating the excretion of feces and N consisted of two steps:

1. Simulations in "NorFor" were conducted to gain values for the feces/manure characteristics covering a wide variation of feed characteristics (N content) and production intensities (milk yield/meat production).
2. The results from the simulations were used to develop regression equations between feces/manure characteristics and parameters related to the diet (N content) and animal

characteristics (milk yield, weight, age etc.).

Calculations of N-factors based on these equations have been made back to 1990 for cattle. For poultry and pigs, N-factors have been estimated for 2011 in Karlengen et al. (2012). The factors used until this update were estimated in 1988 (Sundstøl & Mroz 1988), and are regarded as still valid for 1990. A linear interpolation has been used for the years between 1990 and 2011. For the remaining animal categories the N in excreta are considered constant throughout the time series. The factors are shown in Table 5.4. The factors for total N are used in the estimations of N₂O emissions, and ammonium N are used in the estimations of NH₃ and NO_x emissions. Norwegian values are also used for the fraction of total excretion per species for each management system (MS) and for pasture. The fractions are updated every year.

Table 5.4. N in excreta from different animals¹. 2015. kg/animal/year unless otherwise informed in footnote

	Total N	Ammonium N
Dairy cow	129.0	73.5
Beef cow	64.9	36.0
Replacement heifer ²	86.6	47.6
Bull for slaughter ²	69.8	42.2
Finishing heifer ²	67.5	41.1
Young cattle ³	43.97	25.59
Horses	50.0	25.0
Sheep < 1 year	7.7	4.3
Sheep > 1 year	11.6	6.38
Goats	13.3	7.9
Pigs for breeding	23.5	15.7
Pigs for slaughtering ⁴	3.2	2.13
Hens	0.670	0.29
Chicks bred for laying hens ⁴	0.046	0.017
Chicks for slaughtering ⁴	0.030	0.011
Ducks, turkeys/ goose for breeding	2.0	0.8
Ducks, turkeys/ goose for slaughtering ⁴	0.4	0.18
Mink	4.3	1.7
Foxes	9.0	3.6
Reindeer	6.0	2.7
Deer	12.0	5.4

¹ Includes pasture.

² Factors for excreted nitrogen apply for the whole life time of animals, and nitrogen is calculated when animals are slaughtered/replaced.

³ Average factor for all heifers for slaughter and replacement and bulls for slaughter, per animal and year.

⁴ Per animal. For these categories, life time is less than a year. This means that the number of animals bred in a year is higher than the number of stalls (pens).

Source: Karlengen et al. (2012) and estimations by Statistics Norway

Ammonia, together with NO_x, volatilised from manure storage is part of the estimations of indirect N₂O emissions from atmospheric deposition. A model is used for calculating the emissions of ammonia from manure management and from spreading of manure. The principle of the model is illustrated in Figure 5.4.

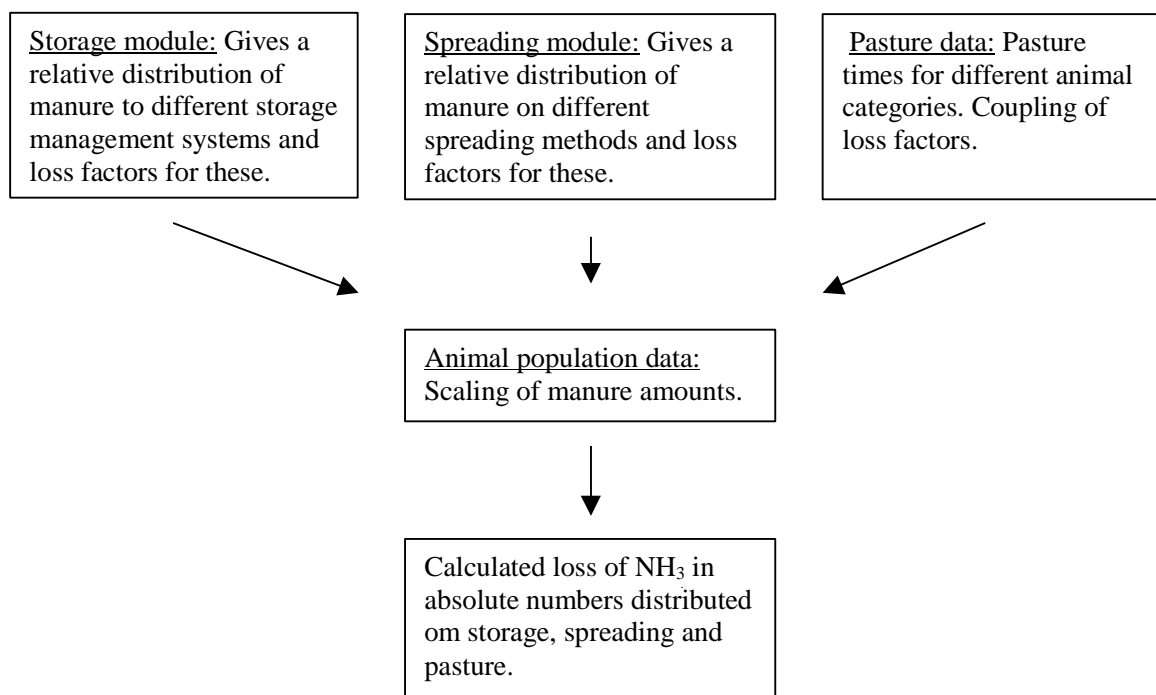


Figure 5.4 The principle of the NH₃ model

The storage module in the NH₃ model gives the relative distribution of manure nitrogen to the different storage management systems. Total NH₃ emissions from storage are estimated by multiplying the different emission factors for the storage systems by the amount of manure nitrogen (ammonium N) for each storage system and summarizing the results. The amount of ammonium nitrogen in the manure is estimated by the number of animals and ammonium nitrogen excretion factors for each type of animal (see Table 5.4).

5.4.2.3 Activity data

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

Surveys for assessing use of management systems have been carried out in 2000, 2003 and 2013. The distribution of manure systems in 2015 is given in Table 5.5.

Table 5.5 Fraction of total excretion per animal category for each management system and for pasture (MS) used in the estimations. 2015

	In-house slurry pit	Tank without cover	Tank with cover	In-house deep litter	Dry lot	Heaps	Pasture range and paddock
Dairy cattle	0.60	0.05	0.12	0.02	0.01	0.03	0.17
Other cattle	0.50	0.04	0.10	0.02	0.01	0.02	0.31
Swine	0.63	0.11	0.20	0.02	0.00	0.04	0.00
Poultry	0.03	0.00	0.00	0.00	0.00	0.97	0.00
Sheep	0.29	0.00	0.00	0.07	0.01	0.18	0.45
Goat	0.16	0.00	0.00	0.03	0.02	0.43	0.37
Horse	0.19	0.00	0.00	0.03	0.03	0.51	0.25
Fur bearing animals	0.25	0.00	0.00	0.04	0.03	0.68	0.00

Source: Data for storage systems from Statistics Norway (Statistics Norway et al. 2015), data for pasture times from (TINE BA Annually) (Dairy cattle, goat), Statistics Norway's Sample Survey 2001 (Statistics Norway 2002b)

Data on storage systems for other years are not available. Separate estimations of the effects on emissions of the assumed changes in storage systems since 1990 show that these assumed changes do not impact significantly. For the intermediate years 2004-2012 between the surveys of 2003 and 2013, the distribution of management system has been estimated using a linear interpolation of changes between 2003 and 2013, for each system. The 2013 data on storage systems will be used in approaching years until newer data becomes available. The surveys on management systems do not include pasture. Data for pasture times for dairy cattle and dairy goat are however annually updated in the Cow Recording System, while for the other animals, data from Sample survey of agriculture and forestry for 2001 at Statistics Norway is used. The data source for pasture times for dairy cattle in 2015 has however not been updated since 2013.

In the manure surveys of 2000 and 2013, the manure of each management system is distributed by all combinations of the following regions and productions:

Regions:

- South-Eastern Norway
- Hedmark and Oppland
- Rogaland
- Western Norway
- Trøndelag
- Northern Norway

Production¹⁹:

- Cattle
- Pigs
- Sheep
- Goats and horses
- Poultry

5.4.2.4 Emission factors

Emission factors vary with production and storage system; in the model there is no variation between regions. The factors are based on data from Denmark, Germany and Netherlands, since measurements of NH₃-losses in storage rooms have so far not been carried out in Norway. The factors are shown in Table 5.6.

Table 5.6. Emission factors for various storage systems and productions. Per cent losses of N of ammonium N

	Storage system						
	Manure cellar for slurry	Open manure pit for slurry	Manure pit for slurry with lid	Open flag-stones	Indoor built up/deep litter	Outdoor built up/enclosure	Storage for solid dung and urine
	Gutter	Gutter	Drainage to gutter				
Cattle, milking cow:							
Loss from animal room	5	5	5	5	8	8	5
Loss from storage room	2	9	2	2	15	15	15
Total loss	7	14	7	7	23	23	20
Pigs:							
Loss from animal room	15	15	15	15	15	15	20
Loss from storage room	4	6	2	2	25	25	30
Total loss	19	21	17	17	40	40	50
Sheep and goats:							
Loss from animal room	5	5	5	5	8	8	5
Loss from storage room	2	6	2	2	10	10	10
Total loss	7	11	7	7	18	18	15
Poultry:							
Loss from animal room	12	10	12	12	25	25	25
Loss from storage room	15	15	15	15	25	25	25
Total loss	27	25	27	27	50	50	50
Other animals:							
Loss from animal room	5	NO	NO	NO	15	15	15
Loss from storage room	10	NO	NO	NO	15	15	15
Total loss	15	NO	NO	NO	30	30	30

Source: Morken et al. (2005)

The factors are combined with activity data from the Statistics Norway survey of different

¹⁹ The grouping of animals are different in the two surveys. Cattle is one category in the 2000 survey and three categories in the 2013 survey. Goats are grouped with sheep in the 2000 survey, but with horses in the 2013 survey. Horses are grouped with other animals in the 2000 survey. Fur bearing animals are not included in the 2013 survey, but added to the horse/goat category. All manure from fur bearing animals are considered to be stored in heaps.

storage systems in 2000 (Gundersen & Rognstad 2001), the Sample survey of agriculture and forestry 2003 (Statistics Norway 2004) and Use of inorganic and organic fertilisers in agriculture 2013 (Statistics Norway et al. 2015), and emission factors for NH₃ emissions from storage of manure and stalled animals are calculated for production and region (Table 5.7). To estimate losses, these emission factors are in turn multiplied with the amount of manure (based on number of animals and N-factors per animal, Table 5.4). The changes in storage systems from 2003 to 2013 have been linearly interpolated in the intermediate years. From 1990-2002 and from 2013, the number of animals is the only activity data that differs from year to year.

Table 5.7. Average emission factors for the manure storage systems used, distributed on type of animal production and region. 2015. Per cent of ammonium N

	South-Eastern Norway	Hedmark/ Oppland	Rogaland	Western Norway	Trøndelag	Northern Norway
Cattle	10.3	9.4	7.8	7.6	7.8	7.8
Pigs	23.5	21.9	19.2	19.1	20.2	19.2
Sheep	13.3	12.6	8.9	10.7	11.3	11.9
Poultry	50.0	48.5	49.1	49.9	48.9	50.0
Goats, horses and fur-bearing animals	29.4	27.6	29.8	24.7	29.7	19.2

Source: Statistics Norway, NH₃-model estimations.

5.4.2.5 Uncertainties

Uncertainty estimates are provided in Appendix C.

5.4.2.5.1 Activity data

Emissions are estimated from the animal population. The data for the number of animals are considered to be known within ± 5 per cent.

For the emissions of NH₃ from manure management, Norwegian data for N in excreta are used. The nitrogen excretion factors are uncertain, but the range is considered to be within ± 15 per cent (Rypdal 1999). The uncertainty has not been estimated for the revised nitrogen excretion factors from Karlengen et al. (2012), and in the key category analysis the uncertainty estimate for the country specific nitrogen excretion factors from 1999 is still used as the best available estimate. This can be considered as a conservative estimate of the uncertainty since it is expected that the new nitrogen excretion factors have a lower uncertainty. The uncertainty is connected to differences in excretion between farms in different parts of the country, the fact that the survey farms may not have been representative, general measurement uncertainty and the fact that fodder and fodder practices have changed since the factors were determined.

There is also an uncertainty connected to the division between different storage systems for manure, which is considered to be within ± 10 per cent, and the division between storage and pasture, which is considered to be within ± 15 per cent.

5.4.2.5.2 Emission factors

Ammonia emissions from agriculture are estimated based on national conditions. There are uncertainties in several parameters as fraction of manure left on pastures, amount of manure, conditions of storage, conditions of spreading and climate conditions. An uncertainty analysis for

the revised NH_3 model, which has been in use since 2003, has not been performed. However, the revision of the model is believed to have reduced the uncertainty. Also the new estimations of nitrogen excretion from animals (Karlengen et al. 2012) are believed to have reduced uncertainty further.

5.4.3 NO_x emissions from manure management

5.4.3.1 Description

Emissions from NO_x from animal manure is included in the inventory in the 2017 submission. The amounts of N from NO_x emissions from manure management systems, spreading of manure and droppings on pastures are however small, being about 5 per cent of the N in NH_3 emissions from the same sources.

5.4.3.2 Method

In Norway, all animal excreta that are not deposited during grazing are managed as manure. Norwegian values for N in excreta from different animals according to Table 5.4 are used. How these are estimated is described in section 5.4.2.2. Norwegian values are also used for the fraction of total excretion per species for each management system (MS) and for pasture. The fractions are updated every year.

NO_x volatilised from manure storage is part of the estimations of indirect N_2O emissions from atmospheric deposition.

5.4.3.3 Activity data

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

Surveys for assessing use of management systems have been carried out in 2000, 2003 and 2013. The distribution of manure systems used in the 2015 inventory is given in Table 5.8. Norway has developed calculation models for estimating N_2O and NH_3 emissions. The basic data from the 2013 Manure Survey is however treated a little differently in the two models, and different manure storage categories are used. For estimating NO_x emissions, the setup for the N_2O estimations was more suitable to use than the NH_3 model. This means that the NO_x estimations are in better coherence with the calculations for the N_2O emissions than with the NH_3 calculations regarding the use of the data from the manure surveys. This also explains the differences between table Table 5.8 and Table 5.5.

Table 5.8 Fraction of total excretion per specie for each management system and for pasture (MS) used in the estimations of NO_x . 2015

	Pit storage below animal confinement	Liquid / slurry without cover	Liquid / slurry with cover	Solid storage	Cattle and swine deep bedding	Dry lot	Pasture range and paddock	Poultry manure
Dairy cattle	0.01	0.15	0.68	0.00	0.00	0.00	0.17	

Mature non dairy cattle	0.01	0.09	0.38	0.08	0.03	0.10	0.31	
Young cattle	0.01	0.10	0.54	0.02	0.01	0.02	0.31	
Pigs	0.01	0.20	0.74	0.02	0.00	0.03	0.00	
Sheep	0.14	0.01	0.28	0.07	0.02	0.04	0.45	
Goat	0.12	0.00	0.20	0.03	0.03	0.25	0.37	
Horse	0.14	0.00	0.24	0.04	0.03	0.30	0.25	
Poultry								1.00
Fur bearing animals				1.00				
Reindeer, deer and other animals							1.00	

Source: Data for storage systems from Statistics Norway (Statistics Norway et al. 2015), data for pasture times from TINE BA (Annually) (Dairy cattle, goat), Statistics Norway's Sample Survey 2001 (Statistics Norway 2002a) (non-dairy cattle, sheep) and expert judgements (poultry, fur bearing animals and other).

5.4.3.4 Emission factors

Table 5.9 NO_x emission factors for manure management per manure management system

	EF (AWMS) (kg NO-N/kg of Nex in AWMS)
Pit storage below animal confinement	0.0001
Liquid / slurry without cover	0.0001
Liquid / slurry with cover	0.0001
Solid storage	0.01
Cattle and swine deep bedding	0.01
Dry lot	0.01
Poultry manure	0.01
Pasture range and paddock	0

Source:EEA (2016)

N excretions is estimated as urine-N, which is the same N excretion factor that is used in the estimations of NH₃ from manure management systems.

5.4.4 NMVOC emissions from manure management

5.4.4.1 Description

Livestock production is a source of emissions of NMVOC during feeding with silage and manure management. The emissions from housing are included in the inventory in the 2017 submission.

5.4.4.2 Method

The emissions have been estimated using Tier 1 methodology from EEA (2016), where animal population numbers are multiplied with a default emission factor.

5.4.4.3 Activity data

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

5.4.4.4 Emission factors

Emission factors are taken from EEA (2016), see Table 5.10 Default Tier 1 EF for NMVOC Table 5.10.

Table 5.10 Default Tier 1 EF for NMVOC

	EF, with silage feeding	EF, without silage feeding
Livestock	NMVOC, kg AAP ⁻¹ . a ⁻¹	
Dairy cattle	17.937	8.047
Non-dairy cattle ¹	8.902	3.602
Sheep	0.279	0.169
'Swine' (Fattening pigs ²)	-	0.551
'Swine' (Sows)	-	1.704
Goats	0.624	0.542
Horses	6.028 ⁵	4.275
Laying hens (laying hens and parents)	-	0.165
Broilers (broilers and parents)	-	0.108
Turkeys ³	-	0.489
Other poultry (ducks, geese) ³	-	0.489
Other animals (Fur animals)	-	1.941
Other animals (Reindeer ⁴)	-	0.045

(1) Includes young cattle, beef cattle and suckling cows

(2) Includes piglets from 8 kg to slaughtering

(3) Based on data for turkeys

(4) Assume 100% grazing

(5) EMEP/EEA factor is 7.781, but has been reduced to 6.028 which is mean value between the factors with and without silage because silage for horses normally is dryer than other silage

Source: EEA (2016) and expert judgement by Dag Austbø (see note 5)

The share of silage in fodder intake is registered for dairy cows in the Cow recording system, see Table 5.11.

Table 5.11 Silage fodder intake, dairy cows, and estimated emission factor, dairy cows.

	Silage fodder, average share of daily intake	Estimated NMVOC, kg AAP ⁻¹ . a ⁻¹
1990	38.1	11.8
1991	38.1	11.8
1992	38.1	11.8
1993	38.1	11.8
1994	40.4	12.0
1995	39.6	12.0
1996	40.0	12.0
1997	40.0	12.0
1998	41.2	12.1
1999	41.4	12.1
2000	41.6	12.2
2001	42.7	12.3
2002	42.2	12.2
2003	43.3	12.3

2004	42.9	12.3
2005	43.6	12.4
2006	43.2	12.3
2007	43.9	12.4
2008	44.4	12.4
2009	44.8	12.5
2010	45.6	12.6
2011	45.3	12.5
2012	45.1	12.5
2013	45.3	12.5
2014	45.3	12.5

Source: Cow recording system (TINE BA Annually)

Amounts of silage fodder is not registered systematically for the other ruminants, and the share of silage in fodder intake is therefore based on expert judgments for these groups of animals.

Table 5.12 Silage as share of total feed intake and estimated emission factor for NMVOC emission per animal, other ruminants

Livestock	Silage fodder, average share of daily intake	Estimated average EF, NMVOC, kg AAP-1. a-1
Growing cattle and mature non dairy cattle	50 ¹	6.25
Sheep	33 ²	0.21
Goats	40 ³	0.57
Horses	33 ⁴	4.85

² Expert judgement Finn Avdem, Nortura. July 2016

³ Judgement by Statistics Norway. November 2016.

⁴ Expert judgement Dag Austbø, Norwegian University of Life Sciences. June 2016

Source: EEA (2016)

5.4.5 PM emissions from manure management

5.4.5.1 Method

Tier 1 methodology from EEA guidebook 2013 p. 16-17.

5.4.5.2 Activity data

Emissions are estimated from the animal population. How the animal population is estimated is described in section 5.2.

5.4.5.3 Emission factors

Default Tier 1 emission factors are used, see Table 5.13.

Table 5.13 Default Tier 1 estimates of EF for particle emissions from animal husbandry (housing)

Livestock	EF, TSP kg AAP-1. a-1	EF, PM10 kg AAP-1. a-1	EF, PM2,5 kg AAP-1. a-1	Source
Dairy cows	1.38	0.63	0.41	EEA (2016)
Other cattle	0.59	0.27	0.18	EEA (2016)

Fattening pigs	0.75	0.34	0.06	EEA (2015)
Weaners	0.21	0.1	0.02	EEA (2015)
Sows	1.53	0.69	0.12	EEA (2015)
Sheep	0.139	0.0556	0.0167	EEA (2016)
Goats	0.139	0.0556	0.0167	EEA (2016)
Horses	0.48	0.22	0.14	EEA (2016)
Laying hens (laying hens and parents)	0.119	0.119	0.023	EEA (2015)
Broilers (broilers and parents)	0.069	0.069	0.009	EEA (2015)
Other poultry (ducks, geese, turkeys)	0.52	0.52	0.07	EEA (2015)
Fur animals	0.018	0.0081	0.0042	EEA (2016)

5.4.6 Completeness

Major missing emission components are not likely.

5.4.7 Source specific QA/QC

Statistics Norway, in cooperation with the Norwegian University of Life Sciences (NMBU), made improvements in 2003 in the calculation model for NH₃ emissions from the agricultural sector. Data sources used for the recalculations in the revised NH₃ model are coefficients from the Norwegian University of Life Sciences, and three surveys from Statistics Norway; two manure surveys (Gundersen & Rognstad 2001) and the sample survey of agriculture and forestry 2001 (Statistics Norway 2002b).

Statistics Norway's detailed manure survey gave more extended activity data which are better related to emission source categories, for manure management and spreading. New loss factors for different manure management categories are also used in the revised NH₃-model. These factors are closer connected to specific activities.

In 2014, a new manure survey for 2013 was carried out by Statistics Norway (Statistics Norway 2015). The results are implemented in the estimations of CH₄ and N₂O emissions from manure. Statistics Norway's detailed manure survey gave more extended activity data which are better related to emission source categories, for manure management and spreading. New loss factors for different manure management categories are also used in the revised NH₃-model. These factors are closer connected to specific activities.

5.5 Crop production and agricultural soils

NFR 3D

Last update: 01.03.17

5.5.1 Description

The use of synthetic fertilisers, animal excreta nitrogen, sewage sludge and other organic

fertilisers applied to soils, and droppings on pastures result in emissions of NH_3 . Agricultural activities are also a source of NO_x , NMVOC and non-combustion emissions of particles.

5.5.2 NH_3 emissions from agricultural soils

5.5.2.1 Method

5.5.2.1.1 NH_3 emissions from use of inorganic N-fertilizers

NFR 3Da1

The calculations of NH_3 emissions from the use of synthetic fertiliser are based on the amounts of nitrogen supplied and emission factors for the percentage of nitrogen emitted as NH_3 during spreading. More information about the calculation of $\text{frac}_{\text{GASF}}$ is given in Miljødirektoratet (2016) Annex IX, section 3.3. The amount of nitrogen supplied is estimated based on data for total annual amount of fertiliser sold in Norway and its nitrogen content, corrected for the amount of synthetic fertiliser applied in forest. The resulting amount is expected to be the amount applied on agricultural fields.

5.5.2.1.2 NH_3 emissions from animal manure applied to soils

NFR 3Da2a

In Norway, all animal excreta that are not deposited during grazing are used as manure and applied to soils. Further, it is assumed that animals do not emit N_2O themselves, NH_3 emissions in storage, and N_2O emissions in storage and manure application are all estimated individually and the emission estimates are based on the same nitrogen pool.

NH_3 emissions from spreading of manure depend on several factors, e.g. climate and time of spreading of manure, type of cultivation and cultivation practices and characteristics of the soil. In the IPCC default method, a NH_3 volatilisation fraction of 20 per cent is used for the total N excretion by animals in the country. However, in the Norwegian emission inventory, yearly updated national ammonia volatilisation values are used, because this is considered to give more correct values for Norway. The estimated national volatilization fractions have differed between 18-20 per cent since 1990.

Emissions of ammonia are calculated for spreading of manure on cultivated fields and meadow. The total amount of manure nitrogen that is spread is estimated by the number of animals and nitrogen excretion factors for each type of animal, and is thereafter distributed on different spreading methods based on national data. The nitrogen basis for the estimated amounts of nitrogen that volatilises as NH_3 during spreading has been corrected for the amount of nitrogen in the NH_3 that volatilises during storage, unlike the method used in the N_2O estimations. N lost as N_2O and leaching during storage is however not deduced from the N basis. Total emissions from spreading are estimated by emission factors for each different spreading methods used multiplied by the amount of manure nitrogen spread with the respective method.

5.5.2.1.3 NH_3 emissions from sewage sludge applied to soils

NFR 3Da2b

To calculate NH_3 emissions from sewage sludge used as fertiliser, the fraction of N in manure

lost as NH_3 is used ($\text{frac}_{\text{GASM}}$). The loss equals to total N in sewage sludge multiplied by $\text{frac}_{\text{GASM}}$.

5.5.2.1.4 *NH_3 emissions from other organic fertilisers applied to soils*

NFR 3Da2c

Emissions of NH_3 from other organic fertilisers applied to soils are estimated by multiplying estimated amounts of N in organic fertilisers with the $\text{frac}_{\text{GASM}}$ -factor. The annual amount of nitrogen in other organic fertilisers applied in agriculture during the period 1990-2013 was assessed in 2014 (Aquateam COWI AS 2014). Other organic fertilisers comprise three main categories; biomanure and other biological residues from biogas plants, compost from composting plants and other commercial organic fertiliser products sold. This was a practically non-existent source of nitrogen before 2000. Since then, it has varied very much over the years. In 2013, the nitrogen from this source contributed to about 25 per cent of the nitrogen in the sewage sludge applied. A new assessment has to take place to get figures for 2014 and later.

5.5.2.1.5 *NH_3 emissions from urine and dung deposited by grazing animals*

NFR 3Da3

Animal population data, data for pasture times, and factors for the nitrogen amount in excreta for different animal categories give the nitrogen amounts for the animal categories on pastures. The amount of animal manure dropped on pastures is given by estimations of total N in manure excreted from animals and data for pasture times (Table 5.5). It is assumed that the share of time the animals spend on pastures corresponds to the share of total N produced that is dropped during grazing. The emissions are calculated by the estimated amount of N deposited during grazing multiplied with specific emission factors by animal category (see Table 5.19) are used.

5.5.2.2 Activity data

NH_3 emissions from use of synthetic fertiliser

The Norwegian Food Safety Authority calculates a total value for annual consumption of synthetic fertilisers in Norway based on sales figures. These data are corrected for the amount of fertiliser used in forests, which is provided by the Norwegian Institute of Bioeconomy Research.

For the calculation of the emission of NH_3 we need a specification of the use of different types of synthetic fertiliser since the NH_3 emission factor vary between different types. This is given by the Norwegian Food Safety Authority for the years from 2000. Due to lack of data for the years before 2000, we have to assume that the percentual distribution between the usage of different fertiliser types is the same as in 1994 for these years.

Animal manure applied to soil and pasture

There are several sources of activity data on spreading of manure. The main sources are the manure survey in 2000 and in 2013 by Statistics Norway and Statistics Norway et al. (2015) various sample surveys of agriculture and forestry 1990-2007 and the animal population. Table 5.14 shows the estimated changes in emissions after implementing the data from the 2013 survey.

Table 5.14. Estimated NH₃-emissions from manure management, pasture and application of manure in 2013, based on old and new survey data. Tonnes NH₃

	Based on survey data from 2000 and 2003	Based on survey data from 2013
Total	22 084	21 093
Manure storage	6 015	6 483
Pasture	919	938
Application of manure	15 150	13 673

Source: Statistics Norway, emission statistics

The survey of 2013 (Statistics Norway et al. 2015) shows that the changes in practices for manure handling from 2003 to 2013 had lowered the NH₃ emissions in 2013 by about 5 per cent, all other factors remaining equal. There was however a distribution of emissions from storage systems to application. The first increased about 8 per cent, while the latter decreased approximately 10 per cent.

Animal population is updated annually. The animal population estimation methodology is described in section 6.2. Data from the manure survey do only exist for 2000 and 2013, while the data from the sample surveys have been updated for several, but not all, years. The manner of spreading the manure also affects the NH₃ emission estimates.

Data for time on pasture and share of animals on pasture are collected from the Sample Survey in Statistics Norway 2001 and from TINE BA (TINE BA is the sales and marketing organisation for Norway's dairy cooperative and covers most of the milk production). The data from TINE BA comprises pasture data for goats and milking cows and are updated annually. All other pasture data are from the Statistics Norway Sample survey 2001. The parameters used in the calculations and their sources are shown in Table 5.15.

Table 5.15. Parameters included in the estimation of NH₃ emissions from manure

Parameters (input)	Sources
Number of animals	Statistics Norway (applications for productions subsidies, no. and weight of approved carcasses), the Cow Recording System at TINE BA
Nitrogen factors for manure	Karlengen et al. (2012), various sources, compiled by Statistics Norway
Area where manure is spread, split on cultivated field and meadow.	Statistics Norway (Sample Surveys of Agriculture, various years), Gundersen and Rognstad (2001) and Statistics Norway et al. (2015)
Area and amount where manure is spread, split on spring and autumn.	Gundersen and Rognstad (2001) and Statistics Norway et al. (2015)
Amount of manure is spread, split on spring and autumn.	Statistics Norway (Sample Surveys of Agriculture, various years)
Addition of water to manure	Gundersen and Rognstad (2001) and Statistics Norway et al. (2015), expert judgements, Statistics Norway's Sample Survey 2007
Spreading techniques	Gundersen and Rognstad (2001) and Statistics Norway et al. (2015), expert judgements,
Usage and time of harrowing and ploughing.	Statistics Norway (Sample Surveys of Agriculture), Gundersen and Rognstad (2001) and Statistics Norway et al. (2015), expert judgements,
Pasture times for different animal categories	TINE BA (Annually) (Dairy cattle, goats), Statistics Norway's Sample Survey 2001 (Statistics Norway 2002b) (non-dairy cattle, sheep), expert judgements.

5.5.2.3 Emission factors

Synthetic fertiliser

Different types of synthetic fertilisers are being used, resulting in different emissions of NH₃. Their respective share is based on sales statistics provided annually by the Norwegian Food Safety Authority for the years from 2000. For earlier years the distribution is based on data from 1994. The NH₃ emission factors (per cent loss of N) for the different types of fertilisers are shown in Table 5.16.

Table 5.16. Emission factors for NH₃-N for different fertilisers. Per cent loss of applied N

Fertiliser	Emission factor (per cent of applied N)
Urea	15
Ammonium sulphate and Ammonium nitrate	5
Calcium nitrate	0
Calcium ammonium nitrate	1
NPK (Nitrogen, phosphorus, potassium)	1
Other	1

Source: ECETOC (1994) and Norsk Hydro, pers. comm.²⁰

Animal manure applied to soil and pasture

²⁰ Norsk Hydro (1995): Personal information, Kaarstad, Norsk Hydro.

Emission factors for spreading of stored manure vary with spreading method (Gundersen & Rognstad 2001), water content (Statistics Norway 2007), type and time of treatment of soil (Gundersen & Rognstad 2001), time of year of spreading (Gundersen & Rognstad 2001); Gundersen 2015, (Statistics Norway 2007), cultivation, and region. The basic factors used are shown in Table 5.17.

Table 5.17. Emission factors for $\text{NH}_3\text{-N}$ for various methods of spreading of manure. Per cent of ammonium N

			Western and northern Norway			Southern and eastern Norway		
			Spring	Summer	Autumn	Spring	Summer	Autumn
Meadow								
Surface spreading			0.5	0.6	0.4	0.5	0.6	0.4
Injection			0.1	0.1	0.05	0.1	0.1	0.05
Water mixing			0.3	0.3	0.2	0.3	0.3	0.2
Dry manure			0.04	0.1	0.1	0.04	0.1	0.1
Open fields								
Method	Time before down-moulding	Type of down-moulding						
Surface spreading	0-4 hrs	plow	0.2		0.2	0.15		0.3
Surface spreading	+ 4 hrs	plow	0.5		0.35	0.4		0.4
Surface spreading	0-4 hrs	harrow	0.4		0.35	0.35		0.35
Surface spreading	+ 4 hrs	harrow	0.5		0.45	0.45		0.45
Water mixing	0-4 hrs	plow	0.1		0.1	0.1		0.15
Water mixing	+ 4 hrs	plow	0.25		0.2	0.2		0.25
Water mixing	0-4 hrs	harrow	0.2		0.2	0.2		0.2
Water mixing	+ 4 hrs	harrow	0.3		0.25	0.25		0.25
Dry manure			0.04		0.1	0.04		0.1

Source: Morken and Nesheim (2004).

Table 5.18. Average NH_3 emission factors for cultivated fields and meadows after time of spreading and region. 2015. Per cent of ammonium N

	South-Eastern Norway		Hedmark/Oppland			Rogaland	Western Norway		Trøndelag		Northern Norway	
	Field	Meadow	Field	Meadow	Field	Meadow	Field	Meadow	Field	Meadow	Field	Meadow
Spring	15.6	40.0	20.4	42.3	4.0	38.0	4.0	43.2	20.8	42.4	7.4	42.6
Autumn	17.4	28.9	20.0	30.7	10.0	27.4	10.0	31.4	24.6	30.8	12.9	31.0

Source: Statistics Norway, NH_3 -model estimations

The factors in Table 5.17 are combined with the activity data in the Sample survey of agriculture and forestry 2007 (Statistics Norway 2007) and a time series on mixture of water in manure, and emission factors for NH_3 emissions from spreading of manure distributed to meadow and cultivated fields, time of season and region are calculated (see Table 5.18). These factors are in turn connected to activity data that are updated in the years since 1990, i.e. number of animals (amount of manure), time of spreading and type of cultivation of the areas where the manure is spread.

The emission factors used for the calculation of the NH_3 emissions from grazing animals are shown in Table 5.19. These are the same as the emission factors used in Germany (Dämmgen et

al. 2002) and Denmark (Hutchings et al. 2001).

Table 5.19. NH₃ emission factors from droppings from grazing animals on pasture. Per cent of ammonium N applied

	N-loss/N applied
Cattle	7.5
Sheep and goats (from 2013 sheep only) ¹	4.1
Reindeer	4.1
Other animals (from 2013 including goat) ¹	7.5

¹ Goats are grouped with sheep in the 2000 survey, and with horses in the 2013 survey.

Source: Dämmgen et al. (2002), Hutchings et al. (2001)

5.5.3 NO_x emissions from agricultural soil

NFR 3Da1

5.5.3.1 Method

The sum of all nitrogen applied to soil has been multiplied with the default tier 1 emission factor to estimate the nitric oxide emission from crop production. Thereafter the amount of NO is translated to amount of NO₂.

5.5.3.2 Activity data

Total N from the following sources are included:

- Synthetic fertilizers
- Animal manure spread
- Sewage sludge
- Other organic fertilisers

Synthetic fertiliser

The Norwegian Food Safety Authority calculates a total value for annual consumption of synthetic fertilisers in Norway based on sales figures. These data are corrected for the amount of fertiliser used in forests, which is provided by the Norwegian Institute of Bioeconomy Research.

Animal manure spread

In Norway, all animal excreta that are not deposited during grazing are used as manure and applied to soils. The total amount of N in manure used as fertiliser is equivalent to total N excreted from the animals deducted for the amount dropped during grazing. How the amount of nitrogen in animal manure are calculated is described further in section 5.4.2.3.

Sewage sludge applied to soils

Statistics Norway (waste water statistics) annually gives values for the amount of sewage sludge, and the fraction of the sewage sludge that are applied on fields. The N-content in the sludge is given in Statistics Norway (2001), and the same value of 2.82 per cent is used for all years.

Other organic fertilisers applied to soils

How the amount of nitrogen in other organic fertiliser are estimated is described further in section 5.5.2.1.4.

5.5.3.3 Emission factors

Tier 1 default emission factor for NO emissions from agricultural soils has been used, see Table 5.20.

Table 5.20. Tier 1 default emission factor for NO emissions from agricultural soils

Pollutant	Value	Unit
NO	0.026	kg kg ⁻¹ fertilizer-N applied

Source: EEA (2015B)

5.5.4 NMVOC emissions from agricultural soils

NFR 3Da1

5.5.4.1 Method

The tier 1 methodology has been used, multiplying cultivated area in Norway with the default emission factor from EEA.

5.5.4.2 Activity data

The activity data used are fully cultivated area given by Statistics Norway.

5.5.4.3 Emission factors

The recommend average emission factor of 0.86 kg NMVOC per ha from EEA (2016) is used. There are great variations in NMVOC emissions, dependent on crop, temperatures, yield etc. The average factor is based on a 50-50 distribution between grass and cropland. In Norway, about two thirds of the agricultural land is grassland. This may indicate an underestimation, but lower average temperatures compared to the average for the whole EMEP area has the opposite effect.

5.5.5 Particle emissions from farm-level agricultural operations

NFR 3Dc

Agriculture is responsible for various types of non-combustion emissions of particles. This is for example dust from crops that are harvested, soil dust from work with agricultural machines, wood particles from felling of trees etc.

5.5.5.1 Method

The tier 1 methodology described in EEA (2016) is used. The area of crop land in Norway is multiplied with tier 1 emission factors, which gives emissions per area unit.

5.5.5.2 Activity data

The area of crop land (open fields and gardens) is used since these emissions are mainly from combine harvesting and soil cultivation (EEA 2016). The emissions may therefore be slightly overestimated since parts of the cropland is not plowed or harrowed every year.

5.5.5.3 Emission factors

Table 5.21. Tier 1 emission factors for emissions of particles from farm-level agricultural operations. kg/ha

Pollutant	Value (kg ha ⁻¹)
PM10	1.56
PM2.5	0.06
TSP	1.56

Source: EEA (2016)

5.5.6 Uncertainties

5.5.6.1 Activity data

There are several types of activity data entering the calculation scheme:

Sales of nitrogen fertiliser: The data are based on sales figures during one year (The Norwegian Food Safety Authority). The uncertainty in the sales figures is within ± 5 per cent (Rypdal & Zhang 2000). In addition, there is a possible additional error due to the fact that sales do not necessarily equal consumption in a particular year, due to storage. The share of the various types of nitrogen fertiliser is assumed to be the same as in an investigation in 1994, and the error connected to this approach has probably increased over the years. The effect for the uncertainty in activity data due to these two factors has not been quantified, but it is assumed that it can be more important than the uncertainty in the sales figures.

Amount of nitrogen in manure: The figures are generated for each animal type, by multiplying the number of animals with a nitrogen excretion factor. The nitrogen excretion factors are uncertain. The range is considered to be within ± 15 per cent (Rypdal 1999). The uncertainty is connected to differences in excreted N between farms in different parts of the country, that the survey farms may not have been representative, general measurement uncertainty and the fact that fodder and feeding practices have changed since the factors were determined. This uncertainty was substantially reduced in 2013 when the nitrogen factors were assessed in a research project (Karlengen et al. 2012).

The uncertainty connected to the estimate of the amount of manure is higher than for the amount of synthetic fertiliser used.

Fate of manure: There is significant uncertainty connected to the allocation of manure between what is used as fertiliser and droppings on pastures.

Atmospheric deposition of agricultural NH₃ emissions: The data are based on national figures for NH₃ emission from agriculture. These are within ± 30 per cent (Rypdal 1999).

5.5.6.2 Emission factors

NH₃

The uncertainty in the estimate of NH₃ emissions from use of fertiliser is assessed to be about ± 20 per cent (Rypdal & Zhang 2001). This uncertainty could be lower if better data on fertiliser composition were obtained. The uncertainty is higher for animal manure (± 30 per cent (Rypdal & Zhang 2001)). This is due to uncertainties in several parameters (fraction of manure left on

pastures, amount of manure, conditions of storage, conditions of spreading and climate conditions) (Rypdal & Zhang 2001). Other factors that could lead to uncertainty are variation in storage periods, variation in house types and climate, and variation in manure properties. Include NO_x, PM and NMVOC.

5.5.7 Completeness

Major missing emission components are not likely.

5.5.8 Source specific QA/QC

Statistics Norway, in cooperation with the Norwegian University of Life Sciences (NMBU), made improvements in 2003 in the calculation model for NH₃ emissions from the agricultural sector. Data sources used for the recalculations in the revised NH₃ model are coefficients from the Norwegian University of Life Sciences, and two surveys from Statistics Norway; a manure survey (Gundersen & Rognstad 2001) and the sample survey of agriculture and forestry (2001). New factors for nitrogen excretion from animals and a revision of animal statistics has been made in 2012, to better reflect the actual nitrogen excretion from each animal category and to have a more correct linkage between the nitrogen excretion factors used and the different animal categories. Data from the manure survey of 2013 was implemented in the estimations of NH₃ emissions from manure in the 2016 submission (Statistics Norway et al. 2015).

5.5.9 Use of pesticides

NFR 3Df

Hexachlorobenzene (HCB) was earlier used as a pesticide, but is now forbidden. The use of this substance is not known in products in Norway today, but it can arise unintentionally and constitute a contamination in some products, among them pesticides. Pesticides can contain among other things pentachlorophenol, atrazine, simazine, picloram, pentachloronitrobenzene (PCNB, quintozone), chlorothalonil, endosulfan and chlopyralid (SYKE 2013). Emissions from the use of pesticides that can include a contamination of HCB are part of the emission inventory estimations. Information about the concentration of HCB in some of the above mentioned pesticides are shown in Table 5.22. This information is collected from Finland (SYKE 2013) and in the estimations it is supposed that half of the HCB remnants in the pesticides are emitted to air. The amounts sold of the substances that can contain contaminants of HCB have been given by the Norwegian Food Safety Authority and the Product Register in the Norwegian Environment Agency. The amount of the effective substance sold in Norway have been used as activity data for the period 1996 to 2008. Since 2008, no substances containing HCB have been sold in Norway. For the years 1990-1995 the value for 1996 is used due to lack of data.

Table 5.22 HCB-contamination in pesticides. mg/kg

Pesticide	
Chlorothalonil	10
Chlopyralid	2.5
Endosulfan	0.1
Simazine	1

Source: SYKE (2013)

5.5.9.1 Uncertainties

No uncertainty analysis has been made for this source.

5.5.9.2 Completeness

Major missing emission components are not likely.

5.5.9.3 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

5.6 Field burning of agricultural wastes

NFR 3F

Last update: 29.02.16

5.6.1 Description

Burning of agricultural residues gives emissions of a large range of standard combustion products. Emissions of NO_x, CO, NH₃, NMVOC, SO₂, particles and the heavy metals Pb, Cd, Hg, As, Cu and Cr, and benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4) and dioxins are included in the inventory.

5.6.2 Method

The emissions from the burning of crop residues are being calculated in accordance with a Tier 1 approach (EEA 2009):

$$E_{\text{Pollutant}} = AR_{\text{residue_burnt}} * EF_{\text{Pollutant}}$$

$E_{\text{Pollutant}}$ = emission (E) of pollutant

$AR_{\text{residue_burnt}}$ = activity rate (AR), mass of residue burnt (dry matter)

$EF_{\text{Pollutant}}$ = emission factor (EF) for pollutant

5.6.3 Activity data

The annual amount of crop residue burned on the fields is calculated based on crop production data for cereals and rapeseed from Statistics Norway, and estimates of the fraction burned made by the Norwegian Crop Research Institute and Statistics Norway. The fraction of crop residue burned on field was updated in 2012 by the Norwegian Agricultural Authorities²¹. This reduced the fraction for 2011 from 7.5 to 4 per cent. For cereals a water content of 15 per cent is used (IPCC 1997). The activity data are consistent with the data used in the estimations of N₂O from crop residues.

²¹ Johan Kollerud, Norwegian Agricultural Authorities, unpublished material 2012.

5.6.4 Emission factors

Table 5.23. Emission factors for agricultural residue burning.

Components	Emission factors	Unit	Source
Precursors			
NO _x	2.3	kg/ tonnes crop residue (d.m.) burned	(EEA 2016)
CO	66.7	kg/ tonnes crop residue (d.m.) burned	(EEA 2016)
SO ₂	0.5	kg/ tonnes crop residue (d.m.) burned	(EEA 2016)
NM VOC	5	kg/ tonnes crop residue (d.m.) burned	(EEA 2016)
NH ₃	2.4	kg /tonnes crop residue (d.m.) burned	(EEA 2016)
Heavy metals			
Pb	0.11	g/ tonnes crop residue (d.m.) burned	(EEA 2016)
Hg	0.14	g/ tonnes crop residue (d.m.) burned	(EEA 2016)
Cd	0.88	g/ tonnes crop residue (d.m.) burned	(EEA 2016)
As	0.0064	g/ tonnes crop residue (d.m.) burned	(EEA 2016)
Cr	0.08	g/tonnes crop residue (d.m.) burned	(EEA 2016)
Cu	0.073	g/ tonnes crop residue (d.m.) burned	(EEA 2013; EPA 2002)
Particles			
TSP	5.8	kg/ tonnes crop residue (d.m.) burned	(EEA 2016)
PM ₁₀	5.7	kg/ tonnes crop residue (d.m.) burned	(EEA 2016) (EEA 2016)
PM _{2.5}	5.4	kg/ tonnes crop residue (d.m.) burned	(EEA 2016)
BC	13	% of PM _{2.5}	GAINS model (IIASA)
benzo(a)pyrene	0,39266	g/ tonnes crop residue (d.m.) burned	(Jenkins et al. 1996)
benzo(b)fluoranthene	1,09678	g/ tonnes crop residue (d.m.) burned	(Jenkins et al. 1996)
benzo(k)fluoranthene	0,46806	g/ tonnes crop residue (d.m.) burned	(Jenkins et al. 1996)
indeno(1,2,3_cd)pyrene	0,33582	g/ tonnes crop residue (d.m.) burned	(Jenkins et al. 1996)
Dioxins	0.5	µg I-TEQ/tonnes crop residue (d.m.) burned	(EEA 2016)
PCB	2.7	µg/tonnes crop residue (d.m.) burned	(Nielsen et al. 2015), (EEA 2013)

Heavy metals and POPs

For heavy metals default emission factors from the EMEP/EEA emission inventory guidebook are used (EEA (2016)). The emissions of benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene (PAH-4) are calculated based on emission

factors from Jenkins et al. (1996).

5.6.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

5.6.6 Completeness

The main emission components from burning of agricultural residues are considered to be covered in the emission calculations.

5.6.7 Source specific QA/QC

In 2002, the emissions of NO_x, CO, Pb, Hg, Cd, and dioxins from agricultural residual burning were included in the Norwegian inventory, and in 2003, the emissions of particles, As, Cr and Cu were added. In 2011, also emissions of SO₂, NMVOC and NH₃ were included in the inventory. In 2016, a project to split PAH-4 emissions on individual PAHs; benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene has been performed. The time series were included but it should be noted that the figures for the earlier years have a higher uncertainty than the more recent years.

5.7 Other agricultural emission sources

NFR 3I

Last update: 09.03.15

5.7.1 Description

Straw treated with NH₃ to be utilised as fodder is a source of NH₃ emissions in Norway.

5.7.2 NH₃ emissions from treatment of straw

5.7.2.1 Method

Emissions of NH₃ from treatment of straw depend only on the amount of NH₃ used. The total amount of NH₃ used for treatment of straw in Norway is multiplied with the share of the NH₃ that is not integrated in the straw.

5.7.2.2 Activity data

The amount of NH₃ used per year is obtained from the Budget Committee for Agriculture²². The area of cultivated fields is annually updated from Statistics Norway's agriculture statistics.

5.7.2.3 Emission factor

It is estimated that 65 per cent of the NH₃ applied is not integrated with the straw, and is therefore emitted after the treatment (Morken 2003). The same estimation is being used in

²² NILF (2010): Totalkalkylen for jordbruket. http://www.nilf.no/Totalkalkylen/Bm/2009/BMposter/BM_R_220B.shtml
211

Denmark.

5.7.2.4 Uncertainties

Uncertainty in the estimate of emissions from NH₃ treatment of straw is rather low (± 5 per cent) (Rypdal & Zhang 2001).

5.7.2.5 Completeness

Major missing emission components are not likely.

5.7.2.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

6 WASTE (NFR sector 5)

NFR 5

6.1 Overview

This sector includes solid waste disposal on land (5A), other biological treatment of waste (5B), waste incineration (5C), waste water handling (5D), and other waste (5E).

Emissions from waste incineration included in sector 5C are emissions from flaring, except flaring from energy sectors (included in NFR 1 energy), and emissions from cremation and hospital waste (until 2005). The main emissions from Waste Incineration are included in the energy sector (1A) since most of incineration of municipal, industrial and medical waste in Norway is now done with energy recovery. The source sector 5E Other Waste covers emissions from municipal sewage sludge applied to parks etc., emissions from accidental car fires, building fires, and emissions from recovering processes in the waste trade.

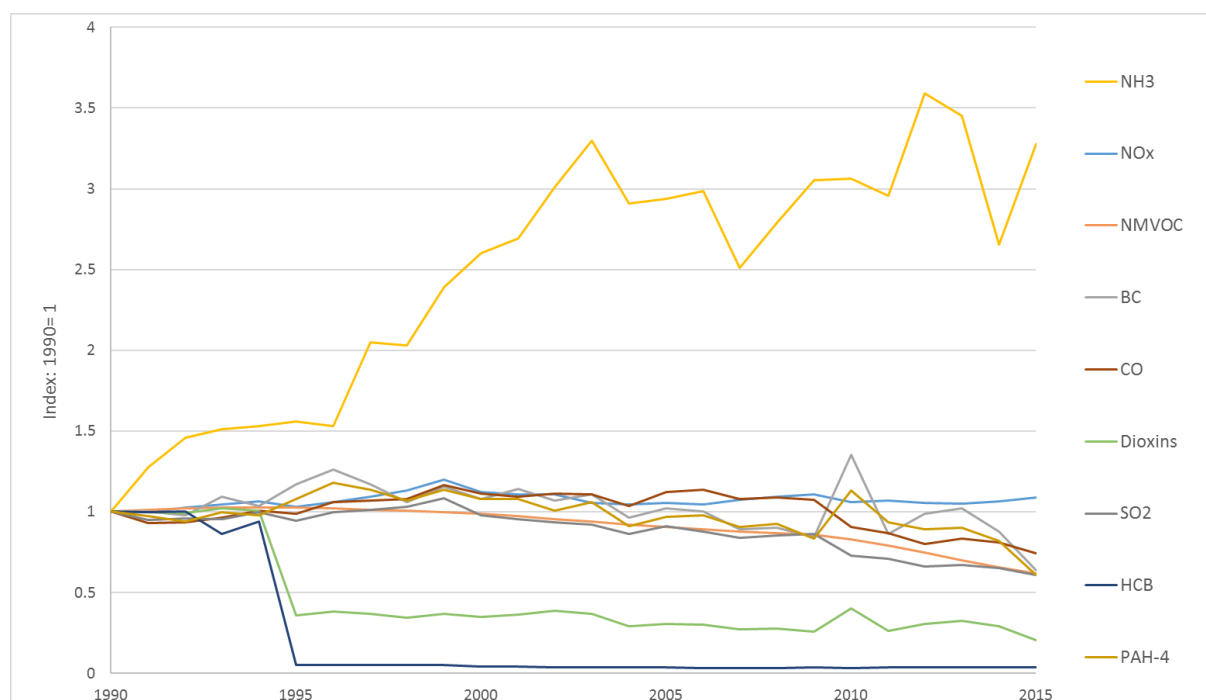


Figure 6.1. Trends for the emissions for most of long-range transboundary air pollutants from waste, relative to 1990

Source: Statistics Norway/ Norwegian Environment Agency

Figure 6.1 shows the trends for the emissions for most of the long-range transboundary air pollutants from waste, relative to 1990. With the exception of NH_3 and NO_x , the emissions of all pollutants have decreased since 1990.

6.2 Solid waste disposal on land

NFR 5A

Last update: 10.01.17

6.2.1 Description

This category is mainly a source of greenhouse gas emissions. Emissions of NMVOC and particulate matter are included in this inventory. Small quantities of CO and NH₃ may be released as well but are considered insignificant and are not estimated in this inventory.

6.2.2 Method

The emissions of NMVOC and particulate matter from solid waste disposal are being calculated in accordance with a Tier 1 approach (EEA 2016) using equation (6.1):

$$(6.1) \quad E_{\text{Pollutant}} = AR_{\text{production}} * EF_{\text{Pollutant}}$$

Where:

$E_{\text{Pollutant}}$ = emission (E) of pollutant

$AR_{\text{production}}$ = activity rate (AR), mass of landfilled waste

$EF_{\text{Pollutant}}$ = emission factor (EF) for pollutant

The emission factors for TSP, PM₁₀ and PM_{2.5}, extracted from (EEA 2016), are shown in the Table 6.1.

Table 6.1. Emission factors for biological treatment of waste. kg/tonnes

TSP	PM ₁₀	PM _{2.5}
0.463	0.219	0.033

Source: EEA (2016)

NMVOC

Small quantities of NMVOC are also emitted. US Environmental Protection Agency (US EPA) evaluates that 98.7 % of the landfill gas is methane and 1.3 % are other VOCs such as perchlorethylene, pentane, butane, etc. (EEA 2016). NMVOC have then been estimated assuming being 1.3 % of the landfill gas. Landfill gas is estimated using IPCC methodology (IPCC 2006).

6.2.3 Activity data

Data over the annual amount of waste deposited is taken from Statistics Norway's waste accounts.

Data over the amount of methane formed by decomposition of biological waste in landfills is taken from Statistics Norway's estimation of methane at MSWDS.

6.3 Compost production

NFR 5B

Last update: 19.02.16

6.3.1 Description

This category covers emissions from the biological treatment of waste: composting. Emissions of NH_3 and CO from home composting and emissions of NH_3 from industrial composting are included in the inventory. This source category is not considered to be significant in Norway in terms of long-range transboundary air pollutants. It can also be a source of NMVOC emissions which are not estimated in the Norwegian inventory.

6.3.1.1 Methodological issues

Emissions of NH_3 from composting of municipal waste have been calculated according to the Tier 2 default methodological guidance given by the 2016 Guidebook (EEA 2016).

6.3.1.2 Activity data

All Norwegian waste treatment plants are obligated to statutory registration and reporting of all waste entering and leaving the plants. All waste streams are weighed, categorized with a waste type and a type of treatment. Data is available for all years since 1995 and for the year 1992. Activity data for the year 1992 and since 1995 are collected from Statistics Norway's, waste statistics. For the years 1990 and 1991, activity data for 1992 are used, while AD for 1993 and 1994 are estimated by linear interpolation of activity data from 1992 and 1995.

Home composting

Emissions from home composting of garden waste and vegetable waste are also included in this inventory. The activity data for this category is available from Statistics Norway for the years 2009-2012. The amount of organic waste from households composted in the period 1990-2008 has been estimated assuming that 3 per cent of all households compost their garden and vegetable food waste (Lystad 2005). The average value of the period 2009-2012, 2.6 per cent, has been used for the following period.

6.3.1.3 Emission factors

Emissions from composting, will depend on both the composition of waste composted, amount and type of supporting material (such as wood chips and peat) used, temperature, moisture content and aeration during the process.

Table 6.2 gives default factors for CO and NH_3 emissions from biological treatment for Tier 2 method used for the estimation of Norwegian emissions.

Table 6.2. Composting emission factors. kg/tonnes

	CO	NH_3
Compost production	NE	0.24
Home composting	0.56	0.66

Source: EEA (2016)

6.4 Waste incineration

NFR 1A1a, 1A2d and 5C

Last update: 10.01.17

6.4.1 Description

In this chapter, the focus will be on waste from flaring, except flaring from energy sectors, and emissions from cremation and hospital waste until 2005. Since 2006, hospital wastes are incinerated in incinerators for municipal wastes and are included in the energy sector.

Emissions from waste incineration in district heating plants are reported under energy (NFR 1A1a), as the energy is utilised, and therefore described in section 3.2.2. In 2015, there were 17 waste incineration plants where household waste was incinerated. In addition, some incineration plants burn waste other than household waste, mainly wooden waste, paper, pasteboard and cardboard. These emissions are reported and described under energy. Waste, other than household waste, is also used as energy source in some manufacturing industries. These emissions are reported and described in the relevant subsectors under 1A2. Flaring off-shore and in refineries is included under sector 1B2c, Flaring in chemical industry is included under sector 2B5.

In Norway, the open burning of private yard waste is under different restrictions according to the respective municipality. These restrictions involve what can be burned, but also the quantity, how, when and where. In some municipalities, a complete ban is imposed. There is no registration of private waste burning and activity data on this subject are difficult to estimate. Citizens are generally encouraged to compost their yard waste or to dispose of it through one of the many waste disposal/recycling sites. Emissions from open burning of waste are not estimated.

PCB containing material are sent abroad, mostly to Sweden, to be destroyed. There is no incineration of PCB in Norway.

6.4.2 Method

Emissions from flaring of landfill gas are estimated by multiplying the amount of gas flared with the emission factors shown in Table 6.3. Emissions from flaring of biogas from industrial waste water treatment plants are estimated. Emissions have been estimated by multiplying the amount of gas flared with the emission factors shown in Table 6.3.

Emissions from cremation and hospital waste are estimated by emission factors multiplied with activity data. For hospital waste, the emissions of lead, cadmium and mercury used in the model are reported to the Norwegian Environment Agency. Emissions of arsenic, chromium and copper have only been reported by two hospitals to the Norwegian Environment Agency for the year 1999. A country specific emission factor has been estimated for each component. This factor is based on the ratio between reported emission figures for 1999 and the quantities of waste burned in 1999. This factor is then multiplied with the amount of waste burned at other hospitals for the years 1995 to 2005. Around 1995, more control device systems were installed

at waste incineration plants as a result of stricter emission requirements. It is assumed that this also applied to incineration of hospital waste. For the years before 1995, it is assumed that the emissions were higher. The emission standards for particulate matter from waste incineration changed from 100 to 30 mg/Nm³. It was assumed that emissions of lead, cadmium, copper and chromium followed the same pattern as particulate matter. It is believed however, that arsenic and mercury has similar properties and it has thus been assumed that emissions of arsenic have been reduced in the same way as mercury. Emissions of mercury were regulated from 0.1 to 0.05 mg/Nm³ from 1994 to 1995. It is therefore assumed that emissions of arsenic before 1995 were twice as large as after 1995. Emissions of particulate matter were reported for all hospitals for the period 1990-1999. Since 2000, emissions from hospitals incinerators have been estimated based on EF and the amount of waste incinerated. Since 2006, all hospital waste has been incinerated at waste incineration plants.

6.4.3 Activity data

Landfill gas

Information on the amount flared is given by the operators of landfills to the Norwegian Environment Agency. Emissions from landfill gas flared is included in 5C. Emissions from landfill gas used for district heating and used in other sectors are reported in the relevant subsectors under 1A1 and 1A4.

Biogas

The amount of biogas flared at some industrial waste water treatment plants are reported to the Norwegian Environment Agency for all years since 1991.

Natural gas

The amount of natural gas flared by the production of methanol is reported under 2B5.

Hospital waste

The amount of hospital waste was reported to Statistics Norway by some hospital incinerators. The hospital incinerators have gradually been closed down, mainly due to new emission limits. Since 2006, no hospital incinerators have been in operation. Nowadays, hospital waste is incinerated in incinerators for municipal waste and emissions are included under 1A1a.

Cremation

Incineration of human bodies is a common practice that is performed on an increasing part of the deceased. The number of cremated bodies is gathered by the Ministry of Culture and published in Statistics Norway's Statistical Yearbook.

6.4.4 Emission factors

Table 6.3 presents emissions factors for the waste incineration sector.

Table 6.3. Emission factors for flare of landfill gas, cremation and hospital waste incineration

Component	Flare landfill gas and biogas kg/tonnes	Cremation Tonnes/body	Cremation Tonnes/body CS EF ¹ 2007->	Hospital waste Tonnes/tonnes
SO ₂	0.02	0.000113		0.0014
CO	0.04	0.00014		0.0028
NO _x	0.17	0.000825		0.0014
Particles PM ₁₀	0.14	0.0000347	3.15504E-06	0.0005
TSP		0.00003856	3.506E-06	0.0005
PM _{2.5}		0.000031	2.81862E-06	0.0005
BC	7% of PM _{2.5}	50 % of PM _{2.5}	50 % of PM _{2.5}	18% of TSP
OC		36 % of PM _{2.5}	36 % of PM _{2.5}	
NM VOC	0	0.000013		0.0007
	g/tonne	kg/body		mg/tonne
Pb	NA	0.00003003	2.73042E-06	Plant-specific emission factors
Cd	NA	0.00000503	4.57344E-07	Plant-specific emission factors
Hg	NA	0.00149	5.59943E-05	Reported
Cu	NA	0.00001243	1.13018E-06	2594.6*
Cr	NA	0.00001356	1.23292E-06	1272.4*
As	NA	0.00001361	1.23747E-06	4705.6
Dioxins	NA	2.7E-11**		0.29685***
PCB	NA	4.1E-07		0.39*
HCB	NA	1.5E-07		2.6*
benzo(a)pyrene	NA	1.32E-08		0.004179
benzo(b)fluoranthene		7.21E-09		0.035821
benzo(k)fluoranthene	NA	6.44E-09		..
indeno(1,2,3-cd)pyrene	NA	6.99E-09		..

NA=Not Applicable.

¹ Country specific emission factor based on measurements of Hg and TSP for the years 2013-2015. EFs for all HM are reduced as much as TSP (91%). The new emission factors are used for all years since 2007.

* Country specific emission factor used for the years after 1995. Emission factors for the years 1990 to 1994 can be given on request.

** Emissions factor is given in kg I-TEQ/body

*** Emissions factor is given in mg I-TEQ/tonne

Source: EEA (2016), Kupiainen and Klimont (2004) and Danish IIR (Aarhus University, 2016)

BC emissions have been estimated using shares of PM_{2.5} as emission factors. Shares given by IIASA (Kupiainen & Klimont 2004) have been used. For cremation, as no share for BC was found in the literature for the use of natural gas in navigation, BC share has been set to be 50 per cent of PM_{2.5}. Indeed, the amount of PM_{2.5} is assumed to be equally shared between BC and organic mass (OM).

6.4.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

6.4.5.1 Activity data

No data on amounts of hospital waste have been reported since 1999. The amount of hospital waste the subsequent years may vary from the data reported in 1998 and 1999. Since 2006, no

hospital incinerators have been in operation.

6.4.5.2 Emission factors

The composition of the hospital waste could be different from the waste the emission factors are based on. In that case, the calculated emissions will be incorrect. Combustion engineering and processes also influence the emissions. These uncertainties have not been calculated.

6.4.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

6.5 Waste water handling

NFR 5D

Last update: 11.01.17

6.5.1 Description

This category covers emissions from the biological treatment of waste water and latrines. This source category is not considered to be significant on the Norwegian level in terms of long-range transboundary air pollutants.

Emissions of NMVOC are estimated in the Norwegian inventory. Emission factors for all other pollutants are not available and may be assumed negligible in most cases.

6.5.2 Method

The emissions of NMVOC from waste water treatment are being calculated in accordance with a Tier 1 approach (EEA 2016) using equation (6.2):

$$(6.2) \quad E_{\text{Pollutant}} = AR_{\text{production}} * EF_{\text{Pollutant}}$$

Where:

$E_{\text{Pollutant}}$ = emission (E) of pollutant

$AR_{\text{production}}$ = activity rate (AR), amount of waste water

$EF_{\text{Pollutant}}$ = emission factor (EF) for pollutant

The emission factors for NMVOC is given in EEA (2016). The emission factors used is 15 mg NMVOC/m³ waste water.

6.5.3 Activity data

Domestic waste water

Total amount of waste water handled by all waste water treatment plants in the country is taken from Statistics Norway's municipal water supply for the years after 2009. For the years 1990 to 2008, the amount of waste water is estimated based on the part of the population connected to treatment plants, using equation (6.3).

$$(6.3) \quad \text{Waste water} = \text{Population} \times \text{NR}_{\text{PEOPLE}} \times \text{EF}$$

Where:

$\text{NR}_{\text{PEOPLE}}$: share of people connected to treatment plants

EF : emission factor (average household consumption per person per year)

Norwegian population data are extracted from Statistics Norway's population statistics. Data for the number of people in Norway connected to waste water treatment plants are extracted from Statistics Norway's waste water statistics. Data for the average household consumption per person per year (2002-2008) are extracted from Statistics Norway's statistics on municipal water supply. Varies between 70-76 m³ water/inhabitant/year. The number for 2001 have been used for all years 1990-2001.

Industrial waste water

The amount water released into recipient is reported by industries to the Norwegian Environment Agency (pulp and paper industry, chemical industry and food processing industries).

6.6 Other emission sources from the waste sector

NFR 5E

Last update: 18.02.16

6.6.1 Description

This category is a catch all for the waste sector. In the Norwegian inventory, emissions from sewage sludge applied on fields other than agricultural soils, accidental car fires, house fires and emissions from recovering processes in the waste trade are included in this category.

6.6.2 Method

6.6.2.1 Sewage sludge applied on fields

NH₃

Emissions of NH₃ are calculated for sewage sludge applied on fields other than agricultural soils. To calculate NH₃ emissions from sewage sludge, the fraction of N in manure lost as NH₃ is used (frac_{GASM}). The loss equals to total N in sewage sludge multiplied by frac_{GASM}. See section 5.5.2.1.3.

6.6.2.2 Car and house fires

Particles, heavy metals and POPs

Emissions of particles, heavy metals, PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene) and dioxins are calculated for car fires and house fires. In addition, SO₂, NO_x, NMVOC and CO are calculated for car fires. Emissions are calculated by multiplying the annual number of car and house fires with emission factors. Four

types of buildings are separated with different emission factors: detached houses, undetached houses, apartment buildings and industrial buildings.

6.6.2.3 Waste trade

NH₃, particles, heavy metals and POPs

Emissions from recovering processes in the waste trade include emissions of NH₃, particles, heavy metals (As, Cd, Cr, Cu, Hg, Pb), and PAHs (benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and indeno(1,2,3-cd)pyrene). The emission figures are reported annually by the actual plants to the Norwegian Environment Agency.

6.6.3 Activity data

6.6.3.1 Sewage sludge applied on fields

Statistics Norway's waste water statistics annually gives values for the amount of sewage sludge and the fraction of the sewage sludge that is applied on fields.

6.6.3.2 Car and house fires

Data on the number of car and house fires are provided annually by the Directorate for Civil Protection and Emergency Planning. These figures only include fires reported to the fire service.

6.6.4 Emission factors

6.6.4.1 Sewage sludge applied on fields

The N-content in the sludge is given in Statistics Norway (2001), and the same value of 2.82 per cent is used for all years.

6.6.4.2 Car fires

The emission factor for particles is given by EPA (2002). EPA recommends the factor of 0.9 kg/car for combustion of wrecked cars without car tyres, and a factor for combustion of car tyres of 1.4 kg/car. This results in an overall emission factor of 2.3 kg/car. The emission factor for dioxins emissions from car fires is found in Hansen (2000). Emissions factors for heavy metal and PAHs from car fires is found in the Danish IIR (Aarhus University, 2016). Emission factors for mercury from car fires is found in the French IIR (CITEPA 2016). No data are available for HCB and PCBs. NH₃ is assumed not to be emitted. It is difficult to estimate the amount of material burned in a car fire. It is assumed that the average weight of a car is 1383 kg, average weight loss is assumed to be 18,2 per cent or 252 kg (CITEPA 2016). Emissions from vehicle fires are calculated by multiplying the mass of vehicle fires with selected emission factors. Emission factors are not available for different vehicle types, whereas it is assumed that all the different vehicle types leads to similar emissions.

6.6.4.3 House fires

It is difficult to estimate the amount of material burned in a house fire. In Finstad *et al.* (2002b) a calculation was made that has been used to scale the chosen emission factors, to reflect how much of the building that is lost in a fire. This scaling calculation is based on the amount of damage estimated in monetary value, and value on how much of the building and the furniture that is burned. The emission factors used for particles in the inventory are given by scaling the

emission factors used for combustion of fuelwood in the households (Haakonsen and Kvingedal 2001). The emission factors for heavy metals are given by scaling the emission factors for combustion of wood waste in the industry (EPA 2002). For dioxins, OSPAR (Norwegian pollution control authority 2001) gives the emission factor of 170 µg I-TEQ per tonne burned material. Emissions factors for PAHs is found in Danish IIR (Aarhus University 2016). The scaled emission factors used for the different building types are given in Table 6.4.

Table 6.4. Emission factors used for car fires and house fires, unit/fire

	Car	Detached house	Undetached house	Apartment building	Industrial building
SO ₂ (tonnes)	0.0013	NE	NE	NE	NE
NO _x (tonnes)	0.0005	NE	NE	NE	NE
NM VOC (tonnes)	0.0021	NE	NE	NE	NE
CO (tonnes)	0.016	NE	NE	NE	NE
TSP (tonnes)	0.0023	0.14382	0.06162	0.04378	0.02723
PM ₁₀ (tonnes)	0.0023	0.14382	0.06162	0.04378	0.02723
PM _{2.5} (tonnes)	0.0023	0.14382	0.06162	0.04378	0.02723
BC	NE	9 % of PM _{2.5}	9 % of PM _{2.5}	9 % of PM _{2.5}	9 % of PM _{2.5}
Pb (kg)	0.206	0.00042	0.00018	0.00013	8E-05
Cd (kg)	0.0004	0.00085	0.00036	0.00026	0.00016
Hg (kg)	0.0001	0.000087	0.000037	0.000026	0.000016
As (kg)	6.5E-05	0.00135	0.00058	0.00041	0.00025
Cr (kg)	0.00096	0.00129	0.00055	0.00039	0.00024
Cu (kg)	0.0067	0.00299	0.00128	0.00091	0.00057
benzo(a)pyrene (kg)	0.0037	0.008	0.0064	0.0037	0.0096
benzo(b)fluoranthene (kg)	0.0041	0.0126	0.0101	0.0059	0.0152
benzo(k)fluoranthene (kg)	0.0041	0.0044	0.0036	0.0021	0.0054
indeno(1,2,3-cd)pyrene (kg)	0.0059	0.0086	0.0069	0.004	0.0104
Dioxins (mg)	0.048	1.43817	0.61621	0.43779	0.27234

Source: Statistics Norway, Danish IIR (Aarhus University, 2016) and French IIR (2016)

6.6.5 Uncertainties

Uncertainty estimates for long-range transboundary air pollutants are given in Appendix C.

6.6.6 Source specific QA/QC

There is no source specific QA/QC procedure for this sector. See section 1.6 for the description of the general QA/QC procedure.

7 Other and Natural emissions

There is no long-range transboundary air pollution reported for Norway as Other or Natural emissions.

8 Recalculations and Improvements

8.1 Recalculations

8.1.1 Overall description of the recalculations for the long-range transboundary air pollutants

As part of the continual process of improving the emission estimates, the Norwegian emission inventory has been recalculated. The process involves correcting discovered errors and utilising new or improved information where this has become available. The entire time series, at present 1990-2014, are recalculated when the method for a certain source category is revised. The figures in the inventory are therefore, as far as possible, consistent through the whole time series.

The most important recalculations in the 2017 submission are:

1. Emissions from road traffic have been updated due to revised emission factors. NMVOC and CO has decreased throughout the period. NO_x has decreased for most years, but increased for 2011-2014.
2. NMVOC og NO_x from agriculture (3B manure management systems and 3D agricultural soils) have been included in the emission inventory according to EMEP guidelines 2013.
3. Emissions of PM_{2,5}, PM₁₀ and TSP from agriculture have been revised according to EMEP guidelines 2016. Emissions from agricultural operations previously defined under source 3I are now defined under source 3D Agricultural soils. Emissions of PM from 3B (manure management systems) are included for the first time.
4. Emissions of CO from production of aluminium have been included for all years 1990-2014. This has led to an increase in emissions of CO by 13.9 to 57.8 per cent.
5. PAH-4 have been split into benzo(a)pyren, benzo(b)fluoranthen, benzo(k)fluoranthen and indeno(1,2,3_cd)pyrene. Emission factors used for estimation of emissions of benzo(a)pyren, benzo(b)fluoranthen, benzo(k)fluoranthen and indeno(1,2,3_cd)pyrene are taken from EEA Guidebook 2013 and 2016, Finstad et al (2001) and Danish IIR (Aarhus University 2016). The PAH profile has been measured at some aluminum producing plants. This has also led to updated distribution of PAH-4 emissions from aluminum production.
6. Emissions of SO₂, NO_x, NH₃, NMVOC, CO and heavy metals from car fires have been included. This has led to increased emissions. The increase for 2014 in emission of lead, chromium and copper were 153, 0.7 and 5 kg respectively.

In combination with some minor changes from other sources, the recalculations have caused several changes in the emission figures, see Table 8.6, Table 8.7 and Table 8.8.

8.1.2 Specific description of the recalculations

8.1.2.1 Energy

Energy figures for the last year in the previous submission (2013) have been extensively revised,

because the energy figures for 2013 used in the previous submission were preliminary. There will always be some changes in the energy figures for the last year, e.g. some figures on energy use in manufacturing industries will be adjusted, which will lead to adjustments in other sectors, as total use of oil products in the energy accounts sum up to national sales of petroleum products. Now the final figures for energy use are available and are used in the emission calculations. Changes in emission figures caused by such revisions will not be commented on specifically under each NFR code.

1A1a Public electricity and heat production

- Consistency. For one plant, there are minor changes in SO₂ figures 2011-2014, due to replacement of previously used 2010 figures with reported figures for the actual years.
- Revised activity data. Minor changes due to revisions in figures on waste combustion 2012-2014.
- Revised activity data. For one plant reported As emissions for the years 1997, 1999 and 2011 to 2014 have replaced the reported figures from 2004 which was used earlier. These changes have led to a small increase in 1997, and decreases for all other years.

1A2a Iron and steel

- Correction of error. Small reductions in emissions of SO₂ and NO_x from one plant in 2013 due to an error in the previous calculations.

1A2c Chemicals

- Correction of error. Small reductions in emissions of NO_x from one plant in 2011 due to an error in the previous calculations.

1A2f Other manufacturing

- Reported emissions of Pb from glass production are assumed to include both process and combustion emissions. Combustion emissions of lead from glass production have been removed.

1A3B_ Road transport

- Revised emission factors. A revised version of the Handbook of Emission Factors (HBEFA) has been taken into use. The new version contains updated hot emission factors for several emission components for the years 1990-2014. As a consequence of the update, the NO_x emissions from passenger cars (1A3bi) are higher, especially Euro-5 and Euro-6 diesel vehicles. The emissions of NMVOC and CO from road transport are lower as a consequence of the update. Emission factors for mopeds are also used for snow scooters. Thus, the revision also affects source categories 1A4b-ii and 1A4c-ii.

1A3bvii Road transport: Automobile road abrasion

- Correction of error. Minor changes in emissions of cadmium, particulates and PAH in 2014 due to errors in the previous calculation.

1A3d National navigation

- Accuracy. The emission factor for NO_x was updated for the years 2008-2014, and this resulted in an important increase in emissions from national navigation. The previously

used emission factors overestimated the effect of economic incentives to reduce emissions. The economic incentives to reduce NO_x-emissions from ships started in 2008, therefore it was decided to keep the “old” emission factor for 2007 and update the emission factors for the years 2008 and onwards. The update was based on reported figures of energy use and NO_x-emissions from as good as all relevant ships for a period of three months in 2016. The emission factor obtained is assumed to be of high quality. A linear interpolation was chosen to estimate emission factors for the years between 2007 and 2016. The changes also affect emissions from fishing (1A4c-iii).

1A4bii and 1A4cii Residential and Agriculture/forestry/fishing – Off-road vehicles

- Revised emission factors: Changes for road transport (see 1A3b above) also affects emissions from snow scooters reported in several categories under 1A4.

1A4ciii Fishing

- Accuracy: Changes in NO_x emission factors for navigation also affects fishing. See 1A3d above.

1A5a Other stationary (including military)

- Consistency. Military use of natural gas in 2009 has been reallocated from boiler to ship, in accordance with the practice for other years. This causes emission reductions in 1A5a.

1A5b Other, Mobile (including military, land based and recreational boats)

- Consistency. Military use of natural gas in 2009 has been reallocated from boiler to ship, in accordance with the practice for other years. This causes emission increases in 1A5b.

1B2ai Oil and natural gas: Transport

- Revised emission data. Increase in NMVOC emissions 2013 due to revised reported figure from one field.
- Revised emission data. Higher NMVOC emissions 2014 due to revised reported figures from many fields.
- Reallocation. 78 tonnes NMVOC reallocated from 1B2c to 1B2ai 2010-2014.

1B2av Oil and natural gas: Distribution of oil products

- Revised emission data. Minor changes in emissions of NMVOC 2013-2014 due to revisions in reported emission data.

1B2c Oil and natural gas: Venting and flaring (oil, gas, combined oil and gas)

- Revised emission data. Reported emission figures for NMVOC from two fields have been adjusted somewhat downwards for 2014.
- Reallocation. 78 tonnes NMVOC reallocated from 1B2c to 1B2ai 2010-2014.

8.1.2.2 Industrial processes

2A3 Glass production

- Correction of errors. Reported emission of As have been removed for the years 1990-1992 (43 kg) and 2005-2014 (0,05 kg).

2A5b Construction and demolition

- Revised emission factor. Emission factor from Guidebook 2016 have replaced earlier used emission factors for estimation of emissions of particulate matter from sand pits and rock-crushing plants and for building and construction. Resulted in a decrease of PM_{2.5} emissions of about 0.7 to 1.7 per cent.

2A6 Other mineral products

- Correction of error. For one plant, reported emission figures on SO₂ 2011-2014 replace previous zero registrations. Small figures 2011-2012, for 2013 and 2014 respectively 100 and 162 tonnes.

2A6 Other process use of carbonates

- *Revised emission data.* For one plant, reported emission figures on Cd and Cr for 2014 replace previous registrations. Emissions of Cd decreases by 0.03 kg and emissions of Cr decreases by 18.8 kg
- Revised emission factors. Emission factors for particulate matter from Guidebook 2013 replace previously used factors. This causes emission reductions in the order of 1 000 – 4 300 tonnes for TSP and 200 – 700 tonnes for PM₁₀, whereas emissions of PM_{2.5} have risen by 100 - 400 tonnes.

2B2 Nitric Acid Production

- Addition. Emissions of BC have been included for all years. Emissions of BC have increased by 0.1 to 0.3 per cent.

2B5 Calcium carbide

- Addition. Emissions of BC have been included for all years. Emissions of BC have increased by 0.1 per cent or less.

2B6 Titanium dioxide production

- Addition. Emissions of BC have been included for all years. Emissions of BC have increased by 0.1 per cent or less.

2B6 Production of plastic

- Addition. Emissions of BC have been included for all years. Emissions of BC have increased by less than 0.1 per cent.

2B10A Ethylene dichloride and vinyl chloride monomer

- Correction of error. Emission increase for HCB in 2005, due to the replacement of an erroneous figure. (1.1 gram has been replaced by 11 gram).

2B10 Chemical industry: Production of soap and panit and varnish production

- Addition. Emissions of BC have been included for all years. Emissions of BC have increased by less than 0.1 per cent .

2C2 Ferroalloys production

- Revised activity data. For three plants, revision of activity data used in the calculations has caused emission changes of NMVOC and dioxins 1990-2011 and 2013-2014.

2C3 Aluminium production

- Addition. Emissions of CO have been included for all years 1990-2014. Emissions of CO have increased by 13.9 to 57.8 per cent.
- Revised activity data. Revision of activity data used in the calculations has caused minor emission changes for NO_x in 2010, 2013 and 2014.
- Revised emission factor. The PAH profile has been measured at some aluminum producing plants. This has also led to updated distribution of PAH-4 emissions from aluminum production. Emissions of PAH-4 from aluminium production have been reduced by 1 to 16 tonnes. Largest reduction before 2005.
- Addition. Emissions of BC have been included for all years 1990-2014. Emissions of BC have increased by 0.2 to 0.6 per cent.

2H2 Food and beverages industry

- Revised emission factor. Emission factor from Guidebook 2016 have replaced earlier used emission factors used estimation of NMVOC emissions from production of bread and beer. Resulted in an increase of emissions of about 0.1 to 0.6 per cent.

2G Tobacco

- Addition. Emissions of NH₄, from use of tobacco have been included for all years 1990-2014.
- Correction of error. Emissions of Hg have been reduced. The emission factor has been changed from 0.0001 g/kg tobacco to 0.00001 g/kg tobacco.
- Correction of error. Emissions factors for Cu was earlier by a mistake used for estimation of emissions of Cr, and the emission factor for Cr was earlier by a mistake used for estimation of emissions of Cu. This mistake is now corrected. Emissions of Cr have decreased for all years (0.5 to 1.2 kg) and emissions of Cu have increased for all years (0.6 to 1.3 kg).

2H1 Pulp and paper industry

- Addition. Emissions of NO_x, NMVOC and CO from production of pulp and paper have been included for all years 1990-2014.

2I Wood processing

- Addition. Emissions of TSP from wood processing have been included for all years 1990-2014.

8.1.2.3 Agriculture

3B and 3D NH₃ emissions from manure management and agricultural soils

- Correction of error and revised activity data in the livestock population characterisation

Changes in activity data

Number of milk cows:

The figures for the years 2009-2014 have been revised. This has given a substantially higher numbers of cows, particularly for 2014 (see Table 8.1).

Table 8.1. Number of milk cows in the 2016 and the 2017 submissions

	2017 submission	2016 submission	Increase	Increase, per cent
2009	235 480	210 554	24 927	11.8
2010	232 294	209 094	23 200	11.1
2011	224 721	201 165	23 557	11.7
2012	229 767	203 592	26 175	12.9
2013	225 163	196 085	29 078	14.8
2014	221 032	177 759	43 272	24.3

Source: The cow recording system at TINE AS

Statistics Norway and the Cow recording system at TINE AS have assessed the data sources and data needs, and have established a cooperation to ensure that the correct data is used in the future.

Number of heifers for replacement

Also these figures have been revised. Table 8.2 shows the adjustment from the previous time series.

Table 8.2. Number of heifers

	2017 submission	2016 submission	Increase	Increase, per cent
1990	143 904	151 025	-7 121	-4.7
1991	142 210	149 240	-7 029	-4.7
1992	141 906	148 908	-7 001	-4.7
1993	140 179	147 085	-6 906	-4.7
1994	137 906	144 680	-6 774	-4.7
1995	138 359	145 140	-6 781	-4.7
1996	140 351	147 216	-6 865	-4.7
1997	137 684	144 394	-6 710	-4.6
1998	135 775	142 372	-6 597	-4.6
1999	137 441	144 100	-6 659	-4.6
2000	129 500	135 725	-6 224	-4.6
2001	127 008	133 092	-6 085	-4.6
2002	124 347	130 297	-5 949	-4.6
2003	123 763	129 668	-5 905	-4.6
2004	117 946	122 011	-4 065	-3.3
2005	118 090	123 006	-4 916	-4.0
2006	114 306	118 408	-4 101	-3.5
2007	109 011	112 974	-3 963	-3.5
2008	110 954	110 954	0	0.0
2009	114 416	109 286	5 131	4.7
2010	111 122	109 150	1 973	1.8
2011	110 311	108 516	1 796	1.7
2012	108 455	106 679	1 777	1.7

2013	109 643	107 650	1 993	1.9
2014	109 115	99 556	9 559	9.6

Data are available from 2004. The numbers of heifers 1990-2003 are estimated on basis of the relationship between number of heifers 2004-2011 and the number of milk cows; hence the whole time series is adjusted.

Number of sheep

In the calculations of emissions from sheep manure, the sheep population is divided in sheep over one year and sheep under one year. The method for estimating the population was revised in 2016, see description in chapter 5.2.3. The difference in estimated numbers compared to previous figures is shown in Table 8.3.

Table 8.3. Number of sheep in the 2016 and 2017 submission

	Sheep > 1 year			Sheep < 1 year		
	2017 submission	2016 submission	Increase, per cent	2017 submission	2016 submission	Increase, per cent
1990	714 384	1 028 867	-30.6	622 862	519 541	19.9
1991	747 530	1 082 391	-30.9	631 730	534 781	18.1
1992	746 691	1 094 910	-31.8	638 394	550 021	16.1
1993	689 097	1 021 346	-32.5	611 272	540 422	13.1
1994	771 234	1 119 823	-31.1	673 584	576 623	16.8
1995	783 922	1 138 821	-31.2	683 599	592 999	15.3
1996	771 837	1 107 716	-30.3	682 781	596 173	14.5
1997	735 326	1 076 217	-31.7	651 839	568 317	14.7
1998	748 821	1 100 478	-32.0	641 322	546 808	17.3
1999	749 076	1 102 068	-32.0	629 186	622 551	1.1
2000	766 098	1 129 458	-32.2	643 141	639 746	0.5
2001	793 551	1 138 073	-30.3	662 425	654 334	1.2
2002	770 379	1 104 909	-30.3	665 456	641 394	3.8
2003	776 557	1 111 513	-30.1	688 644	668 572	3.0
2004	726 018	1 083 216	-33.0	697 210	477 713	45.9
2005	717 098	1 057 911	-32.2	685 466	467 648	46.6
2006	678 520	1 002 006	-32.3	663 761	466 899	42.2
2007	672 788	1 019 998	-34.0	652 250	448 634	45.4
2008	678 462	1 035 624	-34.5	647 166	447 373	44.7
2009	692 873	1 061 636	-34.7	663 630	464 270	42.9
2010	691 450	1 054 092	-34.4	659 895	461 592	43.0
2011	687 440	1 050 191	-34.5	653 663	480 506	36.0
2012	685 326	1 048 699	-34.6	634 325	463 456	36.9
2013	689 345	1 042 449	-33.9	642 338	458 413	40.1
2014	683 479	1 058 705	-35.4	676 867	476 015	42.2

Changes in emissions

Milk cows

About one quarter of total NH₃ emissions in 2014 was caused by storing and spreading of manure from milk cows. An increase in estimated number of cows of almost 25 per cent in 2014 means that the correction in number of milk cows alone led to an increase in estimated NH₃ emissions of about 6 per cent. The estimated increase in the other years the number of cows was adjusted, was substantially lower.

Heifer for replacement

About 6 per cent of total NH₃ emissions in 2014 were caused by storing and spreading of manure from heifer for replacement. An increase in estimated number of heifer of almost 10 per cent in 2014 means that the correction in number of heifers alone led to an increase in estimated NH₃ emissions of about 0,6 per cent.

Sheep

About 8 per cent of total NH₃ emissions in 2014 were caused by storing and spreading of manure from sheep. Number of sheep under one year increased about 40 per cent in 2014, while the number over one year went down about 35 per cent. The total emissions from sheep decreased due to this revision. The emissions of NH₃ were about 0.4 per cent lower compared to what the emissions would have been with the animal figures used in the 2016 submissions. Emissions in the years before 2004 decreased slightly more due to the revision of number of sheep.

Slaughter age for bulls

The slaughter age for bulls was reduced 2004-2014 with 0.4-0.8 months. In 2014 the reduction was 0.75 months. The emissions was reduced with 37 tonnes NH₃, corresponding 0.14 percent of the total emission in 2014.

3B and 3D NO_x and NMVOC emissions from manure management and agricultural soils

- Addition. NO_x and NMVOC emissions are included in the emission estimates from agriculture for the first time. For the year 2015, the emissions from agriculture made out 4.2 and 7.3 per cent of total national emissions. Sources and methods are described in chapter 5.4.3, 5.4.4, 5.5.3 and 5.5.4.

3B and 3D Particulate matter (PM₁₀, PM_{2.5} and TSP) from manure management and agricultural soils

- Addition and revised emission factors and activity data. There are three changes in the estimations of particulate emissions from agriculture:
 - Emissions from animal manure are included in the emission estimates for the first time.
 - The emission factors from agricultural operations are changed and increased substantially.
 - Activity data for emissions from agricultural operations is changed from grain area to open fields and gardens

Emission factors particulate matter, other agricultural operations.

Table 8.4 shows the changes in emission factors from previous submission.

Table 8.4. Emission factor for emissions of particulate matter from agricultural operations. kg per hectare

	2016 submission¹	2017 submission
PM _{2.5}	0.00292 ¹	0.06
PM ₁₀	0.00292 ¹	1.56
TSP	0.00292 ¹	1.56

¹ Aggregate of factors for the three sources that make out "agricultural operations"

Changes in activity data

Activity data is changed from area of grains to area of open fields and gardens. Table 8.5 shows old and new figures.

Table 8.5. Activity data used for estimating emissions of particulate matter from agricultural operations. 1000 hectares

	2016 submission: grain area	2017 submission: area of open fields and gardens
1990	356.7	438.9
1991	369.1	434.5
1992	359.4	430.1
1993	352.7	425.7
1994	348.6	421.4
1995	338.1	417.0
1996	334.1	412.6
1997	332.1	408.2
1998	335.6	403.8
1999	323.7	399.5
2000	326.8	398.4
2001	321.1	399.7
2002	321.0	391.4
2003	323.1	385.4
2004	323.9	384.5
2005	321.3	380.4
2006	314.7	373.1
2007	309.1	366.9
2008	306.4	360.9
2009	304.3	356.6
2010	300.3	353.6
2011	298.0	347.6
2012	292.7	343.4
2013	286.1	335.0
2014	282.7	332.6

PM emissions from animal manure

Emissions of PM from animal manure is included for the first time, and is estimated according to EMEP/EEA guidebook 2013. These emissions constitute about 80 per cent of PM₁₀, 96 per cent of PM_{2.5} and 85 per cent of TSP from agriculture.

8.1.2.4 Waste

5A Managed Waste Disposal sites

- Addition. Emissions of NMVOC and particulate matter have been included for all years. NMVOC emissions from managed waste disposal sites represent more than 99 per cent of the total NMVOC of the waste treatment sector (NFR 5).
Particulate matter emissions from managed waste disposal sites represent between 0,4 and 1 per cent of total particulate matter emissions from of the waste treatment sector (NFR 5) during the whole period.

5B1 Home composting of garden and vegetable food waste

- Addition. Emissions of CO and NH₃ have been included for all years. CO emissions from composting represents between 11 and 27 per cent of total CO emissions from the waste treatment sector during the whole period. NH₃ represents between 5 and 32 per cent of total CO emissions from the waste treatment sector during the whole period.

5C1BIII Clinical waste incineration

- Correction of emission factor. HCB emission factor for all years after 1995 has been corrected. HCB emissions of HCB for burning of hospital waste have increased by 31 per cent for all years 1995 to 2005.
- Correction of emission factor. PCB emission factor for the years 1990 to 1994 has been corrected. PCB emissions for burning of hospital waste have decreased by 31 per cent for all years 1990 to 1994.

5C1 Cremation

- Correction of emission factor. Hg and particulate matter emission factors has been corrected for all years. Emission factor from Guidebook 2016 has been used for the years 1990-2006. Since 2007, country specific emission factor based on measurements have been used.
Hg emissions have decreased by 70 per cent for the period 1990-2006 and by 99 per cent after 2007. Particulate matter ustlipp have been multiplied by around 1000 for the period 1990-2006 and by around 100 after 2007.

5D Domestic Wastewater

- Addition. Emissions of NMVOC have been included for all years 1990-2014 (3-5 tonnes). NMVOC emissions from waste treatment have increased by mellom 0.3 to 1.1 per cent during the period.

5 E Other Waste

- Correction of emission factor. House fires. The emission factor for Hg for all years has been corrected. By a mistake the emission factor for Cd was earlier used. Emissions of Hg have been reduced by around 40 per cent for the period.
- Addition of emission. Car fires. Emissions of SO₂, NO_x, NMVOC, CO, Pb, Cd, Hg, As, Cr, Cu and PAHs (benzo(a)pyren, benzo(b)fluoranthen, benzo(k)fluoranthen and indeno(1,2,3_cd)pyrene) have been included for all years. Total emissions of lead have increased by 0,2 to 5,3 per cent due to this inclusion. Minor increases for the other components.

8.1.3 Implications of the recalculations for long-range transboundary air pollutants

Implications for emissions levels

Table 8.6 shows the effects of recalculations on the emission figures for the main pollutants 1990-2013, Table 8.7 the effect on the PM emissions and Table 8.8 the effects on the POP and heavy metal emission figures.

Table 8.6. Recalculations in 2016 submission compared to the 2015 submission. Main pollutants

	SO ₂	NO _x	NMVOC	CO	NH ₃
	tonnes	tonnes	Tonnes	tonnes	tonnes
1990	3	7816	10610	78204	-593
1991	3	7815	10633	80784	-798
1992	3	7808	10791	79308	-696
1993	3	7579	10662	81101	-469
1994	3	7530	10900	79611	-666
1995	3	7583	10914	79259	-705
1996	3	7564	11196	82766	-554
1997	3	7455	11068	89715	-553
1998	3	7461	11183	100520	-527
1999	3	7227	11483	103798	-685
2000	3	7217	11206	107717	-761
2001	3	6958	10877	109951	-706
2002	3	6957	10983	112045	-625
2003	3	7258	11220	130198	-671
2004	3	7300	11358	147651	-494
2005	3	7406	11545	156835	-475
2006	3	7331	11566	157146	-500
2007	3	7451	11600	155568	-583
2008	3	8362	11770	157391	-596
2009	3	9444	11694	127536	348
2010	3	11018	12134	126948	153
2011	1	13403	12149	129793	109
2012	3	15022	11510	131521	337
2013	69	15776	13192	135033	155
2014	-141	17199	19822	135453	842

Source: Statistics Norway

Table 8.7 Recalculations in 2017 submission compared to the 2016 submission.

Particulate matter

	TSP	PM ₁₀	PM _{2.5}	BC
	tonnes	tonnes	Tonnes	Tonnes
1990	-4698	1243	358	58
1991	-3477	1418	452	55
1992	-2534	1538	493	38
1993	-3156	1441	476	45
1994	-1776	1709	338	45
1995	-2130	1716	322	42
1996	-2352	1726	316	39
1997	-2923	1588	336	37
1998	-3370	1527	259	35
1999	-3415	1546	278	29
2000	-3737	1521	328	31
2001	-3953	1470	297	37
2002	-3932	1506	306	37
2003	-3835	1516	366	24
2004	-3810	1615	370	24
2005	-5105	1452	347	23
2006	-4775	1469	385	15
2007	-6265	1382	372	15
2008	-5220	1665	317	7
2009	-4624	1721	406	-2
2010	-4149	1729	420	0
2011	-4686	1630	423	2
2012	-7460	1285	510	-34
2013	-5952	1632	426	-29
2014	-6049	1519	222	-65

Source: Statistics Norway

Table 8.8. Recalculations in 2017 submission compared to the 2016 submission.

POPs and heavy metals

	Lead	Cadmium	Mercury	Arsenic	Chromium	Copper	PAH-4	Dioxins	HCB	PCB
	Kg	Kg	Kg	Kg	Kg	Kg	Kg	mg	kg	kg
1990	286	1	-50	-42	0	11	-16326	83	0	-3
1991	249	1	-49	-42	0	10	-12690	63	0	-3
1992	247	1	-50	-42	0	10	-12917	62	0	-3
1993	262	1	-51	0	0	10	-13223	-19	0	-2
1994	276	1	-51	0	0	11	-11734	2	0	-3
1995	274	1	-49	0	1	10	-11153	-15	0	0
1996	305	1	-50	0	1	11	-12548	6	0	0
1997	315	1	-50	1	1	12	-14061	-22	0	0
1998	316	1	-49	0	1	12	-11822	2	0	0
1999	349	1	-51	0	1	13	-10625	-21	0	0
2000	329	1	-51	0	1	12	-9460	-13	0	0
2001	316	1	-51	0	1	12	-10359	-147	0	0
2002	323	1	-51	0	1	12	-12168	-136	0	0
2003	329	1	-49	0	1	12	-9296	-132	0	0
2004	288	1	-49	0	0	11	-10739	-146	0	0
2005	318	1	-50	0	1	11	-12500	-136	10	0
2006	327	1	-51	0	1	12	-7329	-161	0	0
2007	289	1	-72	0	1	10	-5389	-163	0	0
2008	295	1	-73	0	1	11	-5634	-151	0	0
2009	300	1	-74	0	1	11	-1607	-152	-1	69
2010	211	0	-74	0	0	8	-1450	-203	0	0
2011	192	0	-76	0	0	7	-1441	-155	0	0
2012	159	0	-78	0	0	6	-1079	-153	0	0
2013	167	0	-79	1	0	9	-1424	-149	0	0
2014	242	4	-83	-1	183	298	-1178	-232	1	125

Source: Statistics Norway

8.1.3.1 Implications for emission trends

As a result of the different recalculations for 1990-2014 there have been some changes in the trends. The differences are shown in the tables below.

Table 8.9. Trends in emissions 1990-2014. This submission vs. previous submission.

Main Pollutants. Per cent change 1990-2014

	SO ₂	NO _x	NMVOC	CO	NH ₃
2017 submission	-68.5	-20.9	-47.7	-54.3	11.8
2016 submission	-68.2	-26.7	-52.6	-67.7	5.6

Source: Statistics Norway

Table 8.10. Trends in emissions 1990-2014. This submission vs. previous submission.

Particulate Matter. Per cent change 1990-2014

	TSP	PM ₁₀	PM _{2.5}	BC
2017 submission	-22.2	-29.4	-33.7	-28.2
2016 submission	-18.9	-30.7	-33.7	-25.9

Source: Statistics Norway

Table 8.11. Trends in emissions 1990-2014. This submission vs previous submission.

POPs and heavy metals. Per cent change 1990-2014

	Lead	Cadmium	Mercury	Arsenic	Chromium	Copper	PAH-4	Dioxins	HCB	PCB
2017 submission	-97.3	-72.3	-81.5	-64.3	-71.5	18.7	-67.5	-85.7	-98.9	-87.6
2016 submission	-97.4	-72.6	-76.4	-64.7	-73.1	17.6	-78.8	-85.5	-98.9	-87.7

Source: Statistics Norway

8.2 Planned improvements

8.2.1 Overview

There are several areas where improvement actions are needed to improve the Norwegian emission inventory system. In this chapter the main issues are listed.

8.2.2 General

- Many of the emission factors used in the inventory are relatively old, some over 10 years, and need to be analysed. Some of them also lack good documentation and source references. Work has begun to go through all emission factors, register digitally those that have sufficiently documentation and flag those that do not for future revision. This work needs to be continued.

8.2.3 Energy

- The energy statistics used as input to the Norwegian emission inventory is being improved continuously.
- The technical system for the energy balance is being substituted to a more robust and reliable solution. The project includes improvements in methodologies. Results will be

published in may 2017 and incorporated in the next emission inventory. This may lead to reallocations of fuel use with resulting changes in emissions.

- The methodology for estimating emissions of NO_x from energy use in ships will be updated and improved in 2017.

8.2.4 Industrial processes and product use

- The general improvements described in the section 8.2.2 are relevant for this sector.
- For the next round of reporting, we intend to use updated EFs for TSP, PM₁₀ and PM_{2.5} for 2A5B (construction and demolition).
- For the next round of reporting, we intend to include emissions of CO and NH₃ from 2B1 (ammonia production).
- For the next round of reporting, we intend to use updated EF for PCB for 2C6 (zinc production).

8.2.5 Agriculture

- Norway uses a country-specific model to calculate NH₃ emissions from manure management. The model calculates emissions at county level and is complicated to update. There is a need to review the model to make it better fit the requirements for inventory reporting purposes, and to be able to keep the time series updated and consistent. There is also a need to review the emission factors used in the model.

8.2.6 Waste

No improvement has been planned for the waste sector.

9 Projections

9.1 Introduction

This chapter describes in some detail the projections of greenhouse gas emissions and long-range transboundary air pollutants in Norway up to 2030. In line with international reporting guidelines under the Framework Convention on Climate Change these projections are based on an extension of measures and policies implemented by the 4th quarter of 2014.²³ The base year for the projections is 2012. After the projections were made, the methodology for calculating historical emissions has been changed, and hence the emissions have changed. The methodology changes affected mainly emissions of greenhouse gases, but also some changes were done to other emissions. The two major changes performed for projected CLRTAP gases is the NO_x emission factor for diesel motors off shore and off road vehicles. Both have decreased the NO_x emission level. In Table 9.1 emissions for 1990 and 2011-2013 updated with the effect of changed methodology. Only the effects of the new emissions factors for off road vehicles are taken into account in the projections in Table 9.1. See description of recalculation in this IIR.

New projections will be published during this spring.

9.2 The baseline scenario

In the baseline scenario, the development of NO_x emissions is highly dependent on emissions from oil and gas production, which in 2013 constitute of 32 per cent of total Norwegian emissions. Strong growth in the sector has led to an increase in NO_x emissions. In shipping and road traffic, the two other main contributors to NO_x emissions, activity growth has to a larger extent than in oil and gas production, been counteracted by implementation of abatement technology. This is also the case for SO₂ and NMVOC. In the projections, oil and gas production is anticipated to increase until 2015, and is then stabilised on that level up to 2020 and thereafter decrease to 2030. Emissions of NO_x from transport are estimated to continue to be reduced. Table 9.1 shows projected emissions of NO_x, NMVOCs, NH₃ and SO₂. The estimates are based on the same assumptions as for the greenhouse gases. Included in the NO_x projections are the effect of the first and second agreement of NO_x reduction between the Ministry of Climate and Environment and the industry for the period 2008-2017. The aim of the first and second agreement is to reduce NO_x emissions by 34 kilotons.

²³ The projections were presented in the National Budget 2015.

Table 9.1. Anthropogenic emissions of NO, nmVOCs, SO₂ and NH₃. Thousand tonnes

	1990	2011	2012	2013 ¹	2020	2030
NO _x	192.1	172.3	165.2	156.3 (154)	144	123
SO ₂ ²	52.3	18.8	17.3	17.1 (21)	18	18
NM VOC	291.1	134.2	135.8	134.8 (132)	127	114
NH ₃	24.3	26.9	27.2	27.2 (25)	27	27

¹ The Norwegian commitment according to the Gothenburg Protocol in brackets.

Source: Norwegian Environment Agency

9.3 Methodology and key assumptions

The emission projections for Norway presented in this report uses various sources and methodologies. For energy-related emissions, the projections are largely based on macroeconomic model simulations supplemented by available micro studies.

The baseline scenario is based on measures adopted and implemented by the 4th quarter of 2014. The projections are based on information up to September 2014.

9.3.1 Macroeconomic assumptions and CO₂ emissions from the mainland economy

In the projections, current policies are assumed to be continued. Accordingly, CO₂-taxes are maintained at today's level in real terms. Also the carbon price for industries included in the emission trading scheme is assumed to stay at the current level in real terms.

In the projections, total factor productivity is on average assumed to grow annually by almost 1.6 pct. in the mainland economy. As a result, emissions per unit produced in each sector will continue to fall over time. Moreover, for some emission sources it is assumed stronger than average growth in emission technology. Consequently, the amount of emissions in these sectors will not increase proportionally to the production or the use of factor input.

The model-based projections of emissions from road transport have been adjusted, taking on board calculations from Statistics Norway's road model. Road transport with personnel cars is anticipated to stabilize per person but the total transport will grow with the projected growth in population. Use of bio fuels for road transportation is assumed to be about 5-6 per cent of fuel sales by 2020 and 2030. Overall, the projections imply lower growth in the emissions of greenhouse gases from road transport than experienced in the period 1990-2007. Emissions of NO_x from road transport are projected to decrease substantially.

Emissions in 2020 and 2030 will depend on structural developments, particularly in the energy-intensive industries. Production and emissions from energy intensive manufacturing are strongly correlated with electricity consumption in this sector. In the long run, as a technical assumption, demand for electricity from energy intensive manufacturing is anticipated to stay relatively stable at the 2014-level.

Norway is the sixth largest hydro power producer in the world. Emissions from electricity

production are small in Norway, as about 95 per cent (2011) of the supply of electricity originates from hydro power. In the projections, the production of renewables (e.g. hydro and wind power) is exogenously determined.

Table 9.2 lists key macroeconomic assumptions underpinning the Norwegian emission projections. In the baseline scenario, average annual GDP growth is estimated at 1.9 per cent in 2012-2020 and at 2.1 per cent in 2020-2030. Growth in the mainland economy, i.e. total GDP excluding petroleum activities and ocean transport, is estimated at 2.4 per cent in 2012-2030.

Table 9.2 Key macroeconomic assumptions

	2012	2020	2030
	Billion 2012 NOK	Annual average growth rate	
Gross domestic product	2 965	1.9	2.1
- Petroleum activities and ocean transport	670	0.0	-0.9
- Mainland Norway	2 295	2.4	2.4
Manufacturing	201	2.6	1.6
Consumption	1 176	3.2	4.1
Gross fixed capital formation	660	2.4	0.3
- Petroleum activities and ocean transport	182	1.9	-2.3
- Mainland Norway	478	2.7	1.3
Population in 1000	5 051	1.1	0.9
Number of persons employed in 1000	2 684	1.0	0.3
	Level		
Oil price (2015-NOK)	685	545	545
EU-ETS price (2014-NOK)	55	60	90
Electricity price (NOK/KWh 2014-NOK)	0.26	0.27	0.41
Net domestic energy use		Annual average growth rate	
- Petroleum products (PJ)	278.7 ¹	1.4	0.2
- Electricity (TWh)	118.7	1.1	0.3

¹ Including energy-sectors and excluding sea transport in international waters.

Sources: Statistics Norway and Ministry of Finance.

10 Reporting on gridded emissions and LPS

Last update: 18.02.15

10.1 Gridded emissions

Information about the geographical distribution of emissions is useful for modelling and control purposes. The spatial distribution of emissions introduces another dimension (axis) to the general model.

10.1.1 EMEP grid squares

Emissions by EMEP $0.1^\circ \times 0.1^\circ$ grid square are reported to the UNECE and used in models of long-range air pollution. The emissions are allocated to grid squares as follows:

- Emissions from large point sources are allocated directly to the appropriate squares
- Emissions at sea from national sea traffic are allocated to squares on the basis of a AIS-data analysis.
- The remaining emissions are allocated to squares according to the following :
 - When figures for the activity used to calculate emissions are available *directly* at geographical level, these figures are used. Examples are fuel combustion in manufacturing industries and emissions from animals.
 - When the activity at the geographical level is unknown, the national emissions are allocated *indirectly* using surrogate statistical data. For example, fuel combustion in service industries is allocated using employment figures. In a number of cases the activity is known directly at the intermediate level (county), but allocation within counties uses surrogate data.

10.1.2 Scope

Gridded emissions were last reported in 2012 for the years 1990, 1995, 2000, 2005, and 2010, at the EMEP $50\text{km} \times 50\text{km}$ grid. Gridded emissions were reported on G-NFR sources. Gridded emissions of the following components are reported: NO_x , NMVOC, SO_x , NH_3 , $\text{PM}_{2.5}$, PM_{10} , and CO. In addition, gridded emissions of Pb, Cd, Hg, dioxins, PAH-4, and HCB were reported in 2014 for the emission years 2005 and 2010 at the EMEP $50\text{km} \times 50\text{km}$ grid, see chapter 10.1.3.

Gridded emissions on the EMEP $0.1^\circ \times 0.1^\circ$ will be reported in 2017 for the years 1990, 1995, 2000, 2005, 2010 and 2015 for the following pollutants: NO_x , NMVOC, SO_x , NH_3 , TSP, $\text{PM}_{2.5}$, PM_{10} , CO, As, Cd, Cr, Cu, Hg, Pb, Dioxins, HCB, PCB, BC, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and Indeno(1,2,3-cd)pyrene.

10.1.3 Recent improvements

In 2012, emissions were reported using a $50\text{ km} \times 50\text{ km}$ grid. In 2017, emissions will be reported using a $0.1^\circ \times 0.1^\circ$ grid.

10.1.4 Planned improvements

In the next reporting, the Norwegian Environment Agency plans to reduce uncertainty in the methodologies used to allocate emissions on the grid.

10.2 LPS

Large point sources were last reported in 2012 for the emission years 2005 and 2010. Emission were reported for a total of 89 LPS for 2005 and 76 LPS in 2010. In 2017, will LSP be reported according to the new guidelines.

LPS data will be reported for NO_x, NMVOC, SO_x, NH₃, TSP, PM_{2.5}, PM₁₀, CO, As, Cd, Cr, Cu, Hg, Pb, Dioxins, HCB, PCB, BC, benzo(a)pyrene, benzo(b)fluoranthene, benzo(k)fluoranthene and Indeno(1,2,3-cd)pyrene, where emissions exceed the reporting limits included in the 2014 reporting guidelines.

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Tier 1 Key Category Analysis- Norway – 2017 submission

Methodology

The submission includes tier 1 key category analysis for the years 1990 and 2015 for the components SO₂, NO_x, NH₃, NMVOC, CO, TSP, PM₁₀, PM_{2.5}, Pb, Hg, Cd, dioxins, PAH, HCB and PCB.

The same procedure has been performed for 1990 and 2015. The emissions are analysed using the NFR14 sources (from the NFR 2014-2 reporting template) for both years. For each component the sources have been sorted according to their share of emissions, and the percentage of emissions of the component has been calculated. Sources are assigned as key until 95% of total emissions are covered.

For convenience, the analysis was performed with a few exceptions from the NFR14:

- Gasoline evaporation (1 A 3 b v) is included in 1 A 3 b i-iv

These exceptions do not change the ranking of the other categories, but they may affect which categories are included at the margin.

The result tables 1-18 are sorted by share of total emissions in 2015 for each component separately. Key categories in 1990 which were not key in 2015 are placed at the bottom of each table.

When a source has become key in 2015, this may either mean that the emissions from this source have increased, or that it has decreased less than other sources. The key category analysis does not give information about the magnitude of emissions from each source, and can thus not be used to evaluate trends in emission levels for any given source.

Results

74 per cent of the 93 sources with emissions in 2015 were key to at least one component. This means that 24 sources have emissions, but are not key category for any component. This is especially prominent within the solvents and agriculture categories.

Some sources are key category to a wide range of components. This is the case in particular for public electricity and heat production (1 A 1 a), manufacture of solid fuels and other energy industries (1 A 1 c), road traffic (1 A 3 B i-iii), national navigation (1 A 3 d ii), residential plants (1 A 4 b i), ferroalloys production (2 C 2), aluminium production (2 C 3) and field burning of agricultural residue (3 F). The latter is not key to any primary gases, except CO, but key to particulates and a range of heavy metals and POPs.

Iron and steel production (2 C 1) is key category only to POPs and heavy metals.

Looking at the three most dominant sources of emissions for each component in 2015, it

becomes clear that there are some sources that are responsible for a large proportion of emissions. This is the case for emissions from manufacture of solid fuels and other energy industries, passenger cars, residential plants, ferroalloys production and aluminium production. Some distinctive characteristics of the Norwegian society can explain why some sources are dominant key categories for emissions from many components. For instance, long and cold winters lead to high demand for heating of houses, and wood-burning is common. The wood-burning leads to high emissions of CO, particulate matter, cadmium and POPs from residential plants. Due to a history of cheap electricity (hydroelectric power), Norway has a high share of energy-demanding industry. Thus, industries such as ferroalloys and aluminium production dominate the emissions for SO₂, heavy metals and PAH.

Key categories for SO₂

Production of ferroalloys was the dominant source for emissions of SO₂ in both 1990 and 2015, with, respectively, 23 and 35 per cent of the total (Table A1). The importance of public electricity and heat production has grown considerably, from 2 per cent in 1990 to 11 per cent in 2015. Several sources which were key in 1990 are no longer so in 2015. Particularly, this is the case for road traffic, due to lower sulphur content in petrol and auto diesel. On the other hand, emissions from venting and flaring were not key in 1990, but have become so in 2015.

Table A1. Key categories for SO₂ emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
2C2	Ferroalloys production	23.0 %	34.5 %
1A1a	Public electricity and heat production	2.1 %	11.2 %
2C3	Aluminium production	8.2 %	7.8 %
2C7c	Other metal production	0.9 %	5.9 %
1B2aiv	Fugitive emissions oil: Refining / storage	6.9 %	5.6 %
2B5	Carbide production	8.5 %	4.4 %
1A3dii	National navigation (shipping)	8.0 %	3.9 %
2A1	Cement production	1.1 %	3.4 %
1A1b	Petroleum refining	0.7 %	3.4 %
1A1c	Manufacture of solid fuels and other energy industries	0.8 %	3.4 %
1A4ciii	Agriculture/Forestry/Fishing: National fishing	3.6 %	2.6 %
1A4bi	Residential: Stationary	2.5 %	1.9 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	1.6 %	1.9 %
1A4ai	Commercial/institutional: Stationary	1.9 %	1.7 %
1A2f	Stationary combustion in manufacturing industries and construction: Non-metallic minerals	1.7 %	1.6 %
1B2c	Venting and flaring (oil, gas, combined oil and gas)	0.1 %	1.3 %
2A6	Other mineral products	0.5 %	1.0 %
2H1	Pulp and Paper	3.6 %	0.6 %
1A3biii	Road transport: Heavy duty vehicles and buses	3.4 %	0.1 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	3.1 %	0.7 %
1A2e	Stationary combustion in manufacturing industries and construction: Food processing, beverages and tobacco	2.5 %	0.5 %
2B10a	Chemical industry: Other	2.4 %	0.0 %
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	2.3 %	0.9 %
1A3bi	Road transport: Passenger cars	2.2 %	0.1 %
2B6	Titanium dioxide production	1.0 %	0.0 %
1A2gvii	Mobile Combustion in manufacturing industries and construction:	0.9 %	0.1 %
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery	0.9 %	0.0 %
1A3bii	Road transport: Light duty vehicles	0.7 %	0.1 %
2H3	Other industrial processes	1.1 %	.

Key categories for NO_x

Manufacture of solid fuels and other energy industries was the most important emission source of NO_x in 2015, with a little less than a third of the emissions (Table A2). In 1990, Road transport: Passenger cars was the dominant source, and the three road transport groups together were responsible for one third of the emissions. In 2015, this share was reduced to one fifth. The actual emissions were, however, more than halved in the period, partly due to an increased share of cars with catalysts. Some new sources have become key categories since 1990. This is the case for commercial/institutional mobile combustion and civil aviation (domestic and international).

Table A2. Key categories for NO_x emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A1c	Manufacture of solid fuels and other energy industries	12,4 %	31,0 %
1A3dii	National navigation (shipping)	15,2 %	11,8 %
1A3bi	Road transport: Passenger cars	17,8 %	8,3 %
1A3biii	Road transport: Heavy duty vehicles and buses	12,2 %	8,2 %
1A4cii	Agriculture/Forestry/Fishing: National fishing	11,3 %	8,0 %
2C2	Ferroalloys production	4,7 %	5,5 %
1A3bii	Road transport: Light duty vehicles	2,5 %	3,3 %
3DA1	Inorganic N-fertilizers (includes also urea application)	2,2 %	2,7 %
1A2gvii	Mobile Combustion in manufacturing industries and construction:	3,5 %	2,4 %
3DA2a	Animal manure applied to soils	1,2 %	1,8 %
1A4aii	Commercial/institutional: Mobile	0,5 %	1,7 %
1A1a	Public electricity and heat production	0,7 %	1,3 %
1A5b	Other, Mobile (including military, land based and recreational boats)	1,5 %	1,2 %
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery	1,9 %	1,2 %
1A3ai(i)	1 A 3 a i (i) International Aviation (LTO)	0,2 %	1,2 %
1A3aii(i)	Domestic aviation LTO (civil)	0,4 %	1,2 %
1A2f	Stationary combustion in manufacturing industries and construction: Non-metallic minerals	1,7 %	1,0 %
2B2	Nitric acid production	1,2 %	0,8 %
1A4bii	Residential: Household and gardening (mobile)	0,6 %	0,8 %
1A4bi	Residential: Stationary	1,0 %	0,8 %
1B2aiv	Fugitive emissions oil: Refining / storage	0,6 %	0,8 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,8 %	0,6 %
1A1b	Petroleum refining	0,8 %	0,4 %
1A3c	Railways	0,7 %	0,5 %
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	0,7 %	0,2 %

Key categories for NH₃

There has been little change in the key categories for NH₃ from 1990 to 2015 (Table A3).

Agricultural sources are dominant. However, field burning of agricultural residue, which was a key category in 1990, was no longer so in 2015. For passenger cars and sewage sludge applied to soils and other waste, there has been an opposite development – these were key categories only in 2015.

Table A3. Key categories for NH₃ emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
3DA2a	Animal manure applied to soils	52,5 %	52,4 %
3DA1	Inorganic N-fertilizers (includes also urea application)	5,4 %	7,9 %
3B1a	Manure management - Dairy cattle	6,2 %	5,2 %
3B1b	Manure management - Non-dairy cattle	3,5 %	4,3 %
3B3	Manure management - Swine	4,4 %	4,2 %
3DA3	Urine and dung deposited by grazing animals	4,3 %	3,6 %
1A3bi	Road transport: Passenger cars	0,7 %	3,0 %
3B4gi	Manure management - Laying hens	1,9 %	2,8 %
3DA2b	Sewage sludge applied to soils	0,6 %	2,3 %
2B2	Nitric acid production	2,0 %	2,2 %
3B2	Manure management - Sheep	2,7 %	2,1 %
3B4e	Manure management - Horses	0,8 %	1,7 %
3B4gii	Manure management - Broilers	1,2 %	1,7 %
5E	Other waste	0,5 %	1,3 %
3I	Agriculture other	7,4 %	1,2 %
3F	Field burning of agricultural residue	4,1 %	0,4 %

Key categories for NMVOC

NMVOC emissions are spread on a wide range of sources. Offshore loading of oil is the dominant emission source (Table A4), but due to increased use of emission reducing technology, this source has become less dominant during the period from 1990 to 2015. These emissions were responsible for 39 per cent of the total in 1990, but only 21 per cent in 2015. In 2015, solvents were the second most important source for NMVOC. Due to decreases in emissions from other sources, particularly oil loading, the share is higher than in 1990, although the actual emissions have been reduced. NMVOC emissions from passenger cars was the second largest emission source in 1990, but this source's share of total emissions was reduced from 18 per cent in 1990 to 3 per cent in 2015.

Table A4. Key categories for NMVOC emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1B2ai	Fugitive emissions oil: Exploration, production, transport	38,7 %	21,2 %
2D3i	Other solvent use	11,1 %	16,6 %
2D3a	Domestic solvent use including fungicides	.	8,7 %
1A4bii	Residential: Household and gardening (mobile)	3,6 %	7,3 %
1B2c	Venting and flaring (oil, gas, combined oil and gas)	1,1 %	6,8 %
1B2aiv	Fugitive emissions oil: Refining / storage	3,1 %	4,9 %
1A4bi	Residential: Stationary	2,4 %	4,5 %
2D3d	Coating applications	4,3 %	4,4 %
1A3bi	Road transport: Passenger cars	17,8 %	3,0 %
1B2av	Distribution of oil products	3,4 %	2,8 %
3B1b	Manure management - Non-dairy cattle	1,4 %	2,3 %
3B1a	Manure management - Dairy cattle	1,3 %	1,8 %
1A1c	Manufacture of solid fuels and other energy industries	0,4 %	1,7 %
1B2b	Fugitive emissions from natural gas (exploration, production, processing, transmission, storage, distribution and other)	0,4 %	1,5 %
1A1a	Public electricity and heat production	0,1 %	1,3 %
1A3biv	Road transport: Mopeds & motorcycles	0,7 %	1,1 %
2C2	Ferroalloys production	0,5 %	0,9 %
2H2	Food and beverages industry	0,5 %	0,8 %
3B4gii	Manure management - Broilers	0,2 %	0,8 %
1A3dii	National navigation (shipping)	0,6 %	0,7 %
5A	Solid waste disposal on land	0,4 %	0,5 %
3B4gi	Manure management - Laying hens	0,2 %	0,5 %
3DA1	Inorganic N-fertilizers (includes also urea application)	0,3 %	0,5 %
1B1a	Fugitive emission from solid fuels: Coal mining and handling	0,4 %	0,4 %
2D3e	Degreasing	0,4 %	0,3 %
1A3bii	Road transport: Light duty vehicles	1,6 %	0,2 %
2D3g	Chemical products	0,9 %	0,2 %
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery	0,7 %	0,3 %
1A3biii	Road transport: Heavy duty vehicles and buses	0,7 %	0,3 %
2D3f	Dry cleaning	0,5 %	0,1 %

Key categories for CO

Aluminium production was the most important emission source for CO in 2015. This share grew from 13 per cent in 1990 to 38 per cent in 2015. In 1990, passenger cars was the dominant source, with 42 per cent of the total CO emissions; this share was reduced to only 8 per cent in 2015 (Table A5). Emissions from combustion in households, primarily of fire wood, increased its emission share from 18 per cent in 1990 to 24 per cent in 2015, although the actual emissions were reduced by 40 per cent. Emission reductions from some sources have caused that several minor sources have become key categories in 2015, even though their actual emissions may have been reduced.

Table A5. Key categories for CO emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
2C3	Aluminium production	12,7 %	38,2 %
1A4bi	Residential: Stationary	18,5 %	24,3 %
1A4bii	Residential: Household and gardening (mobile)	4,8 %	10,7 %
1A3bi	Road transport: Passenger cars	42,1 %	7,9 %
1A1c	Manufacture of solid fuels and other energy industries	0,6 %	2,7 %
2B5	Carbide production	4,9 %	2,3 %
1A1a	Public electricity and heat production	0,1 %	2,1 %
1A3biv	Road transport: Mopeds & motorcycles	0,8 %	2,0 %
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery	1,0 %	1,2 %
1A3biii	Road transport: Heavy duty vehicles and buses	0,7 %	1,0 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	0,5 %	0,8 %
3F	Field burning of agricultural residue	3,3 %	0,8 %
1A3bii	Road transport: Light duty vehicles	5,4 %	0,8 %
1A4ciii	Agriculture/Forestry/Fishing: National fishing	0,4 %	0,7 %
2C4	Magnesium production	2,4 %	.

Key categories for particulates

The dominant emission source for particulates of all sizes both in 1990 and 2015 is burning of fuel wood in small stoves in households (Tables A6-A8). For PM_{2.5}, more than half of the emissions came from this source in both 1990 and 2015. The importance of the different other emission sources vary some what between the different PM fractions, but other mineral products is important for TSP and PM₁₀ and ferroalloys production for both PM₁₀ and PM_{2.5}.

Table A6. Key categories for TSP emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A4bi	Residential: Stationary	32,7 %	31,7 %
2A6	Other mineral products	6,3 %	10,0 %
2D3b	Road paving with asphalt	5,1 %	9,1 %
1A3bvi	Road transport: Automobile tyre and brake wear	4,3 %	8,3 %
1A3bvii	Road transport: Automobile road abrasion	11,0 %	8,2 %
2C2	Ferroalloys production	5,7 %	3,6 %
1A1a	Public electricity and heat production	0,7 %	2,6 %
2A5b	Construction and demolition	1,6 %	2,6 %
2C3	Aluminium production	3,6 %	2,3 %
1A1c	Manufacture of solid fuels and other energy industries	0,7 %	2,1 %
1A3dii	National navigation (shipping)	1,9 %	1,6 %
3B4gii	Manure management - Broilers	0,5 %	1,4 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,3 %	1,2 %
1A4ciii	Agriculture/Forestry/Fishing: National fishing	1,1 %	1,1 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	1,6 %	1,0 %
1B2c	Venting and flaring (oil, gas, combined oil and gas)	1,6 %	1,0 %
3B4gi	Manure management - Laying hens	0,5 %	1,0 %
3DC	Farm-level agricultural operations including storage, handling and transport of agricultural products	1,0 %	1,0 %
3B3	Manure management - Swine	0,6 %	0,8 %
2B2	Nitric acid production	1,9 %	0,8 %
1A4bii	Residential: Household and gardening (mobile)	0,5 %	0,7 %
3B1b	Manure management - Non-dairy cattle	0,6 %	0,7 %
1A3bi	Road transport: Passenger cars	0,4 %	0,6 %
3B4giv	Manure management - Other poultry	0,2 %	0,6 %
3B1a	Manure management - Dairy cattle	0,7 %	0,6 %
3F	Field burning of agricultural residue	3,4 %	0,5 %
1A3bii	Road transport: Light duty vehicles	0,7 %	0,5 %
2B5	Carbide production	2,0 %	0,1 %
1A3biii	Road transport: Heavy duty vehicles and buses	1,5 %	0,4 %
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery	1,1 %	0,4 %
2H3	Other industrial processes	0,8 %	0,0 %
2H1	Pulp and Paper	0,6 %	0,1 %
1A2gvii	Mobile Combustion in manufacturing industries and construction:	0,4 %	0,2 %

Table A7. Key categories for PM₁₀ emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A4bi	Residential: Stationary	41,9 %	44,4 %
2A6	Other mineral products	4,2 %	7,1 %
2C2	Ferroalloys production	7,5 %	5,1 %
1A3bvii	Road transport: Automobile road abrasion	4,3 %	3,5 %
1A1a	Public electricity and heat production	0,7 %	3,4 %
2C3	Aluminium production	4,6 %	3,1 %
1A1c	Manufacture of solid fuels and other energy industries	0,9 %	3,0 %
2D3b	Road paving with asphalt	1,4 %	2,8 %
1A3dii	National navigation (shipping)	2,5 %	2,2 %
3B4gii	Manure management - Broilers	0,7 %	2,0 %
2A5b	Construction and demolition	1,1 %	1,8 %
1A4ciii	Agriculture/Forestry/Fishing: National fishing	1,5 %	1,5 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,3 %	1,5 %
1A3bvi	Road transport: Automobile tyre and brake wear	0,7 %	1,5 %
3B4gi	Manure management - Laying hens	0,7 %	1,4 %
3DC	Farm-level agricultural operations including storage, handling and transport of agricultural products	1,3 %	1,4 %
1B2c	Venting and flaring (oil, gas, combined oil and gas)	2,0 %	1,4 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	1,9 %	1,3 %
1A4bii	Residential: Household and gardening (mobile)	0,7 %	1,0 %
2B2	Nitric acid production	2,0 %	0,9 %
1A3bi	Road transport: Passenger cars	0,5 %	0,9 %
3B4giv	Manure management - Other poultry	0,2 %	0,8 %
3F	Field burning of agricultural residue	4,4 %	0,7 %
1A3bii	Road transport: Light duty vehicles	0,9 %	0,7 %
1A4ai	Commercial/institutional: Stationary	0,1 %	0,6 %
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery	1,4 %	0,6 %
1A3biii	Road transport: Heavy duty vehicles and buses	2,0 %	0,5 %
2B5	Carbide production	2,7 %	0,1 %
2H1	Pulp and Paper	0,7 %	0,1 %
2H3	Other industrial processes	0,5 %	0,0 %
1A2gvii	Mobile Combustion in manufacturing industries and construction:	0,5 %	0,2 %
2G	Other product use	0,4 %	0,3 %
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	0,4 %	0,2 %
5E	Other waste	0,4 %	0,4 %

Table A8. Key categories for PM_{2.5} emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A4bi	Residential: Stationary	51,0 %	56,9 %
2C2	Ferroalloys production	9,4 %	6,8 %
1A1a	Public electricity and heat production	0,4 %	4,1 %
1A1c	Manufacture of solid fuels and other energy industries	1,1 %	3,8 %
1A3dii	National navigation (shipping)	3,0 %	2,8 %
1A4ciii	Agriculture/Forestry/Fishing: National fishing	1,7 %	1,9 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,3 %	1,9 %
2C3	Aluminium production	2,5 %	1,8 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	2,3 %	1,8 %
1B2c	Venting and flaring (oil, gas, combined oil and gas)	2,0 %	1,7 %
1A3bvi	Road transport: Automobile tyre and brake wear	0,6 %	1,4 %
1A4bii	Residential: Household and gardening (mobile)	0,9 %	1,3 %
1A3bi	Road transport: Passenger cars	0,6 %	1,1 %
2A6	Other mineral products	0,7 %	1,0 %
3F	Field burning of agricultural residue	5,2 %	0,9 %
2B2	Nitric acid production	1,8 %	0,9 %
1A3bii	Road transport: Light duty vehicles	1,1 %	0,9 %
1A3bvii	Road transport: Automobile road abrasion	0,9 %	0,8 %
1A4ai	Commercial/institutional: Stationary	0,1 %	0,7 %
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery	1,6 %	0,7 %
1A3biii	Road transport: Heavy duty vehicles and buses	2,4 %	0,7 %
2D3b	Road paving with asphalt	0,2 %	0,5 %
5E	Other waste	0,5 %	0,5 %
2G	Other product use	0,6 %	0,4 %
2B5	Carbide production	3,3 %	0,2 %
1A2gvii	Mobile Combustion in manufacturing industries and construction:	0,6 %	0,3 %
2H1	Pulp and Paper	0,5 %	0,0 %
1A2D	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	0,5 %	0,2 %

Key categories for lead (Pb)

There has been a dramatic change in dominant sources for emissions of lead from 1990 to 2015 (Table A9). In 1990, road traffic, particularly passenger cars, was by far the most important source, with 90 per cent of total lead emissions, due to high lead content in petrol. In 2015, petrol no longer contained significant amounts of lead, and other sources had become dominant. The most significant emission source in 2015 was automobile tyre and brake wear, with 30 per cent of the emissions. Iron and steel production was the second most important source in 2015, with 18 per cent of the total emissions. Due to the reduced importance of road traffic, far more sources were key in 2015 than in 1990.

Table A9. Key categories for Pb emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A3bvi	Road transport: Automobile tyre and brake wear	0,6 %	30,4 %
2C1	Iron and steel production	1,3 %	18,1 %
1A3aii(i)	Domestic aviation LTO (civil)	0,3 %	7,4 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,0 %	7,2 %
1A2f	Stationary combustion in manufacturing industries and construction: Non-metallic minerals	0,1 %	4,3 %
2C7c	Other metal production	0,1 %	4,3 %
2C3	Aluminium production	0,3 %	3,6 %
2C2	Ferroalloys production	0,7 %	2,9 %
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	0,1 %	2,6 %
1A4ai	Commercial/institutional: Stationary	0,0 %	2,4 %
1A3bi	Road transport: Passenger cars	82,3 %	2,3 %
5E	Other waste	0,2 %	2,1 %
2B6	Titanium dioxide production	0,1 %	1,8 %
1A3biii	Road transport: Heavy duty vehicles and buses	0,6 %	1,8 %
1A4bi	Residential: Stationary	0,1 %	1,1 %
2B5	Carbide production	0,3 %	0,9 %
1A3bii	Road transport: Light duty vehicles	6,1 %	0,9 %
1A1a	Public electricity and heat production	1,1 %	0,9 %
1A3dii	National navigation (shipping)	0,1 %	0,7 %
1A4bii	Residential: Household and gardening (mobile)	3,0 %	0,1 %
1A3biv	Road transport: Mopeds & motorcycles	0,9 %	0,0 %
2A3	Glass production	0,8 %	0,0 %

Key categories for mercury (Hg)

Mercury emissions stem from a wide variety of sources (Table A10). In 1990, ferroalloys production and other product use (emissions from thermometers, fluorescent tubes and other instruments) dominated, with more than half of the total emissions. In 2015, ferroalloys production was still the largest contributor, but the share was reduced from 35 per cent in 1990 to 10 per cent in 2015.

Table A10. Key categories for Hg emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
2C2	Ferroalloys production	35,2 %	10,0 %
1A1c	Manufacture of solid fuels and other energy industries	0,7 %	9,6 %
1A3dii	National navigation (shipping)	2,8 %	8,0 %
1A4cii	Agriculture/Forestry/Fishing: National fishing	1,7 %	7,0 %
1A4bi	Residential: Stationary	2,2 %	6,3 %
1A1a	Public electricity and heat production	6,9 %	6,2 %
1A4ai	Commercial/institutional: Stationary	1,0 %	5,7 %
2G	Other product use	20,0 %	5,2 %
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	1,7 %	4,6 %
2C7c	Other metal production	.	4,0 %
1A3bi	Road transport: Passenger cars	0,9 %	3,7 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,8 %	3,5 %
2C1	Iron and steel production	7,0 %	2,8 %
3F	Field burning of agricultural residue	4,0 %	2,7 %
1A2f	Stationary combustion in manufacturing industries and construction: Non-metallic minerals	0,7 %	2,6 %
2A6	Other mineral products	0,4 %	1,9 %
1A3ai(i)	1 A 3 a i (i) International Aviation (LTO)	0,1 %	1,8 %
1A3aii(i)	Domestic aviation LTO (civil)	0,2 %	1,7 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	1,6 %	1,7 %
2A1	Cement production	1,8 %	1,6 %
1A3bvi	Road transport: Automobile tyre and brake wear	0,2 %	1,5 %
2B6	Titanium dioxide production	0,2 %	1,3 %
1A5b	Other, Mobile (including military, land based and recreational boats)	0,3 %	1,0 %
1A3biii	Road transport: Heavy duty vehicles and buses	0,1 %	1,0 %
2B10a	Chemical industry: Other	5,0 %	0,2 %
5C1bv	Cremation	1,5 %	0,4 %
1A2e	Stationary combustion in manufacturing industries and construction: Food processing, beverages and tobacco	1,1 %	0,9 %

Key categories for cadmium (Cd)

Whereas field burning of agricultural residues was the most important emission source for cadmium in 1990, combustion in households, particularly of fire wood, dominated in 2015 (Table A11). Metal production was responsible for more than one third of the emissions in 1990, but its share was reduced significantly in 2015. More minor sources were key in 2015 than in 1990.

Table A11. Key categories for Cd emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A4bi	Residential: Stationary	7,0 %	21,5 %
1A1a	Public electricity and heat production	6,3 %	9,3 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,4 %	9,0 %
3F	Field burning of agricultural residue	23,6 %	8,8 %
1A3bvii	Road transport: Automobile road abrasion	3,4 %	6,3 %
2C3	Aluminium production	6,1 %	4,8 %
2C2	Ferroalloys production	16,9 %	4,6 %
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	8,1 %	4,6 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	2,5 %	4,5 %
2C6	Zinc production	10,2 %	4,0 %
1A3bi	Road transport: Passenger cars	1,1 %	3,9 %
1A4ai	Commercial/institutional: Stationary	0,3 %	2,6 %
1A1c	Manufacture of solid fuels and other energy industries	0,4 %	2,6 %
1A2f	Stationary combustion in manufacturing industries and construction: Non-metallic minerals	1,4 %	2,3 %
1A3biii	Road transport: Heavy duty vehicles and buses	0,4 %	2,1 %
2C1	Iron and steel production	3,1 %	1,4 %
1A3bii	Road transport: Light duty vehicles	0,1 %	1,1 %
1A3dii	National navigation (shipping)	0,9 %	0,9 %
1A4ciii	Agriculture/Forestry/Fishing: National fishing	0,4 %	0,7 %
5E	Other waste	0,1 %	0,6 %
2B5	Carbide production	5,9 %	0,1 %

Key categories for dioxins

In 1990, other industrial processes, i.e. ore mines, and magnesium production were the largest sources of dioxin emissions (Table A12). The enterprises responsible for these emissions have been shut down since 1990, and thus other sources have become dominant. In 2015, residential plants were responsible for one third of the dioxin emissions in Norway. Most of these emissions came from use of fire wood. Since the major emission sources in 1990 have disappeared, several minor sources have become key in 2015.

Table A12. Key categories for dioxin emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A4bi	Residential: Stationary	5,1 %	34,1 %
1A1c	Manufacture of solid fuels and other energy industries	0,5 %	9,8 %
1A3dii	National navigation (shipping)	1,8 %	8,9 %
5E	Other waste	1,8 %	8,4 %
1A4cii	Agriculture/Forestry/Fishing: National fishing	1,4 %	7,7 %
2C1	Iron and steel production	0,9 %	6,9 %
2C3	Aluminium production	0,6 %	4,9 %
1A1a	Public electricity and heat production	10,6 %	4,8 %
2C2	Ferroalloys production	1,3 %	3,6 %
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals	0,1 %	2,0 %
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print	1,0 %	1,4 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	0,3 %	1,2 %
1A3bi	Road transport: Passenger cars	1,7 %	1,1 %
1A5b	Other, Mobile (including military, land based and recreational boats)	0,1 %	0,7 %
2H3	Other industrial processes	42,6 %	.
2C4	Magnesium production	25,1 %	.
5C1biii	Clinical waste incineration	4,2 %	.

Key categories for PAH

Aluminium production was the most important emission source of PAH emissions in both 1990 and 2015, although it was less dominant in 2015 (Tables A13-16). A strong reduction in emissions from aluminium production from 1990 to 2015 has brought about that road transport has become key in 2015, although the emission increase in the different road transport groups not has been substantial. Combustion of fire wood in households is also an important emission source for PAH.

Table A13. Key categories for benzo(a)pyrene emissions, 1990 and 2015. Key categories are given in bold italic

	Source	1990	2015
2C3	Aluminium production	68,5 %	45,4 %
1A4bi	Residential: Stationary	18,6 %	24,6 %
1A3bi	Road transport: Passenger cars	2,2 %	10,8 %
1A3bii	Road transport: Light duty vehicles	0,4 %	4,3 %
2C2	Ferroalloys production	0,7 %	3,7 %
2B5	Carbide production	1,5 %	3,2 %
1A3biii	Road transport: Heavy duty vehicles and buses	0,5 %	2,0 %
3F	Field burning of agricultural residue	3,8 %	1,4 %
2C7c	Other metal production	2,2 %	0,4 %

Table A14. Key categories for benzo(b)fluoranthene emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
2C3	Aluminium production	70,7 %	45,5 %
1A4bi	Residential: Stationary	13,4 %	17,3 %
1A3bvi	Road transport: Automobile tyre and brake wear	1,8 %	8,6 %
1A3biii	Road transport: Heavy duty vehicles and buses	1,5 %	5,7 %
1A3bi	Road transport: Passenger cars	1,2 %	5,6 %
2C2	Ferroalloys production	0,7 %	3,9 %
2B5	Carbide production	1,6 %	3,3 %
1A3bii	Road transport: Light duty vehicles	0,2 %	2,2 %
3F	Field burning of agricultural residue	4,9 %	1,8 %
1A2f	Stationary combustion in manufacturing industries and construction: Non-metallic minerals	0,2 %	1,3 %
2C7c	Other metal production	2,3 %	0,4 %

Table A15. Key categories for benzo(k)fluoranthene emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
2C3	Aluminium production	77,7 %	52,9 %
1A3biii	Road transport: Heavy duty vehicles and buses	3,3 %	13,1 %
1A3bi	Road transport: Passenger cars	1,6 %	8,8 %
1A4bi	Residential: Stationary	6,7 %	8,6 %
2C2	Ferroalloys production	0,8 %	4,4 %
2B5	Carbide production	1,7 %	3,8 %
1A3bii	Road transport: Light duty vehicles	0,3 %	3,6 %
3F	Field burning of agricultural residue	4,1 %	1,5 %
2C7c	Other metal production	2,5 %	0,5 %

Table A16. Key categories for indeno(1,2,3_cd)pyrene emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
2C3	Aluminium production	57,8 %	34,5 %
1A4bi	Residential: Stationary	22,7 %	25,2 %
1A3bi	Road transport: Passenger cars	4,6 %	15,2 %
1A3bii	Road transport: Light duty vehicles	0,8 %	5,8 %
1A3biii	Road transport: Heavy duty vehicles and buses	1,4 %	4,6 %
2C2	Ferroalloys production	0,6 %	2,7 %
2B5	Carbide production	1,3 %	2,3 %
5E	Other waste	1,4 %	2,2 %
1A3dii	National navigation (shipping)	0,9 %	1,8 %
1A4cii	Agriculture/Forestry/Fishing: National fishing	0,9 %	1,8 %
3F	Field burning of agricultural residue	5,5 %	1,7 %
2C7c	Other metal production	1,9 %	0,3 %

Key categories for HCB

In 1990, magnesium production was by far the largest source for HCB emissions, with almost 99 per cent of the total (Table A13). This production had ceased to exist in 2015, and road traffic had become the dominant source. The three groups of road transport were together responsible for more than half of the emissions. Aluminium production was the second most

important source in 2015. However, HCB emissions in Norway are now negligible.

Table A17. Key categories for HCB emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A3bi	Road transport: Passenger cars	0,0 %	35,0 %
2C3	Aluminium production	0,1 %	13,1 %
1A3bii	Road transport: Light duty vehicles	0,0 %	12,0 %
2B10a	Chemical industry: Other	0,0 %	9,3 %
1A1a	Public electricity and heat production	0,6 %	7,7 %
1A4bi	Residential: Stationary	0,1 %	5,9 %
1A3biii	Road transport: Heavy duty vehicles and buses	0,0 %	5,1 %
1A3dii	National navigation (shipping)	0,0 %	2,4 %
1A4ciii	Agriculture/Forestry/Fishing: National fishing	0,0 %	1,9 %
1A1c	Manufacture of solid fuels and other energy industries	0,0 %	1,3 %
1A2gviii	Stationary Combustion in manufacturing industries and construction: Other	0,0 %	1,2 %
2C1	Iron and steel production	0,0 %	1,2 %
2C4	Magnesium production	98,8 %	.

Key categories for PCB

Road traffic is by far the most important source for emissions of PCB. There has, however, been a strong shift between the different road traffic groups. In 1990, most of the emissions came from passenger cars, whereas heavy duty vehicles and buses were dominant in 2015 (Table A18).

Table A17. Key categories for PCB emissions, 1990 and 2015. Key categories are given in bold italic.

	Source	1990	2015
1A3biii	Road transport: Heavy duty vehicles and buses	6,1 %	82,3 %
5E	Other waste	1,5 %	4,4 %
1B2c	Venting and flaring (oil, gas, combined oil and gas)	2,6 %	4,3 %
1A3bi	Road transport: Passenger cars	79,1 %	4,3 %
1A3bii	Road transport: Light duty vehicles	5,9 %	2,0 %
1A4bii	Residential: Household and gardening (mobile)	2,9 %	0,0 %

Emission factors used in the estimations of emissions from combustion

In the calculations, the numbers are used with the highest available accuracy. In this tables though, they are only shown rounded off, which in some cases can lead to the result that the exceptions looks the same as the general factors. The tables include the emission factors used for estimating the acidifying pollutants, heavy metals and persistent organic pollutants. Due to practical reasons, also the emission factors used for the Norwegian greenhouse gas inventory are included in the tables.

For road traffic, this general view of the emission factors only includes last year's factors and not all time series.

In the tables for stationary combustion, dotted cells indicate combinations of fuel and source without consumption.

A description of the sector codes used in the tables is given in Appendix D.

SO₂ and heavy metals - Stationary and mobile combustion

Table B1. General emission factors for SO₂ and heavy metals

	SO ₂ ¹ kg/tonne ²	Pb g/tonne ²	Cd g/tonne ²	Hg g/tonne ²	As g/tonne ²	Cr g/tonne ²	Cu g/tonne ²
Coal	16 ³	0.2 ³	0.003 ³	0.05 ³	0.089 ³	0.065 ³	0.087 ³
Coke	18	0.2 ³	0.003 ³	0.05 ³	0.089 ³	0.065 ³	0.087 ³
Petrol coke	18	0.2	0.003	0.05	0.089	0.065	0.087
Charcoal	0.32	0.8	0.38	0.02	0.01	0.68	0.18
Motor gasoline	0.001	0.03⁴	0.01	0.0084	0.05	0.05	1.7
Aviation gasoline	0.4	675.7	0.01	0	0.05	0.05	1.7
Kerosene (heating)	0.346	0.07	0.01	0.03	0.05	0.04	0.05
Jet kerosene	0.274	0.07	0.01	0.03	0.05	0.05	0.05
Auto diesel	0.015⁵	0.1	0.01	0.0023	0.05	0.05	1.7
Marine gas oil/diesel	1.158	0.1	0.01	0.05	0.05	0.04	0.05
Light fuel oils	0.928	0.1	0.01	0.05	0.05	0.04	0.05
Heavy distillate	4.375	0.1	0.01	0.05	0.05	0.04	0.05
Heavy fuel oil	17.84⁶	1	0.1	0.2	0.057	1.35	0.53
Natural gas (1000 Sm ³) ...	0	0.00025	0.002	0.001	0.004	0.021	0.016
LPG	0	0	0	0	0.004	0.021	0.016
Refinery gas	0	0	0	0	0.004	0.021	0.016
CO gas	0	0	0	0	0.004	0.021	0.016
Fuel gas	0	0	0	0	0.004	0.021	0.016
Landfill gas	0.019	0	0	0	0.004	0.021	0.016
Biogas	0	0.00025	0.0017	0.001	0.0038	0.021	0.016
Fuel wood	0.2	0.05	0.1	0.010244	0.159	0.152	0.354
Wood waste	0.37	0.05	0.1	0.010244	0.159	0.152	0.354
Wood pellets	0.37	0.05	0.1	0.1	0.159	0.152	0.354
Wood briquettes	0.37	0.05	0.1	0.1	0.159	0.152	0.354
Black liquor	0.37	0.05	0.1	0.010244	0.159	0.152	0.354
Municipal waste	1.4	0.00304	0.00015	0.00016	0.022	0.001	0.000985
Special waste	9.2	14	0.6	0.2	1	31	25

¹ Applies to 2015 for petroleum products; the factors change yearly, in accordance with changes in the sulphur content in the products.

² For natural gas: 1000 Sm³.

³ Exceptions: Direct-fired furnaces in cement production = 9.1 and small stoves in households = 20.

⁴ From 1997 - considerably higher earlier years. Earlier used factors are not shown in this Appendix.

⁵ Applies to road traffic. Weighted average of duty-free and dutiable auto diesel.

⁶ Stationary combustion.

Source: Rosland (1987), (Norwegian pollution control authority (1990), (Sandgren *et al.* (1996), Finstad *et al.* (2001) Finstad *et al.* (2002) and Finstad and Rypdal (2003).

Table B2. Exceptions from the general emission factors for heavy metals: Solid fuels in small stoves (households)

	Pb g/tonne	Cd g/tonne	Hg g/tonne	As g/tonne	Cr g/tonne	Cu g/tonne
Coal	2.5	0.15	0.3	1.2	0.9	1.2
Coke	2.5	0.15	0.3	1.2	0.9	1.2

Table B3. Time series for variable emission factors for SO₂ (kg/tonne)

Years	V11 Motor gasoline	V13 Kerosene (heating)	V14 Jet kerosene	V15 Auto diesel				V17 Marine gas oil/diesel	V18 Light fuel oils	V19 Heavy distillate	V20 Heavy fuel oil (LS-oil)	V20 Heavy fuel oil (NS-oil)
	General	General	General	General	M.1A3B.1 Passenger cars	M.1A3B.2 Light duty vehicles	M.1A3B.3 Heavy duty vehicles	General	General	General	General	General
1980	1	0.2	0.2	6.6	.	.	.	6.6	6.6	15	19	46
1987	0.7	0.4	0.4	4.4	.	.	.	4.4	4.4	9	19	44
1989	0.6	0.4	0.4	3.4	.	.	.	3.4	3.4	7.6	18.2	40
1990	0.6	0.3	0.3	3.2	.	.	.	3.2	3.2	6	17	39.4
1991	0.6	0.38	0.38	2.8	.	.	.	2.8	2.8	4.6	16.8	43.6
1992	0.6	0.32	0.32	2.6	.	.	.	2.6	2.6	4.4	16.4	42.6
1993	0.6	0.42	0.42	2.2	.	.	.	2.2	2.2	4.4	16.2	45.8
1994	0.6	0.36	0.36	1.4	.	.	.	1.4	1.4	4.2	14.2	44.8
1995	0.24	0.46	0.46	1.4	.	.	.	1.4	1.4	4.6	11.8	43.4
1996	0.22	0.46	0.5	1.2	.	.	.	1.2	1.2	3.8	12.6	46.6
1997	0.16	0.46	0.46	1.2	.	.	.	1.2	1.2	3.8	12.6	47.2
1998	0.16	0.42	0.42	0.8	.	.	.	1.8	1.8	4.2	12.4	42.8
1999	0.22	0.32	0.32	0.6	.	.	.	1.6	1.6	4.4	12.8	39
2000	0.18	0.36	0.36	1.4	0.1174	0.1174	0.1174	1.8	1.8	4.6	14.4	31
2001	0.18	0.46	0.46	0.8	0.0885	0.0885	0.0885	1.8	1.8	4.8	13.2	44.4
2002	0.2	0.32	0.32	0.6	0.0708	0.0708	0.0708	1.6	1.2	4.8	12	43.8
2003	0.1	0.3	0.3	0.8	0.0748	0.0748	0.0748	2	0.8	4.6	14	44.2
2004	0.06	0.3	0.3	0.8	0.0748	0.0748	0.0748	1.8	0.8	5	14.2	44.2
2005	0.01	0.28	0.28	0.8	0.0278	0.0278	0.0278	1.8	0.8	4.6	13.6	39.2
2006	0.01	0.27	0.27	1.38	0.0393	0.0393	0.0393	2	1.38	4.44	10.4	26.2
2007	0.01	0.296	0.296	0.73	0.0244	0.0244	0.0244	1.53	0.73	4.17	17.8	20
2008	0.01	0.286	0.286	0.786	0.0285	0.0285	0.0285	1.562	0.986	3.098	17.5	28.5
2009	0.01	0.302	0.371	0.016	0.016	0.016	0.016	1.069	0.949	4.31	17.4	27.8
2010	0.01	0.324	0.294	0.015	0.015	0.015	0.015	1.184	0.978	4.31	17.5	28
2011	0.01	0.334	0.296	0.015	0.015	0.015	0.015	1.196	0.984	4.32	17.8	28.4
2012	0.01	0.326	0.294	0.015	0.015	0.015	0.015	1.038	0.658	4.295	17.5	27.4
2013	0.009	0.298	0.252	0.014	0.014	0.014	0.014	1.026	0.642	3.957	15.4	26.4
2014	0.01	0.342	0.252	0.014	0.014	0.014	0.014	1.054	0.648	4.263	15.5	27.0
2015	0.01	0.346	0.274	0.015	0.015	0.015	0.015	1.158	0.928	4.375	17.8	28.6

Table B4. Time series for variable emission factors for heavy metals, stationary combustion. g/tonne

Sector	Source	Fuel	1990-1991			1992-		
			Pb	Cd	Hg	Pb	Cd	Hg
General	S.03	V51	0.0085	0.00047	0.00035	0.00304	0.00015	0.00016

Aviation - NO_x, NMVOC, CO, particles and PAH

Table B5. General emission factors for aviation

Source	Fuel	NO _x kg/ tonne	NMVOC kg/ tonne	CO kg/ tonne	NH ₃ kg/ tonne	TSP, PM ₁₀ , PM _{2.5} kg/tonne	Dioxins µg I-TEQ/ tonne
M.1A3A.111 Jet/turboprop 0-100 m	V14 Jet kerosene	12.968	1.164	10.952	0	0.064	0.06
M.1A3A.112 Jet/turboprop 100-1000 m	V14 Jet kerosene	12.968	1.164	10.952	0	0.064	0.06
M.1A3A.12 Jet/turboprop cruise	V14 Jet kerosene	14.650	0.707	11.351	0	0.102	0.06
M.1A3A.211 Helicopter 0- 100 m	V14 Jet kerosene	6.67	28.8	36.6	0	0.025	0.06
M.1A3A.212 Helicopter 100-1000 m	V14 Jet kerosene	6.67	28.8	36.6	0	0.025	0.06
M.1A3A.22 Helicopter cruise	V14 Jet kerosene	6.67	32	36.6	0	0.007	0.06
M.1A3A.311 Small aircraft 0-100 m	V12 Aviation gasoline	12.968	1.164	10.952	0	0.064	2
M.1A3A.312 Small aircraft 100-1000 m	V12 Aviation gasoline	12.968	1.164	10.952	0	0.064	2
M.1A3A.32 Small aircraft cruise	V12 Aviation gasoline	14.650	0.707	11.351	0	0.102	2

Table B5 (cont.). General emission factors for aviation

Source	Fuel	benzo(a)pyrene g/tonne	benzo(b)fluoranthene g/tonne	benzo(k)fluoranthene g/tonne	indeno(1,2,3-cd)pyrene g/tonne
M.1A3A.111 Jet/turboprop 0-100 m	V14 Jet kerosene	NE	NE	NE	NE
M.1A3A.112 Jet/turboprop 100-1000 m	V14 Jet kerosene	NE	NE	NE	NE
M.1A3A.12 Jet/turboprop cruise	V14 Jet kerosene	NE	NE	NE	NE
M.1A3A.211 Helicopter 0-100 m	V14 Jet kerosene	NE	NE	NE	NE
M.1A3A.212 Helicopter 100-1000 m	V14 Jet kerosene	NE	NE	NE	NE
M.1A3A.22 Helicopter cruise	V14 Jet kerosene	NE	NE	NE	NE
M.1A3A.311 Small aircraft 0-100 m	V12 Aviation gasoline	0.005	0.009	0.003	0.011
M.1A3A.312 Small aircraft 100-1000 m	V12 Aviation gasoline	0.005	0.009	0.003	0.011
M.1A3A.32 Small aircraft cruise	V12 Aviation gasoline	0.005	0.009	0.003	0.011

Numbers in italics have exceptions for some sectors, see table B7, and bold numbers are different for different years, see table B8.

In the estimation update for NO_x, NMVOC, CO and particles, which was based on the new EEA (2013) factors, no distinction are made between flight phases in 0-100 m altitude and 100-1000 m altitude. Furthermore, emission factors for jet/turboprop and small aircraft are weighted together.

Source: Finstad *et al.* (2001), Finstad *et al.* (2002), EEA (2013). PAHs: Jet keorsone: EEA (2013), Aviation gasoline: Aarhus University (2016)

Table B6. Exceptions from the general factors for aviation

Component	Emission factor		Fuel	Source	Sectors
NO _x	13.51	V14	Jet kerosene	M.1A3A.111, M1A3A.211	248422
NO _x	13.29	V14	Jet kerosene	M.1A3A.112, M1A3A.212	248422
NO _x	11.7	V14	Jet kerosene	M.1A3A.12, M.1A3A.22	248422
NM VOC	7.43	V14	Jet kerosene	M.1A3A.111, M1A3A.211	248422
NM VOC	7.36	V14	Jet kerosene	M.1A3A.112, M1A3A.212	248422
NM VOC	4.3	V14	Jet kerosene	M.1A3A.12, M.1A3A.22	248422
CO	23.67	V14	Jet kerosene	M.1A3A.111, M1A3A.211	248422
CO	23.15	V14	Jet kerosene	M.1A3A.112, M1A3A.212	248422
CO	20.9	V14	Jet kerosene	M.1A3A.12, M.1A3A.22	248422
NO _x	12.559	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.11x, M1A3A.31x	235100.2N
NO _x	13.857	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.12, M1A3A.32	235100.2N
NM VOC	0.810	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.11x, M1A3A.31x	235100.2N
NM VOC	0.246	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.12, M1A3A.32	235100.2N
CO	9.903	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.11x, M1A3A.31x	235100.2N
CO	2.547	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.12, M1A3A.32	235100.2N
TSP, PM ₁₀ , PM _{2.5}	0.074	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.11x, M1A3A.31x	235100.2N
TSP, PM ₁₀ , PM _{2.5}	0.142	V12, 14	Aviation gasoline, jet kerosene	M.1A3A.12, M1A3A.32	235100.2N

Table B7. Time series for variable emission factors for aviation. Factors for 1989, 1995, 2000 and 2012 are calculated as given in the table. Factors for 1990-1994, 1996-1999 and 2001-2011 are calculated by linear interpolation. Factors after 2012 are kept constant. In the 2012 calculation source M.1A3A.111 and M.1A3A.112 are weighted together.

Component	Year	General			235100.2N		665100.2
		M.1A3A.111 (LTO 0-100 m)	M.1A3A.112 (LTO 100-1000 m)	M.1A3A.12 (cruise)	M.1A3A.111 (LTO 0-100 m)	M.1A3A.112 (LTO 100-1000 m)	M.1A3A.12 (cruise)
NO _x	1989	6.772	13.049	12.119	7.762	14.958	12.755
	1995	9.296	17.913	11.001	7.745	14.924	11.989
	2000	7.579	14.605	14.032	7.327	14.884	11.750
	2012	12.968	12.968	14.650	12.559	12.559	13.857
NM VOC	1989	0.775	0.127	0.554	0.365	0.060	0.675
	1995	7.725	1.265	0.963	0.773	0.127	3.369
	2000	1.576	0.258	0.507	1.293	0.221	0.366
	2012	1.164	1.164	0.707	0.810	0.810	0.246
CO	1989	19.768	2.145	6.947	14.173	1.538	4.191
	1995	27.204	2.952	12.147	15.118	1.640	8.459
	2000	21.239	2.305	7.808	16.925	2.659	3.866
	2012	10.952	10.952	11.351	9.903	9.903	2.547
TSP, PM ₁₀ , PM _{2.5}	1989	0.039	0.039	0.094	0.048	0.048	0.658
	1995	0.056	0.056	0.102	0.075	0.075	1.325
	2000	0.057	0.057	0.155	0.075	0.075	1.325
	2012	0.064	0.064	0.102	0.074	0.074	0.142

Road traffic - NO_x, NMVOC, CO, NH₃, particles and PAH*Table B8. General emission factors for road traffic*

Source	Fuel	NO _x kg/tonne	NMVOC kg/tonne	CO kg/tonne	NH ₃ kg/tonne	TSP, PM ₁₀ kg/tonne	PM _{2.5} kg/tonne	Dioxins µg I- TEQ/tonne
M.1A3B.1 Passenger car	V11 Motor gasoline	3.512	5.528	36.697	1.036	0.041	0.041	0.1
	V15 Auto diesel	9.222	0.485	2.695	0.019	0.273	0.259	0.1
	V31 Natural gas	0.871	0.065	1.693	0	0.122	0.122	0.05
	V32 LPG	1.163	0	11.999	0	0.033	0.033	0.06
M.1A3B.2 Other light duty cars	V11 Motor gasoline	6.386	9.704	100.020	0.821	0.083	0.083	0.1
	V15 Auto diesel	9.719	0.382	2.518	0.014	0.515	0.489	0.1
M.1A3B.3 Heavy duty vehicles	V11 Motor gasoline	27.969	16.767	22.266	0.018	0	0	0.1
	V15 Auto diesel	12.245	0.329	3.822	0.008	0.203	0.192	0.1
	V31/V37 Natural gas/ Biogas	8.274	0	5.901	0.008	0.026	0.026	0.05
M.1A3B.41 Moped	V11 Motor gasoline	3.254	108.829	193.621	0.052	0	0	0.1
M.1A3B.42 Motorcycle	V11 Motor gasoline	4.108	19.100	192.334	0.058	0	0	0.1

Table B8 (cont.). General emission factors for road traffic

Source	Fuel	benzo(a)pyrene g/tonne	benzo(b)fluoranthene g/tonne	benzo(k)fluoranthene g/tonne	indeno(1,2,3-cd)pyrene g/tonne
M.1A3B.1 Passenger car	V11 Motor gasoline	0.030	0.034	0.024	0.037
	V15 Auto diesel	0.114	0.127	0.100	0.106
	V31 Natural gas	0	0	0	0
	V32 LPG	0.026	0.030	0.021	0.033
M.1A3B.2 Other light duty cars	V11 Motor gasoline	0.029	0.035	0.024	0.038
	V15 Auto diesel	0.114	0.127	0.100	0.106
M.1A3B.3 Heavy duty vehicles	V11 Motor gasoline	0.014	0.083	0.092	0.021
	V15 Auto diesel	0.028	0.169	0.189	0.043
	V31/V37 Natural gas/ Biogas	0	0	0	0
M.1A3B.41 Moped	V11 Motor gasoline	0.040	0.040	NE	NE
M.1A3B.42 Motorcycle	V11 Motor gasoline	0.040	0.040	NE	NE

Bold numbers are different for different years, but only the 2015 data are shown in this Appendix, except for CH₄ (table B10) and N₂O (table B11).

Source: Results from Statistics Norway's use of HBEFA (INFRAS 2009), Finstad *et al.* (2001). PAH-profile: Aarhus University (2016)

Navigation - NO_x, NMVOC, CO, particles and POPs

Table B9. General emission factors for navigation

	NO _x kg/tonne	NMVOC kg/tonne	CO kg/tonne	NH ₃ kg/tonne	TSP, PM ₁₀ kg/tonne	PM _{2.5} kg/tonne	Dioxins µg I- TEQ/ tonne
V17 Marine gas oil/diesel, V18 Light fuel oils	43.76	2.4	2.9	0	1.6	1.5	4
V19 Heavy distillate, V20 Heavy fuel oil	43.76	2.4	2.9	0	5.4	5.1	4
V31 Natural gas (1000 Sm ³)	4.0	0.814	2.143	0	0.032	0.032	0.05

Table B9 (cont.). General emission factors for navigation

	benzo(a)pyrene g/tonne	benzo(b)fluoranthene g/tonne	benzo(k)fluoranthene g/tonne	indeno(1,2,3-cd)pyrene g/tonne
V17 Marine gas oil/diesel	0.006	0.028	0.013	0.051
V19 Heavy distillate	0.003	0.009	0.004	0.009
V20 Heavy fuel oil	0.003	0.008	0.004	0.008
V31 Natural gas (1000 Sm ³)	0.000025	0.000102	0.000039	0.000038

Numbers in italics have exceptions for some sectors, see table B13, and bold numbers are different for different years, see table B14-B16.

Source: Flugsrud and Rypdal (1996), Tørnsjø (2001), Finstad *et al.* (2001), Finstad *et al.* (2002b), Finstad *et al.* (2003), Bremnes Nielsen and Stenersen (2010). PAHs: Aarhus University (2016) and EEA (2013).

Table B10. Exceptions from the general factors for navigation

Component	Emission factor (kg/tonne)		Fuel	Sector
NO _x	36.60	V17, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	230310.N
NO _x	54	V17, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, Heavy fuel oil	230600.1 -230600.3
NO _x	46.58	V17, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	248422
NM VOC	1.4	V17, 18, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	230310.N
NM VOC	2.3	V17, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	248422
NM VOC	5	V17	Marine gas oil/diesel, light fuel oils	230600.1 -230600.3
NM VOC	5	V19, 20	Heavy distillate, heavy fuel oil	230600.1 -230600.3
CO	7.9	V17, 18, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	230310.N
CO	1.6	V17, 18, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	230600.1
CO	7	V17, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	230600.1 -230600.3
CO	2.3	V17, 19, 20	Marine gas oil/diesel, light fuel oils, heavy distillate, heavy fuel oil	248422

Table B11. Time series for variable emission factors for navigation. NO_x

Sector	Fuel	1990-1999	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999			
General	V17-20		56.85	56.80	56.89	56.77	56.82	56.68	57.23	57.47	57.41	56.82			
			2000	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015
General	V17-20		57.82	56.51	55.90	55.55	54.61	53.25	51.90	50.54	49.18	47.83	46.47	45.11	43.76
230310.N	V17, 19, 20	52.11	52.12	51.69	51.58	51.48	50.93	49.90	47.41	45.17	43.64	43.36	40.94	37.97	36.60
248422	V17, 19, 20	50.17	49.82	48.95	48.74	48.52	48.31	48.09	47.88	47.66	47.44	47.23	47.01	46.80	46.58

Source: Flugsrud *et al.* (2010)

Table B12. Time series for variable emission factors for navigation. NMVOC and CO

Sector	Fuel	NMVOC						CO	
		1980-1990	1980-1997	1980-1998	1991-	1998-	1999-	1980-1997	1998-
General	V17-20							3.1	2.9
230310.N	V17-20			1.5			1.4		
230600.1	V17-20							2	1.6
230600.1,230910	V19,20	6.4			5				
230600.1,230910	V 20								
248422	V17-20		2.2			2.3			

Other mobile sources including railways - NO_x, NMVOC, CO, NH₃, particles and POPs

Table B13. General emission factors for other mobile sources

		NO _x kg/ tonne	NMVOC kg/ tonne	CO kg/ tonne	NH ₃ kg/ Tonne	TSP, PM ₁₀ kg/ tonne	PM _{2.5} kg/ tonne	Dioxins µg I- TEQ/ tonne
	V01 Coal	3	1.1	3	0	1.6/1.14	0.82	1.6
M.1A3C Railway	V15 Auto diesel	47	4	11	0.007	3.8	3.8	0.1
M.1A3E.21 Small boats 2 stroke	V11 Motor gasoline	6	240	415	0	8	8	0.1
M.1A3E.22 Small boats 4 stroke	V11 Motor gasoline	12	40	1 000	0	1	1	0.1
	V15 Auto diesel	54	27	25	0	4	4	0.1
M.1A3E.31 Motorized equipment 2 stroke	V11 Motor gasoline	2¹	500	700	0	8	8	0.1
	V11 Motor gasoline	10	110	1 200	0	1	1	0.1
M.1A3E.32 Motorized equipment 4t	V15 Auto diesel	13.0	1.0	6.1	0.008	0.2	0.2	0.1
	V18 Light fuel oils	50	6	15	0.005	7.1	6.75	0.1

Other mobile sources including railways - NO_x, NMVOC, CO, NH₃, particles and POPs

Table B13 (cont.). General emission factors for other mobile sources

		benzo(a)pyrene g/tonne	benzo(b)fluoranthene g/tonne	benzo(k)fluoranthene g/tonne	indeno(1,2,3-cd)pyrene g/tonne
M.1A3C Railway	V01 Coal	0.007	0.01	0.004	0.003
	V15 Auto diesel	0.030	0.050	0	0
M.1A3E.21 Small boats 2 stroke	V11 Motor gasoline	0.040	0.040	0	0
M.1A3E.22 Small boats 4 stroke	V11 Motor gasoline	0.040	0.040	0	0
	V15 Auto diesel	0.030	0.050	0	0
M.1A3E.31 Motorized equipment 2 stroke	V11 Motor gasoline	0.040	0.040	0	0
M.1A3E.32 Motorized equipment 4t	V11 Motor gasoline	0.040	0.040	0	0
	V15 Auto diesel	0.030	0.050	0	0
	V18 Light fuel oils	0.030	0.050	0	0

M.1A3E.1 Snow scooter has the same emission factors as M.1A3B.41 Moped, see table B9.

Bold numbers are different for different years, but only 2015 figures are presented here.

¹Before 1995 the emission factor was 1.3.

Numbers in italics have exceptions for some sectors, see table B18–B19.

Sources: Bang (1993), Bang *et al.* (1999), Finstad *et al.* (2001), Finstad *et al.* (2002b), Finstad *et al.* (2003), Winther and Nielsen (2006), EEA (2013).

Table B14. Exceptions from the general factors for greenhouse gases and precursors for other mobile sources

Component	Emission factor (kg/tonne)	Fuel	Source	Sectors
NO _x	15.1	V15	Auto diesel	M.1A3E.32 Motorized equipment 4 stroke 230100-230210
NO _x	54	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230100
NO _x	52	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230210
NO _x	47	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230710-230892, 234910
NO _x	48	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 232360, 248422
NO _x	46	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 234110-234120
NM VOC	1.8	V15	Auto diesel	M.1A3E.32 Motorized equipment 4 stroke 230100-230200
NM VOC	7.2	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230100
NM VOC	5.7	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230210
NM VOC	4	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230710-230892, 234910
NM VOC	4.8	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 232360, 248422
NM VOC	3.8	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 234110-234120
CO	10.0	V15	Auto diesel	M.1A3E.32 Motorized equipment 4 stroke 230100-230210
CO	25	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230100
CO	20	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230210
CO	11	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 230710-230892, 234910
CO	17	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 234110-234120
CO	18	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke 248422

Bold numbers are different for different years, time series for NO_x are presented in table B20.

Table B15. Exceptions from the general factors for other pollutants for other mobile sources

Component	Emission factor (kg/tonne)		Fuel	Source	Sectors
TSP, PM ₁₀	1.7	V15	Auto diesel	M.1A3E.32 Motorized equipment 4 stroke	230100-230210
TSP, PM ₁₀	3.8	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	230710-230892, 234910
TSP, PM ₁₀	4.2	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	232360
TSP, PM ₁₀	5.3	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	234110-234120
TSP, PM ₁₀	5.4	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	248422
PM _{2.5}	1.6	V15	Auto diesel	M.1A3E.32 Motorized equipment 4 stroke	230100-230210
PM _{2.5}	3.61	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	230710-230892, 234910
PM _{2.5}	3.99	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	232360
PM _{2.5}	5.04	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	234110-234120
PM _{2.5}	5.13	V18	Light fuel oils	M.1A3E.32 Motorized equipment 4 stroke	248422

Bold numbers are different for different years, but only 2015 figures are presented here.

Table B16. Time series for NO_x emission factors for use of auto diesel in motorized equipment 4t

Sector	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004
General	46.3	46.4	46.4	46.4	46.5	46.6	46.7	46.8	46.0	43.9	41.9	40.2	37.8	35.0	31.8
230100-230210	27.1	27.0	26.7	26.5	26.4	26.4	26.5	26.6	26.6	26.5	26.4	26.2	25.8	25.4	24.9

Sector	2005	2006	2007	2008	2009	2010	2011	2012	2013	2014	2015
General	28.6	24.8	21.6	19.8	18.5	17.1	16.1	15.3	14.5	13.5	13.0
230100-230210	24.1	23.3	22.3	21.3	20.3	19.3	18.3	17.5	16.8	15.9	15.1

Source: Winther and Nielsen (2006). Data for 2005 and later are extrapolations.

Table B17. Time series for variable emission factors for other mobile sources

Fuel	Component	1980-1990	1991	1992	1993	1994	1995	1996	1997-
V11 Gasoline	Dioxins	1.32	1.11	0.95	0.69	0.25	0.23	0.11	0.1

NO_x - Stationary combustion

Table B18. General emission factors. kg NO_x/tonne fuel

Source	V01 Coal	V02 Coke	V03 Petrol coke	V41 Fuel wood	V42 Wood waste	V43 Black liquor	V44 Wood pellets	V45 Wood bri- quettes	V04 Char- coal	V31 Natural gas (1000 Sm ³)	V33 Refinery gas
S.01 Direct- fired furnaces	16	20	20	2.68	5.95	5.4
S.02 Gas turbines	6.27	.
S.03 Boilers ..	3	3	3.4	.	0.9	0.9	1.3	1.3	.	2.55	3
S.04 Small stoves ...	3	3	.	0.988	.	.	1.1	.	1.4	.	.
S.1B2C Flares	12	7
	V34 Blast furnace gas	V36 Landfill gas	V35 Fuel gas	V32 LPG	V13 Kerosene (heating)	V17 Marine gas oil/ diesel	V18 Light fuel oils	V19 Heavy distillate	V20 Heavy fuel oil	V51 Municipal waste	V52 Special waste
S.01 Direct- fired furnaces	5.4	.	5.4	.	.	70	.	5	5	.	5
S.02 Gas turbines	16
S.03 Boilers	3	0.01	3	2.3	3	2.5	2.5	2.5	4.2	1.365	4.2
S.04 Small stoves	.	.	.	2.3	2.5	.	2.5	2.5	.	.	.
S.1B2C Flares	.	0.17

Numbers in italics have exceptions for some sectors, see table B27, and bold numbers are different for different years, see table B28.

Source: Rosland (1987). Fuel wood factor based on data from annual surveys on use of fuel wood in households.

Table B19. Exceptions from the general factors for NO_x. Stationary combustion. kg NO_x/tonne fuel

Emission factor		Fuel	Source	Sectors
24	V19, 20, 52	Heavy distillate, heavy fuel oil, special waste	S.01 Direct-fired furnaces	231910.2, 232350
6.13	V31	Natural gas (1000 Sm ³)	S.01 Direct-fired furnaces	232014
9.5	V19, 20	Heavy distillate, heavy fuel oil	S.01 Direct-fired furnaces	232360
8.681	V31	Natural gas (1000 Sm ³)	S.02 Gas turbines	230600.1
1,4	V31	Natural gas (1000 Sm ³)	S.1B2C Flares	230600.1
3	V17, 18, 19	Fuel oils	S.03 Boilers	230500-233320
4.5	V01	Coal	S.03 Boilers	230500-233320
3.4	V02	Coke	S.03 Boilers	230500-233320
5	V20, 52	Heavy fuel oil, special waste	S.03 Boilers	230500-233320
2.9	V35	Fuel gas	S.03 Boilers	232011-232050, 232411-232442
0.01	V34	CO gas	S.03 Boilers	233510-233530
6.27	V33	Refinery gas	S.02 Gas turbines	233511
1.4	V01, 02	Coal, coke	S.04 Small stoves	330000

Table B20. Time series for variable emission factors for NO_x. Stationary combustion. kg NO_x/tonne fuel

Sector	Source	Fuel	1980	1991	1992	1995	1996	1999	2005	2006	2007	2008	2009	2010
			-		-		-	-						
			1990		1994		1998	2004						
General	S.04	V41	0.98		0.98	0.98		0.98	0.98	0.98				0.98
			2	0.981	2	1	0.982	1	5	4	0.987	0.988	0.987	8

Sector	Source	Fuel	2011	2012	2013	2014	2015
General	S.04	V41	0.98		0.98	0.989	0.98
			6	0.985	8		8

			1980										
			-										
Sector	Source	Fuel	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000-
230600.			8.22		8.23	8.44	8.61	8.87	9.12	9.18	9.52	9.08	
1	S.02	V31	3	8.172	4	4	7	4	8	5	8	7	8.681

NMVOC - Stationary combustion

Table B21. General emission factors. kg NMVOC/tonne fuel

Source	V01 Coal	V02 Coke	V03 Petrol coke	V41 Fuel wood	V42 Wood waste	V43 Black liquor	V44 Wood pellets	V45 Wood bri- quettes	V04 Char coal	V31 Natural gas (1000 Sm ³)	V33 Re finery gas
S.01 Direct-fired furnaces	0	0	0	8.85	0	0.1
S.02 Gas turbines	0.24	.
S.03 Boilers ...	1.1	0.6	0.6	.	1.30	.	1.3	1.3	.	0.085	0.1
S.04 Small stoves	1.1	0.6	.	7.0	.	.	6.501	.	10	.	.
S.1B2C Flares	0.06	13.5
	V34 CO gas	V36 Landfill gas	V35 Fuel gas	V32 LPG	V13 Kero sene (heating)	V17 Marine gas oil/ diesel	V18 Light fuel oils	V19 Heavy distillate	V20 Heavy fuel oil	V51 Munici- pal waste	V52 Special waste
S.01 Direct-fired furnaces	0	.	0	.	.	5	.	0.3	0.3	.	0.3
S.02 Gas turbines	0.03
S.03 Boilers	0.1	0	0.1	0.1	0.4	0.4	0.4	0.4	0.3	0.7	0.3
S.04 Small stoves	.	.	.	0.1	0.4	.	0.4	0.4	.	.	.
S.1B2C Flares	.	0

Numbers in italics have exceptions for some sectors, see table B30.

Source: Rosland (1987) and Sandgren *et al.* (1996).

Table B22. Exceptions from the general factors for NMVOC. Stationary combustion. kg NMVOC/tonne fuel

Emission factor	Fuel	Source	Sectors
0	V 19, 20, 52	Heavy distillate, heavy fuel oil, special waste	S.01 Direct-fired furnaces 231910.2, 232350
0.1	V34	CO gas	S.01 Direct-fired furnaces 231910.2
0.085034	V31	Natural gas (1000 Sm ³)	S.01 Direct-fired furnaces 232014
0.9	V19, 20	Heavy distillate, heavy fuel oil	S.01 Direct-fired furnaces 232360
0.8	V01	Coal	S.03 Boilers 230500-233320
0	V32, 34, 35, 42	LPG, CO gas, fuel gas, wood waste	230500-233320, 231711, 232011-232050, 233510-233530
0.6	V17, 18, 19	Fuel oils	S.03 Boilers 330000
10	V01	Coal	S.04 Small stoves 330000
0.6	V13	Kerosene (heating)	S.04 Small stoves 330000

CO - Stationary combustion

Table B23. General emission factors. kg CO/tonne fuel

Source	V01 Coal	V02 Coke	V03 Petrol coke	V41 Fuel wood	V42 Wood waste	V43 Black liquor	V44 Wood pellets	V45 Wood briquette s	V04 Char coal	V31 Natural gas (1000 Sm ³)	V33 Re finery gas
S.01 Direct- fired furnaces	0	26.16	0	16.82.	0	0
S.02 Gas turbines	1.7	.
S.03 Boilers ..	3	26.16	3	.	15	0	15	15	.	0	0
S.04 Small stoves ...	3	26.16	.	93.4	.	.	2.6	.	100	.	.
S.1B2C Flares	1.5	0
	V34 CO gas	V36 Landfill gas	V35 Fuel gas	V32 LPG	V13 Kero sene (heating)	V17 Marine gas oil/ diesel	V18 Light fuel oils	V19 Heavy distillate	V20 Heavy fuel oil	V51 Municipal waste	V52 Special waste
S.01 Direct- fired furnaces	0	.	0	.	.	5	.	0.2	0.2	.	0.2
S.02 Gas turbines	0.7
S.03 Boilers ..	0	0	0	0.5	2	2	2	2	0.4	2.8	0.4
S.04 Small stoves	.	.	.	0.5	2	.	2	2	.	.	.
S.1B2C Flares	.	0.04

Numbers in italics have exceptions for some sectors, see table B32, and bold numbers are different for different years, see table B33.

Table B24. Exceptions from the general factors for CO. Stationary combustion. kg CO/tonne fuel

Emission factor	Fuel	Source	Sectors
0	V 19, 20, 52	Heavy distillate, heavy fuel oil, special waste	S.01 Direct-fired furnaces
0.01	V34	CO gas	S.01 Direct-fired furnaces
0.2	V20, 52	Heavy fuel oil, special waste	S.03 Boilers
0	V32, 42	LPG, wood waste	S.03 Boilers
6.5	V17, 18, 19	Fuel oils	S.03 Boilers
100	V01, 02	Coal, coke	S.04 Small stoves
6.5	V13	Kerosene (heating)	S.04 Small stoves
1.7	V31	Natural gas (1000 Sm ³)	S.1B2C Flares

Table B25. Time series for variable emission factors for CO. Stationary combustion. kg CO/tonne fuel

Sector	Source	Fuel	1980-												
			1997	1998	1999	2000	2001	2002	2003	2004	2005	2006	2007	2008	2009
General	S.04	V41	149.1	148.4	146.3	142.6	137.6	131.0	122.2	111.5	115.5	111.9	110.6	107.9	105.0
Sector	Source	Fuel	2010	2011	2012	2013	2014	2015							
			General	S.04	V41	103.3	101.2	99.2	96.1	96.7	93.4				

NH₃ - Stationary combustion

Table B26. General emission factors. kg NH₃/tonne fuel

Source	V01	V02	V03	V41	V42	V43	V44	V45	V04	V31	V33	V34	V36	V35	V32	V13	V17	V18	V19	V20	V51	V52
	Coal	Coke	Petrol	Fuel	Wood	Black	Wood	Wood	Char-	Natural	Re-	CO	Land-	Fuel	LPG	Kero-	Marine	Light	Heavy	Heavy	Munici-	Special
			coke	wood	waste	liquor	pellets	bri-	coal	gas	finery	gas	fill	gas	gas	sene	gas	fuel	dis-	fuel	oil	pal
								quettes		(1000	gas					(heating)	oil/	oils	tillate		waste	
										Sm ³)							diesel					
S.04																						
Small																						
stoves	.	.	.	0.066	.	.	0.066	.	0
All other																						
sources	0	0	0	.	0	0	0	0	1.09	0	0	0	0	0	0	0	0	0	0	0	0	0

Particulate matter - Stationary combustion

Table B27. General emission factors. kg particle component/tonne fuel

Com- ponent	Source	V01 Coal	V02 Coke	V03 Petrol coke	V41 Fuel wood	V42 Wood waste	V43 Black liquor	V44 Wood pellets	V45 Wood briquettes	V04 Char coal	V31 Natural gas (1000 Sm ³)	V33 Refinery gas
	S.01 Direct-fired furnaces	1.6	1.6	1.6	4.43.	0.122	0.144
TSP ...	S.02 Gas turbines	0.122	.
TSP ...	S.03 Boilers	1.6	1.6	1.6	.	2.69	0	2.69	2.69	.	0.122	0.144
TSP ...	S.04 Small stoves	4.2	2.85	3.5	17.16	.	.	1.1	.	2.4	.	.
TSP ...	S.1B2C Flares	0.002	0.144
	S.01 Direct-fired furnaces	1.14	1.14	1.14	4.22	0.122	0.144
PM ₁₀ ..	S.02 Gas turbines	0.122	.
PM ₁₀ ..	S.03 Boilers	1.14	1.14	1.14	.	2.52	0	2.52	2.52	.	0.122	0.144
PM ₁₀ ..	S.04 Small stoves	2.8	1.71	2.1	16.82	.	.	1.1	.	2.4	.	.
PM ₁₀ ..	S.1B2C Flares	0.002	0.144
	S.01 Direct-fired furnaces	0.82	0.82	0.82	4.13	0.122	0.144
PM _{2.5} .	S.02 Gas turbines	0.122	.
PM _{2.5} .	S.03 Boilers	0.82	0.82	0.82	.	2.52	0	2.52	2.52	.	0.122	0.144
PM _{2.5} .	S.04 Small stoves	0.86	0.86	1.5	16.31	.	.	1.1	.	2.4	.	.
PM _{2.5} .	S.1B2C Flares	0.002	0.144
		V34 CO gas	V36 Landfill gas	V35 Fuel gas	V32 LPG	V13 Kerosene (heating)	V17 Marine gas oil/ diesel	V18 Light fuel oils	V19 Heavy distillate	V20 Heavy fuel oil	V51 Municipal waste	V52 Special waste
	S.01 Direct-fired furnaces	0.144	.	0.144	.	.	0.286	.	*	*	.	5.68
TSP ...	S.02 Gas turbines	0.286
TSP ...	S.03 Boilers	0.144	0.144	0.144	0.136	0.296	0.286	0.286	*	*	0.05	5.68
TSP ...	S.04 Small stoves	.	.	.	0.136	0.3	.	0.3
TSP ...	S.1B2C Flares	.	0.144
	S.01 Direct-fired furnaces	0.144	.	0.144	.	.	0.143	.	*	*	.	4.87
PM ₁₀ ..	S.02 Gas turbines	0.143
PM ₁₀ ..	S.03 Boilers	0.144	0.144	0.144	0.136	0.148	0.143	0.15	*	*	0.05	4.87
PM ₁₀ ..	S.04 Small stoves	.	.	.	0.136	0.16	.	0.155
PM ₁₀ ..	S.1B2C Flares	.	0.144
	S.01 Direct-fired furnaces	0.144	.	0.144	.	.	0.036	.	*	*	.	3.2

PM _{2.5}	S.02 Gas turbines	0.036
PM _{2.5}	S.03 Boilers	0.144	0.144	0.144	0.136	0.037	0.12	0.12	*	*	0.05	3.2
PM _{2.5}	S.04 Small stoves	.	.	.	0.136	0.12	.	0.119
PM _{2.5}	S.1B2C Flares	.	0.144

Numbers in italics have exceptions for some sectors, see table B37, and bold numbers are different for different years, see table B38.

* General emission factors for all sources for heavy distillate and heavy fuel oil are given in table B36 for all years.

Source: Finstad *et al.* (2003). Fuel wood factor based on data from annual surveys on use of fuel wood in households

Table B28. General particle emission factors for heavy distillate and heavy fuel oil for all sources. Factors dependent on sulphur content. kg particle component /tonne fuel

Fuel	Compo- nent	1990	1991	1992	1993	1994	1995	1996-1997	1998	1999	2000-
V19	TSP	0.803	0.714	0.701	0.701	0.688	0.714	0.663	0.688	0.701	0.714
	PM ₁₀	0.690	0.614	0.603	0.603	0.592	0.614	0.570	0.592	0.603	0.614
	PM _{2.5}	0.450	0.400	0.393	0.393	0.385	0.400	0.371	0.385	0.393	0.400
	TSP	1.350	1.339	1.316	1.304	1.190	1.053	1.098	1.087	1.110	1.201
V20	PM ₁₀	1.161	1.151	1.131	1.121	1.023	0.905	0.944	0.934	0.954	1.033
	PM _{2.5}	0.761	0.754	0.741	0.735	0.671	0.593	0.619	0.613	0.625	0.677

Source: Finstad *et al.* (2003).

Table B29. Exceptions from the general factors for particles. Stationary combustion

Emission factor (kg TSP/tonne)	Emission factor (kg PM ₁₀ /tonne)	Emission factor (kg PM _{2.5} /tonne)	Fuel	Source	Sectors
4.06	2.4	1.4	V52	Special waste	S.01 Direct-fired furnaces 230500-233320
5.45	3.54	1.45	V01	Coal	S.01 Direct-fired furnaces 233530
4.2	2.8	0.86	V01	Coal	S.03 Boilers 230100
.	0.143 (V18)	0.036 (V17, 18)	V17, 18	Light fuel oils	S.03 Boilers 230500-233320
4.06	2.4	1.4	V52	Special waste	S.03 Boilers 230500-233320
5.45	3.54	1.45	V01	Coal	S.03 Boilers 233530
0.5	0.5	0.5	V51	Municipal waste	S.03 Boilers 253800
0.3	0.155	0.119	V13	Kerosene (heating)	S.04 Small stoves 330000

Table B30. Time series for variable emission factors for particles. Stationary combustion. kg particle component /tonne fuel

Componen t	Sourc e	Fuel	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001
TSP	S.04	V41	22.24	22.24	22.24	22.25	22.26	22.25	22.26	22.27	22.24	22.05	21.68	21.22
PM ₁₀	S.04	V41	21.80	21.80	21.80	21.80	21.81	21.81	21.81	21.83	21.79	21.61	21.25	20.79
PM _{2.5}	S.04	V41	21.13	21.13	21.13	21.14	21.15	21.14	21.15	21.16	21.12	20.95	20.60	20.16
			2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013
TSP	S.04	V41	20.62	19.82	18.85	19.10	18.80	18.66	18.41	18.16	17.96	17.78	17.65	17.33
PM ₁₀	S.04	V41	20.21	19.42	18.47	18.72	18.42	18.29	18.04	17.79	17.60	17.43	17.29	16.98
PM _{2.5}	S.04	V41	19.59	18.83	17.91	18.15	17.86	17.73	17.49	17.25	17.06	16.89	16.77	16.46
			2014	2015										
TSP	S.04	V41	17.35	17.16										
PM ₁₀	S.04	V41	17.00	16.82										
PM _{2.5}	S.04	V41	16.48	16.31										

POPs (Persistent Organic Pollutants) - Stationary combustion

Table B31. General emission factors for PAH

Component	Source	V01 Coal	V02 Coke	V03 Petrol coke	V41 Fuel wood	V42 Wood waste	V43 Black liquor	V44 Wood pellets	V45 Wood briquettes	V04 Charcoa I (1000 Sm ³)	V31 Natural gas	V33 Refinery gas
benzo(a) pyrene g/tonne	S.01 Direct- fired furnaces	0.00002	0.0000 2	0.0000 2	0.007	0.00002	0.00003
benzo(a) pyrene g/tonne	S.02 Gas turbines	0.00002	.
benzo(a) pyrene g/tonne	S.03 Boilers	0.007	0.007	0.007	.	0.0001	0.0001	0.0001	0.0001	.	0.00002	0.00003
benzo(a) pyrene g/tonne	S.04 Small stoves	2.81	2.85	3.5	0.312	.	.	2.091	2.091.	3.5695	.	.
benzo(b) fluoranthene g/tonne	S.01 Direct- fired furnaces	0.001	0.001	0.001	0.010	0.00003	0.00004
benzo(b) fluoranthene g/tonne	S.02 Gas turbines	0.00003	.
benzo(b) fluoranthene g/tonne	S.03 Boilers	0.010	0.010	0.010	.	0.0075	0.0075	0.0075	0.0075	.	0.00003	0.00004
benzo(b) fluoranthene g/tonne	S.04 Small stoves	4.777	4.845	5.95	0.492	.	.	1.918	1.918	3.2745	.	.
benzo(k) fluoranthene g/tonne	S.01 Direct- fired furnaces	0.0008	0.0008	0.001	0.004	0.00003	0.00004
benzo(k) fluoranthene g/tonne	S.02 Gas turbines	0.00003	.
benzo(k) fluoranthene g/tonne	S.03 Boilers	0.004	0.004	0.004	.	0.0075	0.0075	0.0075	0.0075	.	0.00003	0.00004
benzo(k) fluoranthene g/tonne	S.04 Small stoves	3.653	2.85	3.5	0.115	.	.	0.726	0.726.	1.239	.	.
indeno(1,2,3-cd) pyrene g/tonne	S.01 Direct- fired furnaces	0.00003	0.0000 3	0.0000 4	0.003	0.00003	0.00004

indeno(1,2,3_cd) S.02												
pyrene Gas												
g/tonne turbines	0.00003	.
indeno(1,2,3_cd) S.03												
pyrene Boilers	0.003	0.003	0.003	.	0.0002	0.0002	0.0002	0.0002	.	0.00003	0.00004	
g/tonne												
indeno(1,2,3_cd) S.04												
pyrene Small												
g/tonne stoves	2.248	2.28	2.8	0.225	.	.	1.227	1.227	.	2.0945	.	.

Table B31 (cont.). General emission factors for PAH

Component	Source	V34 Blast furnace gas	V36 Landfill gas	V35 Fuel gas	V32 LPG	V13 Kero- sene (heating)	V17 Marine gas oil/ diesel	V18 Light fuel oils	V19 Heavy distillate	V20 Heavy fuel oil	V51 Municipal waste ¹	V52 Special waste
benzo(a) Direct- pyrene fired g/tonne . furnaces	S.01	0.00001	.	0.0000 4	.	.	NE	.	NE	NE	.	0.077
benzo(a)pyrene g/tonne .	S.02 Gas turbines	.	0.00003	.	.	.	NE
benzo(a)pyrene g/tonne .	S.03 Boilers	0.00001	0.00003	0.0000 4	2.5816E-08	0.00002	NE	0.00002	NE	NE	0.00001	0.077
benzo(a)pyrene g/tonne .	S.04 Small stoves	.	.	. 2.5816E-08	0.003	.	0.003
benzo(b) Direct- fluoranthene fired g/tonne . furnaces	S.01	0.00003	.	0.0001 5	.	.	0.00019	.	0.00019	0.00018	.	0.609
benzo(b) Direct- fluoranthene fired g/tonne . furnaces	S.02 Gas turbines	.	0.00004	.	.	.	0.00019
benzo(b) Direct- fluoranthene fired g/tonne . furnaces	S.03 Boilers	0.00001	0.00004	0.0001 5	3.8724E-08	0.00001	0.00003	0.00001	0.00019	0.00018	0.00002	0.609
benzo(b) Direct- fluoranthene fired g/tonne . furnaces	S.04 Small stoves	.	.	. 3.8724E-08	0.002	.	0.002
benzo(k) Direct- fluoranthene fired g/tonne . furnaces	S.01	0.00001	.	0.0000 6	.	.	0.00019	.	0.00019	0.00018	.	0.069
benzo(k) Direct- fluoranthene fired g/tonne . furnaces	S.02 Gas turbines	.	0.00004	.	.	.	0.00019
benzo(k) Direct- fluoranthene fired g/tonne . furnaces	S.03 Boilers	0.00001	0.00004	0.0000 6	3.8724E-08	0.00002	0.00003	0.00002	0.00019	0.00018	0.00001	0.069
benzo(k) Direct- fluoranthene fired g/tonne . furnaces	S.04 Small stoves	.	.	. 3.8724E-08	0.003	.	0.003
indeno(1,2,3-cd) Direct- pyrene fired g/tonne . furnaces	S.01	0.00001	.	0.0000 5	.	.	0.0003	.	0.00030	0.00028	.	0.061

indeno(1,2,3_cd)	S.02														
pyrene	Gas														
g/tonne	turbines		0.00004					0.0003							
indeno(1,2,3_cd)	S.03														
pyrene	Boilers	0.00001	0.00004	0.0000	5	3.8724E-08	0.00005	0.00004	0.00005	0.00030	0.00028	0.00001	0.061		
indeno(1,2,3_cd)	S.04														
pyrene	Small														
g/tonne	stoves					3.8724E-08	0.007		0.007						

Bold numbers are different for different years, see table B40. NE = Not estimated. ¹Emission factor used for the years after 1995. Emission factors for the years 1990 to 1994 can be given on request.

Source: Finstad *et al.* (2001). Fuel wood factor based on data from annual surveys on use of fuel wood in households. EEA (2013), EEA (2016), . Allerup *et. al* (2015)

Table B32. Time series for variable emission factors for PAH¹. Stationary combustion (g component /tonne fuel)

Component	Source	Fuel	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001
benzo(a)pyrene	S.04	V41	0.74	0.74	0.74	0.74	0.74	0.74	0.74	0.74	0.74	0.72	0.69	0.66
benzo(b)fluoranthene	S.04	V41	1.17	1.17	1.17	1.17	1.17	1.17	1.17	1.17	1.16	1.13	1.09	1.03
benzo(k)fluoranthene	S.04	V41	0.27	0.27	0.27	0.27	0.27	0.27	0.27	0.27	0.27	0.26	0.25	0.24
indeno(1,2,3_cd)pyrene	S.04	V41	0.53	0.53	0.53	0.53	0.53	0.53	0.53	0.53	0.53	0.52	0.50	0.47
			2002	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013
benzo(a)pyrene	S.04	V41	0.61	0.54	0.46	0.50	0.47	0.46	0.44	0.42	0.41	0.39	0.38	0.36
benzo(b)fluoranthene	S.04	V41	0.96	0.85	0.73	0.78	0.74	0.73	0.70	0.66	0.64	0.62	0.59	0.59
benzo(k)fluoranthene	S.04	V41	0.22	0.20	0.17	0.18	0.17	0.17	0.16	0.15	0.15	0.14	0.14	0.13
indeno(1,2,3_cd)pyrene	S.04	V41	0.44	0.39	0.33	0.36	0.34	0.33	0.32	0.30	0.29	0.28	0.27	0.26
			2014	2015										
benzo(a)pyrene	S.04	V41	0.36	0.31										
benzo(b)fluoranthene	S.04	V41	0.57	0.49										
benzo(k)fluoranthene	S.04	V41	0.13	0.12										
indeno(1,2,3_cd)pyrene	S.04	V41	0.26	0.23										

Source: Emission factor: Finstad *et al.* (2001). PAH-profile: EEA (2013)

POPs (Persistent Organic Pollutants) - Stationary combustion

Table B33. General emission factors for dioxins

Com- ponent	Source	V01 Coal	V02 Coke	V03 Petrol coke	V41 Fuel wood	V42 Wood waste	V43 Black liquor	V44 Wood pellets	V45 Wood bri- quettes	V04 Char- coal	V31 Natural gas (1000 Sm ³)	V33 Refinery gas
Dioxins µg I- TEQ/tonne ..	S.01 Direct- fired furnaces	1.6	1.6	1.6	2.95.	0.05	0
Dioxins µg I- TEQ /tonne ..	S.02 Gas turbines	0.05	.
Dioxins µg I- TEQ /tonne ..	S.03 Boilers	1.6	1.6	1.6	.	1	1	1	1	.	0.05	0
Dioxins µg I- TEQ /tonne ..	S.04 Small stoves	10	10	10	5.9	.	.	5.9	.	10	.	.
Dioxins µg I- TEQ /tonne ..	S.1B2C Flares	0.05	0
		V34 CO gas	V36 Landfill gas	V35 Fuel gas	V32 LPG	V13 Kero- sene (heating)	V17 Marine gas oil/ diesel	V18 Light fuel oils	V19 Heavy distillate	V20 Heavy fuel oil	V51 Munici- pal waste	V52 Special waste
Dioxins µg I- TEQ /tonne ..	S.01 Direct- fired furnaces	0	.	0	.	.	4	.	0.1	0.1	.	4
Dioxins µg I- TEQ /tonne ..	S.02 Gas turbines	4
Dioxins µg I- TEQ /tonne ..	S.03 Boilers	0	0	1	0.06	0.1	0.1	0.1	0.1	0.1	0.02	4
Dioxins µg I- TEQ /tonne ..	S.04 Small stoves	.	.	.	0.06	0.06	.	0.2
Dioxins µg I- TEQ /tonne ..	S.1B2C Flares	.	0

Numbers in italics have exceptions for some sectors, see table B42.

Source: Finstad *et al.* (2002).

Table B34. Exceptions from the general factors for POPs. Stationary combustion

Emission factor (ug dioxin/tonne)	Fuel	Source	Sectors
0.2	V18, 19 Heavy distillate, heavy fuel oil	S.03 Boilers	330000

The emission factors for BC, PCB and HCB vary greatly between sectors, combustion technology and years. Table B35 presents some of the factors. In a number of cases, particularly for PCB and HCB, factors are not available and emissions have not been estimated.

Table B35. Emission factors for BC, PCB and HCB

	BC kg/tonne	PCB ug/tonn	HCB ug/tonn
Coal	0.016 ¹	161.170 ¹	17.422
Coke	0.016 ¹	161.170 ¹	17.422
Petrol coke	0.016 ¹	161.170 ¹	NE
Charcoal	0.180 ¹	161.170 ¹	0.150 ¹
Motor gasoline	0.009 ²	20 ³	0.422 ²
Aviation gasoline	0.005 ⁴	NE	NE
Kerosene (heating)	0.086 ⁵	4.008	NE
Jet kerosene	0.005 ⁴	NE	NE
Auto diesel	0.196 ²	986.339 ²	456.017 ²
Marine gas oil/diesel	0.600 ⁶	355.656	80
Light fuel oils	4.008 ⁵	4.008	9.482 ⁷
Heavy distillate	2.210 ⁶	603.400 ⁶	140 ⁶
Heavy fuel oil	2.210 ⁶	603.400 ⁶	140 ⁶
Natural gas (1000 Sm ³) ...	0.009 ⁸	NE	NE
LPG	0.009 ⁸	NE	NE
Refinery gas	0.010	NE	NE
CO gas	0.010	NE	NE
Fuel gas	0.010	NE	NE
Landfill gas	0.010	NE	NE
Biogas	0.010 ⁸	4.518	NE
Fuel wood	0.937 ⁹	61.832 ⁹	84
Wood waste	0.294	47.040	84
Black liquor	NE	19.395	2.381
Wood pellets	0.406 ¹⁰	7.829	84
Wood briquettes	0.813	7.829	84
Municipal waste	0.005	0.032 ¹¹	45.150 ¹²
Special waste	1.008 ¹	0.195 ¹¹	225.750 ¹²

¹ Industrial combustion. ² Private cars in 2015. ³ 1997-2015. ⁴ Cruise=0.001. ⁵ Households.

⁶ Ships. ⁷ Services. ⁸ Stationary combustion. ⁹ 2015. ¹⁰ Households: 0.037. ¹¹ 2006-2015.

¹² 2005-2015.

Uncertainty analysis

Long-range transboundary air pollutants

Source for the uncertainty estimates for long-range transboundary air pollutants is Rypdal and Zhang (2001).

Table C1. Summary of expert judgements of uncertainties in point sources

Production type	Number of plants	Pollutant	Emission determination method and uncertainty evaluation	Assessment (average)
Pulp and paper	6	SO ₂	Continuous emission measurements and estimations from sulphur content of fuel. Diffuse emissions of sulphur compounds when producing sulphite pulp. The latter has a higher uncertainty than both the measured and estimated stack emissions.	± 4 %
Oil refineries	2 (3)	SO ₂	Continuous emission measurements and estimations from sulphur content of fuel.	± 5 %
		NO _x	Based on measurements and calculations.	± 10 %
		NM VOC	Combination of point measurements and calculations. Emissions are variable with possibilities of systematic errors. Emissions from loading of products have lower uncertainty than the fugitive. Differences between the refineries due to different technology, products and operations.	± 45 %
Petrochemical industries and gas terminal	4	NO _x	Annual measurements and/or calculations	± 7 %
		NM VOC	Several emission points. Difficult to measure properly and high variability. Uncertainty is in any case lower than for the refineries as mostly gas is handled (high demand for security).	± 25 %
Cement	2	SO ₂	Continuous measurements and annual measurements/calculations. High variability as cement plants incinerates special waste.	± 12 %
		NO _x	Continuous measurements and annual measurements/calculations. High variability as cement plants incinerates special waste.	± 12 %
Ammonia and fertiliser	2	NO _x	Continuous/weekly measurements.	± 7 %
		NH ₃	Several emission points. Several measurements performed each year. Low variability.	± 10 %
Silicon carbide (SiC)	3	SO ₂	Emissions are estimates based on consumption and sulphur content of coke. The sulphur content is measured independently for every delivery. There is, however, uncertainty connected to the end products and degree of oxidation and definition applied, so reporting can seem inconsistent.	± 20 %
Ferroalloys	16	SO ₂	Emissions are estimates based on consumption and sulphur content of coke and the sulphur in products. The sulphur content is measured independently for every delivery. The sulphur content of products are measured regularly, but shows small variability.	± 2 %
		NO _x	Estimates using emission factors. Emission factors are based on measurements. Emission factors are, however, only available for some types of ferroalloys and emissions are not estimated for the others.	± 10-20 %*
Aluminium	8	SO ₂	Monthly measurements (covering emissions from stack and ceiling)	± 7 %
		NO _x	Emissions are estimated based on emission factors (see table 4).	-
Waste incineration	8	SO ₂	Annual representative measurements. Variable emissions due to the waste fraction incinerated.	± 7 %
		NO _x	Annual representative measurements.	± 10 %

* Additional uncertainty due to possible incomplete reporting.

Table C2. Summary of standard deviation and probability density of activity data

SNAP category	Pollutant source	Important for	Standard deviation (2σ). %	Density shape	Source/Comment
01, 02, 03	Gas combustion	NO _x	± 4	Normal	Directorate of oil and gas
01, 02, 03, 07, 08	Oil combustion (total)	SO ₂ , NO _x	± 3	Normal	Spread in data.
0102	Waste combustion - Energy industries	SO ₂ , NO _x , NMVOC	± 5	Normal	Expert judgement
0202	Coal and coke combustion - Residential	SO ₂ , NO _x , NMVOC	± 20	Normal	Expert judgement
090201	Waste combustion - Other sectors	SO ₂ , NO _x , NMVOC	± 30	Lognormal	Expert judgement
01, 02, 03	Wood combustion - All sectors	SO ₂ , NO _x , NMVOC	± 30	Lognormal	Expert judgement
01, 03	Coal and coke combustion- Industry	SO ₂ , NO _x , NMVOC	± 5	Normal	Spread in data
07, 08	Oil, road/off-road/catalytic/non-catalytic	SO ₂ , NO _x , NMVOC, NH ₃	± 20	Normal	Comparisons of data
0805	Oil combustion - Aviation	SO ₂ , NO _x , NMVOC	± 20	Normal	Expert judgement
0804	Oil combustion - Shipping	SO ₂ , NO _x , NMVOC	± 10	Normal	Comparisons of data
0401	Refineries (throughput)	NMVOC	± 3	Normal	Expert judgement
040301	Aluminium production	NO _x	± 3	Normal	Expert judgement
040302	Ferroalloy production	NO _x	± 3	Normal	Expert judgement
040605	Bread production	NMVOC	± 30	Normal	Expert judgement
040607	Beer production	NMVOC	± 10	Normal	Expert judgement
050202	Loading of crude oil	NMVOC	± 3	Normal	Expert judgement
0505	Gasoline distribution	NMVOC	± 3	Normal	Expert judgement
0601	Solvent use	NMVOC			See emission factor
09	Waste combustion in small scale	SO ₂ , NO _x , NMVOC	± 50	Lognormal	Expert judgement
090201	Methane incineration (landfills)	NO _x , NMVOC	± 5	Normal	Expert judgement
090204	Flaring of natural gas	NO _x , NMVOC	± 4	Normal	As combustion of gas
090204	"Flaring" of crude oil	SO ₂ , NO _x , NMVOC	± 10	Normal	Expert judgement
090203/4	Other flaring	NO _x , NMVOC	± 5	Normal	Expert judgement
090207	Incineration of hospital waste	NO _x , NMVOC	± 20	Normal	Expert judgement
090901	Cremation	SO ₂ , NO _x , NMVOC	± 20	Normal	Expert judgement
10	Animal population	NH ₃	± 5-10	Normal	Expert judgement
10	Agricultural soils - Treatment of straw	NH ₃			See emission factor
1001	Agricultural soils - Fertiliser use	NH ₃	± 5	Normal	Agriculture authorities
1009	Agricultural soils - Manure use	NH ₃	± 20	Normal	Expert judgement

Table C3. Summary of standard deviation and probability density of emission factors

SNAP source category	Pollutant source	Standard deviation (2σ). %	Density shape	Source/Comment
01, 02, 03	SO ₂ - Oil combustion, general	± 1	Normal	Expert judgement. Oil companies
01, 02, 03	SO ₂ - Oil combustion, heavy fuel oil	-50 - +100	Normal	Expert judgement. Oil companies
01, 03	SO ₂ - Coal combustion	-50 - +100	Lognormal	Spread in data
01, 03	SO ₂ - Wood combustion	-50 - +100	Lognormal	Spread in data
0804	SO ₂ - Oil combustion, domestic shipping	± 25	Normal	Expert judgement. Oil companies
01, 02 (+03)	NO _x - Combustion in area sources	± 40-50	Normal	Spread in data
0105	NO _x - Combustion off-shore	± 40	Lognormal	Expert judgement
040301	NO _x - Aluminium production	-50 - +100	Lognormal	Expert judgement
07	NO _x - Road traffic	± 25-30	Normal	Expert judgement, spread in data
0704/0705	NO _x - Motorcycles	± 40	Normal	Expert judgement, spread in data
0801-02, 0806-09	NO _x - Equipment and railways	± 40	Normal	Spread in data
0804	NO _x - Shipping	± 15	Normal	Spread in data
0805	NO _x - Aircraft	± 20	Normal	EEA (2000)
0902	NO _x - Flaring	± 40	Lognormal	Expert judgement
01, 02 (+03)	NM VOC - Combustion in area sources	± 40-50	Normal	Spread in data
0105	NM VOC - Combustion offshore	± 50	Lognormal	Expert judgement
040605/07	NM VOC- Beer and bread production	-50 - +100	Lognormal	EEA (2000)
050201	NM VOC- Oil loading onshore	± 30	Normal	Rypdal (1999), Expert judgement
050202	NM VOC- Oil loading offshore	± 40	Normal	Rypdal (1999), Expert judgement
0505	NM VOC - Gasoline distribution	± 50	Lognormal	EEA (2000)
0601	NM VOC - Solvent use	± 30	Normal	Rypdal (1995)
0701	NM VOC - Road traffic (gasoline vehicles)	± 40-50	Normal	Expert judgement, spread in data
0703	NM VOC - Road traffic (diesel vehicles)	± 20-30	Normal	Expert judgement, spread in data
0704/0705	NM VOC - Motorcycles	± 40	Normal	Expert judgement, spread in data
0801-02, 0806-09	NM VOC - Equipment and railways	± 40	Normal	Spread in data
0804	NM VOC - Shipping	± 50	Normal	Spread in data
0805	NM VOC - Aircraft	± 25	Normal	EEA (2000)
0902	NM VOC - Flaring	± 50	Lognormal	Expert judgement
07	NH ₃ - Road traffic	Factor 3	Lognormal	Expert judgement, spread in data
1001	NH ₃ -Agriculture, fertiliser	± 20	Normal	Expert judgement
1005	NH ₃ -Agriculture, animal manure	± 30	Normal	Expert judgement
10	NH ₃ -Agriculture, treatment of straw	± 5	Normal	Expert judgement

Table C4. Uncertainty in emission level of pollutants. 1990, 1998 and 2010

1990	μ (mean) ktonnes	Relative standard deviation (σ/μ)	Uncertainty 2 σ (% of mean)	Uncertainty 2 σ (ktonnes)
SO ₂	52.7	0.02	4.0	2
NO _x	219.0	0.062	12	27
NM VOC	298.4	0.09	18	54
NH ₃	22.9	0.104	21	5
1998	μ (mean) ktonnes	Relative standard deviation (σ/μ)	Uncertainty 2 σ (% of mean)	Uncertainty 2 σ (ktonnes)
SO ₂	29.8	0.021	4.2	1
NO _x	224.0	0.062	12	27
NM VOC	344.5	0.105	21	72
NH ₃	27.0	0.091	18	5
2010*	μ (mean) ktonnes	Relative standard deviation (σ/μ)	Uncertainty 2 σ (% of mean)	Uncertainty 2 σ (ktonnes)
SO ₂	22.0	0.025	5.0	1
NO _x	156.0	0.062	12	19
NM VOC	194.0	0.074	15	29
NH ₃	23.0	0.105	21	5

* Projected data with uncertainties as if they were historical.

Table C5. Uncertainties in emission trends 1990-1998 and 1990-2010

	Absolute change ($\mu_{2010} - \mu_{1990}$)	% change ($(\mu_{2010} - \mu_{1990}) * 100 / \mu_{1990}$)	Relative standard deviation ($\sigma / (\mu_{2010} - \mu_{1990})$)	Uncertainty 2 σ (absolute change)	Uncertainty 2 σ (%-point of change)
1990-1998					
SO ₂	-23.0	-43	-0.04	1.7	3.2
NO _x	+4.8	+2	+3.00	28	13
NM VOC	+43.8	+15	+0.40	35	12
NH ₃	+4.1	+18	+0.22	1.8	8.0
1990-2010					
SO ₂	-30.7	-58	-0.03	1.8	3.4
NO _x	-62.8	-29	-0.21	26.9	12
NM VOC	-104.9	-35	-0.18	38	13
NH ₃	+0.0	0	61.3	3.1	13

* Projected values with uncertainties as if they were historical.

Economic sectors in the Norwegian emission model

The classification is an aggregated version of the one used in the national accounts. To make the standard sectors more appropriate for emission calculations, a few changes have been made, e.g. "Private households" is defined as a sector. The classification is aggregated from the Norwegian *Standard Industrial Classification*, SIC2007 (Statistics Norway 2009). The sic is identical to the European NACE (rev. 2) classification up to the four-digit level. A national level has been introduced at the five-digit level.

The sector numbers in the model have six or, in a few cases, eight digits. The first two digits refer to the main sectors of the economy: 23 = private sector, 24 = central government, 25 = local government, 33 = private households, and 66 = foreign activity. For clarity, the two first digits are only included for the first sector listed in each main sector in the table below.

The next four digits are approximate SIC codes. The first two of these in most cases correspond to SIC at the two-digit level, but some sector numbers, particularly those used for service industries, are aggregates of several SIC divisions. The detailed relationship is shown in the following table, where the sectors are listed with the corresponding SIC codes.

For emissions from solvents and paraffin wax, figures are available at a somewhat more disaggregated sector level, but since these sectors do not reflect the general detailing level in the emission calculations, they are not included in the table below.

Sector number	SIC code	Sector name
Agriculture and forestry		
230100	01.01-5, 01.7	Agriculture
0160	01.6	Services related to agriculture
0210	02	Forestry and logging
Fishing		
0310-N	03.1	Fishing
0320	03.2	Operation of fish farms
Energy sectors		
0500	05	Coal mining
0600.1	06 part, 49.5	Extraction of crude petroleum and natural gas, offshore: Permanent installations
0600.2	06 part	Extraction of crude petroleum and natural gas, offshore: Moveable installations
0600.3	06 part	Extraction of crude petroleum and natural gas: Plants on shore
1910.2	19.1 part	Coking plants
1922	19.2 part	Manufacture of refined petroleum products
3510	35.12, 35.13, 35.14	Transmission, distribution and trade of electricity
3511	35.11	Production of electricity
3520	35.2	Manufacture and distribution of gas
3530	35.3	Steam and hot water supply
Mining/manufacturing		
0710	07.1, 07.29	Mining of ores except uranium and thorium
0721	07.21	Mining of uranium and thorium ores
0810	08 except 08.92	Quarrying and mining except ores and extraction of peat
0892	08.92	Extraction and agglomeration of peat
0910	09.1, 52.215	Service activities incidental to oil and gas extraction
0990	09.9	Service activities incidental to mining
1010	10.1	Production, processing and preserving of meat and meat products
1020	10.2	Processing and preserving of fish and fish products
1030	10.3	Processing and preserving of fruit and vegetables
1040	10.4	Manufacture of vegetable and animal oils and fats
1050	10.5	Manufacture of dairy products
1060	10.6	Manufacture of grain mill products, starches and starch products
1080	10.7, 10.8	Manufacture of other food products
1090	10.9	Manufacture of prepared animal feeds
1100	11	Manufacture of beverages
1200	12	Manufacture of tobacco products
1300	13	Manufacture of textiles and textile products
1400	14	Manufacture of wearing apparel
1500	15	Manufacture of leather, leather products and footwear
1610	16.1	Sawmilling and planing of wood, impregnation of wood
1620	16.21, 16.22, 16.24, 16.29	Manufacture of products of wood, cork, straw and plaiting materials, except furniture
1630	16.23	Manufacture of builders' supplies
1711	17.11	Manufacture of pulp
1712	17.12	Manufacture of paper and paperboard
1720	17.2	Manufacture of articles of paper and paperboard
1800	18	Printing and service activities related to printing and reproduction of recorded media
1910.1	19.1 part	Manufacture of coke oven products
1921	19.2 part	Manufacture of refined petroleum products except oil refineries
2011	20.11, 20.12, 20.13	Manufacture of basic chemicals
2014	20.14	Manufacture of other organic basic chemicals
2015	20.15	Manufacture of fertilisers and nitrogen compounds

Sector number	SIC code	Sector name
2016	20.16, 20.17	Manufacture of plastics and synthetic rubber in primary forms
2020	20.2	Manufacture of pesticides and other agrochemical products
2030	20.3	Manufacture of paints and varnishes, printing ink and mastics
2040	20.4	Manufacture of soap and detergents and toilet preparations
2050	20.5, 20.6	Manufacture of other chemical products
2100	21	Manufacture of basic pharmaceutical products and pharmaceutical preparations
2200	22	Manufacture of rubber and plastic products
2310	23.1	Manufacture of glass and glass products
2320	23.2, 23.3, 23.4	Manufacture of refractory products, clay building materials and other porcelain and ceramic products
2350	23.5	Manufacture of cement, lime and plaster
2360	23.6, 23.7, 23.9	Manufacture of products of cement, lime and plaster and other non-metallic mineral products
2411	24.101, 24.2, 24.3	Manufacture of basic iron and steel
2412	24.102	Manufacture of ferroalloys
2440	24.4 except 24.42	Other non-ferrous metal production
2442	24.42	Aluminium production
2451	24.51, 24.52	Casting of iron and steel
2453	24.53, 24.54	Casting of light metals and other non-ferrous metals
2510	25.1, 25.2, 25.3	Manufacture of structural metal products, tanks, reservoirs and containers etc. of metal
2570	25.7	Manufacture of cutlery, tools and general hardware
2590	25.4, 25.5, 25.6, 25.9	Manufacture of other metal products
2610	26.1, 26.2	Manufacture of electronic components and computers
2630	26.3	Manufacture of communication equipment
2640	26.4	Manufacture of consumer electronics
2650	26.5, 26.6, 26.7, 26.8	Manufacture of other electronic and optical products
2750	27.5	Manufacture of domestic appliances
2790	27.1, 27.2, 27.3, 27.4, 27.9	Manufacture of other electrical apparatus and equipment
2810	28.1, 28.2	Manufacture of general-purpose machinery
2830	28.3, 28.4, 28.9	Manufacture of special-purpose machinery
2900	29	Manufacture of motor vehicles and parts and accessories for motor vehicles
3011	30.1 except 30.113 and 30.116	Building of ships and boats
3012	30.113, 30.116	Building of oil platforms
3020	30.2	Manufacture of railway and tramway locomotives and rolling stock
3030	30.3	Manufacture of aircraft and spacecraft
3090	30.4, 30.9	Manufacture of other transport equipment
3100	31	Manufacture of furniture
3210	32.1	Manufacture of jewellery, bijouterie and related articles
3290	32.2, 32.3, 32.4, 32.5, 32.9	Other manufacturing
3310	33.1	Repair of fabricated metal products, machinery and equipment
3320	33.2	Installation of industrial machinery and equipment

Water supply, sewerage, waste management and remediation activities

3600	36	Water collection, treatment and supply
3800	37-39	Sewerage, waste collection, treatment and disposal activities; materials recovery

Construction

4120	41.2, 42, 43	Construction
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Sector number	SIC code	Sector name
Wholesale and retail trade		
4700	45, 46, 47	Wholesale and retail trade, repair of motor vehicles and personal and household goods
Transport etc.		
4910	49.1, 49.2	Transport via railways
4932	49.32	Taxi operation
4939	49.31, 49.39	Other land passenger transport
4940	49.4	Freight transport by road
5020.N	50.101, 50.201	Ocean transport
5030	50.102, 50.109, 50.202, 50.203, 50.204, 50.3, 50.4	Inland and coastal water transport
5100.1N	51 part	Domestic air transport
5100.2N	51 part	International air transport
5222	52 except 52.215, 79	Supporting and auxiliary transport activities
5300	53, 61	Post and telecommunications
Accommodation and food service activities		
5500	55, 56	Accommodation, food and beverage service activities
Business services		
5800	58	Publishing activities
6200	62, 63, 95	Information technology services
6600	64, 65, 66	Financial and insurance activities
6810	41.1, 68	Real estate activities
7100	69-71, 73-74, 78, 80-82	Other business activities
7200	72	Research and development
7700	77	Rental and leasing activities
8500	85	Education
8600	75, 86-88	Health and social work
9300	59-60, 90-93	Recreational, cultural and sporting activities
9400	94, 99	Activities of membership organisations
9600	96	Other service activities
Central government		
245222	52, 79	Supporting and auxiliary transport activities
7100	69-71, 73-74, 78, 80-82	Other business activities
7200	72	Research and development
8410	84.1, 84.21, 84.23, 84.24, 84.25, 84.3	Public administration
8422	84.22	Defence
8500	85	Education
8600	75, 86-88	Health and social work
9300	59-60, 90-93	Other service activities
Local government		
253700	37	Sewerage
3800	38	Waste collection, treatment and disposal activities; materials recovery
6000	59-60, 90-93	Other service activities
8410	84.1, 84.21, 84.23, 84.24, 84.25, 84.3	Public administration
8500	85	Education
8600	75, 86-88	Health and social work
9600	96	Other personal service activities

Sector number	SIC code	Sector name
Private households		
330000	n.a.	Private households
Foreign activities in Norway		
665020	n.a.	Foreign activities in Norway, ocean transport
665100.2	n.a.	Foreign activities in Norway, air transport

Source classifications used in the Norwegian emission inventory

Table E1. Source classifications used in the national emission inventory

Oil and gas extraction
Oil and gas extraction (stationary combustion)
Offshore
Natural gas in turbines etc., offshore
Flaring, offshore
Diesel fixed installations
Diesel mobile installations, production
Diesel mobile installations, exploration
Well testing
Onshore installations
Natural gas in turbines etc., onshore installations
Flaring, onshore installations
Oil and gas extraction (process emissions)
Offshore
Cold flaring and leakage
Oil loading at sea
Onshore installations
Oil loading, land
Gas terminals
Manufacturing industries and mining
Manufacturing industries and mining, stationary combustion
Wood processing
Oil refining
Chemical industries
Petrochemistry
Fertiliser
Other chemical industries
Mineral industry
Cement, lime and plaster
Other mineral industries
Metal industry
Other industries and mining
Manufacturing industries and mining, processes
Wood processing
Oil refining
Chemical industries
Petrochemistry
Fertiliser

	Carbides
	Other chemical industries
Mineral industry	
	Cement
	Other mineral industries
Metal industry	
	Iron, steel and ferro-alloys
	Aluminium
	Other metals
	Anodes
Other manufacturing industries and mining	
	Coal mining
	Other mining
	Fermentation (bread and beer)
	Asphalt production plants
	Other industries
Energy supply	
Heating in other industries and households	
	Heating in other industries
	Heating in primary industries
	Heating in construction and building
	Heating in other service industries
	Heating in households
Road traffic	
	Passenger cars
	Passenger cars - petrol
	Passenger cars - diesel
	Light duty vehicles
	Light duty vehicles- petrol
	Light duty vehicles - diesel
	Heavy duty vehicles
	Heavy duty vehicles - petrol
	Heavy duty vehicles - diesel etc.
	Motorcycles and mopeds
	Motorcycles
	Mopeds
Aviation, navigation, fishing, motorized equipment etc.	
	Railways
	Domestic aviation
	Domestic aviation < 1000 m
	Domestic aviation > 1000 m
	Coastal navigation
	Navigation - Coastal traffic etc.

	Navigation - Fishing
	Other mobile combustion
	Small boats
	Snowmobiles
	Tractors, constructions machines and other motorized equipment: diesel
	Tractors, constructions machines and other motorized equipment: petrol
Agriculture	
	Enteric fermentation and manure
	Enteric fermentation
	Manure
	Fertiliser and agriculture, other
	Fertilizer
	Agriculture, other
International transportation	
	International navigation
	International aviation
Other	
	Landfill
	gas
	Road, tyre and brake wear
	Road wear
	Tyre and brake wear
	Railway contact wire abrasion
	Products containing fluorinated gases, solvents etc.
	Products containing fluorinated gases
	Other products, including solvents
	Other
	Fires, cremations etc.
	Gas distribution
	Petrol distribution
	Whitening of industrial waste
	Waste water and waste water handling
	Sources not mentioned elsewhere

Table E2.EMEP/NFR14 source sector categories

1A1a	Public electricity and heat production
1A1b	Petroleum refining
1A1c	Manufacture of solid fuels and other energy industries
1A2a	Stationary combustion in manufacturing industries and construction: Iron and steel
1A2b	Stationary combustion in manufacturing industries and construction: Non-ferrous metals
1A2c	Stationary combustion in manufacturing industries and construction: Chemicals
1A2d	Stationary combustion in manufacturing industries and construction: Pulp, Paper and Print
1A2e	Stationary combustion in manufacturing industries and construction: Food processing, beverages and tobacco
1A2f	Stationary combustion in manufacturing industries and construction: Non-metallic minerals
1A2gvii	Mobile Combustion in manufacturing industries and construction: (please specify in the IIR)
1A2gviii	Stationary combustion in manufacturing industries and construction: Other (please specify in the IIR)
1A3ai(i)	International aviation LTO (civil)
1A3aii(i)	Domestic aviation LTO (civil)
1A3bi	Road transport: Passenger cars
1A3bii	Road transport: Light duty vehicles
1A3biii	Road transport: Heavy duty vehicles and buses
1A3biv	Road transport: Mopeds & motorcycles
1A3bv	Road transport: Gasoline evaporation
1A3bvi	Road transport: Automobile tyre and brake wear
1A3bvii	Road transport: Automobile road abrasion
1A3c	Railways
1A3di(ii)	International inland waterways
1A3dii	National navigation (shipping)
1A3ei	Pipeline transport
1A3eii	Other (please specify in the IIR)
1A4ai	Commercial/institutional: Stationary
1A4aii	Commercial/institutional: Mobile
1A4bi	Residential: Stationary
1A4bii	Residential: Household and gardening (mobile)
1A4ci	Agriculture/Forestry/Fishing: Stationary
1A4cii	Agriculture/Forestry/Fishing: Off-road vehicles and other machinery
1A4ciii	Agriculture/Forestry/Fishing: National fishing
1A5a	Other stationary (including military)
1A5b	Other, Mobile (including military, land based and recreational boats)

1B1a	Fugitive emission from solid fuels: Coal mining and handling
1B1b	Fugitive emission from solid fuels: Solid fuel transformation
1B1c	Other fugitive emissions from solid fuels
1B2ai	Fugitive emissions oil: Exploration, production, transport
1B2aiv	Fugitive emissions oil: Refining / storage
1B2av	Distribution of oil products
1B2b	Fugitive emissions from natural gas (exploration, production, processing, transmission, storage, distribution and other)
1B2c	Venting and flaring (oil, gas, combined oil and gas)
1B2d	Other fugitive emissions from energy production
2A1	Cement production
2A2	Lime production
2A3	Glass production
2A5a	Quarrying and mining of minerals other than coal
2A5b	Construction and demolition
2A5c	Storage, handling and transport of mineral products
2A6	Other mineral products (please specify in the IIR)
2B1	Ammonia production
2B2	Nitric acid production
2B3	Adipic acid production
2B5	Carbide production
2B6	Titanium dioxide production
2B7	Soda ash production
2B10a	Chemical industry: Other (please specify in the IIR)
2B10b	Storage, handling and transport of chemical products (please specify in the IIR)
2C1	Iron and steel production
2C2	Ferroalloys production
2C3	Aluminium production
2C4	Magnesium production
2C5	Lead production
2C6	Zinc production
2C7a	Copper production
2C7b	Nickel production
2C7c	Other metal production (please specify in the IIR)
2C7d	Storage, handling and transport of metal products (please specify in the IIR)
2D3a	Domestic solvent use including fungicides
2D3b	Road paving with asphalt
2D3c	Asphalt roofing
2D3d	Coating applications
2D3e	Degreasing
2D3f	Dry cleaning

2D3g	Chemical products
2D3h	Printing
2D3i	Other solvent use (please specify in the IIR)
2G	Other product use (please specify in the IIR)
2H1	Pulp and paper industry
2H2	Food and beverages industry
2H3	Other industrial processes (please specify in the IIR)
2I	Wood processing
2J	Production of POPs
2K	Consumption of POPs and heavy metals (e.g. electrical and scientific equipment)
2L	Other production, consumption, storage, transportation or handling of bulk products (please specify in the IIR)
3B1a	Manure management - Dairy cattle
3B1b	Manure management - Non-dairy cattle
3B2	Manure management - Sheep
3B3	Manure management - Swine
3B4a	Manure management - Buffalo
3B4d	Manure management - Goats
3B4e	Manure management - Horses
3B4f	Manure management - Mules and asses
3B4gi	Manure management - Laying hens
3B4gii	Manure management - Broilers
3B4giii	Manure management - Turkeys
3B4giv	Manure management - Other poultry
3B4h	Manure management - Other animals (please specify in IIR)
3Da1	Inorganic N-fertilizers (includes also urea application)
3Da2a	Animal manure applied to soils
3Da2b	Sewage sludge applied to soils
3Da2c	Other organic fertilisers applied to soils (including compost)
3Da3	Urine and dung deposited by grazing animals
3Da4	Crop residues applied to soils
3Db	Indirect emissions from managed soils
3Dc	Farm-level agricultural operations including storage, handling and transport of agricultural products
3Dd	Off-farm storage, handling and transport of bulk agricultural products
3De	Cultivated crops
3Df	Use of pesticides
3F	Field burning of agricultural residues
3I	Agriculture other (please specify in the IIR)
5A	Biological treatment of waste - Solid waste disposal on land

5B1	Biological treatment of waste - Composting
5B2	Biological treatment of waste - Anaerobic digestion at biogas facilities
5C1a	Municipal waste incineration
5C1bi	Industrial waste incineration
5C1bii	Hazardous waste incineration
5C1biii	Clinical waste incineration
5C1biv	Sewage sludge incineration
5C1bv	Cremation
5C1bvi	Other waste incineration (please specify in the IIR)
5C2	Open burning of waste
5D1	Domestic wastewater handling
5D2	Industrial wastewater handling
5D3	Other wastewater handling
5E	Other waste (please specify in IIR)
6A	Other (included in national total for entire territory) (please specify in IIR)

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The Norwegian Environment Agency is working for a clean and diverse environment. Our primary tasks are to reduce greenhouse gas emissions, manage Norwegian nature, and prevent pollution.

We are a government agency under the Ministry of Climate and Environment and have 700 employees at our two offices in Trondheim and Oslo and at the Norwegian Nature Inspectorate's more than sixty local offices.

We implement and give advice on the development of climate and environmental policy. We are professionally independent. This means that we act independently in the individual cases that we decide and when we communicate knowledge and information or give advice.

Our principal functions include collating and communicating environmental information, exercising regulatory authority, supervising and guiding regional and local government level, giving professional and technical advice, and participating in international environmental activities.